

1 **Interagency Ecological Program**
2 **2010 Pelagic Organism Decline Work Plan**
3 **and**
4 **Synthesis of Results**
5

6 Prepared by: Randall Baxter (DFG), Rich Breuer (DWR), Larry Brown (USGS), Louise Conrad
7 (DWR), Fred Feyrer (USBR), Stephanie Fong (CVRWQCB), Karen Gehrts (DWR), Lenny
8 Grimaldo (USBR), Bruce Herbold (USEPA), Peter Hrodey (USFWS), Anke Mueller-Solger (DSC),
9 Ted Sommer (DWR), and Kelly Souza (DFG)

10
11 6 December 2010

12
13 Interagency Ecological Program for the San Francisco Estuary
14 A Cooperative Program of
15 California Dept of Water Resources
16 California Dept. of Fish and Game
17 State Water Resources Control Board
18 U.S. Fish and Wildlife Service
19 U.S. Bureau of Reclamation
20 U.S. Geological Survey
21 U.S. Army Corps of Engineers
22 U.S. Environmental Protection Agency
23 National Marine Fisheries Service
24

Contents

27	Executive Summary	9
28	Introduction.....	12
29	Purpose.....	14
30	Species Background.....	15
31	Conceptual Models	17
32	Basic Conceptual Model	19
33	<i>Previous Abundance</i>	19
34	<i>Habitat</i>	22
35	<i>Top-Down</i>	38
36	<i>Bottom-Up</i>	45
37	Species-specific Models.....	53
38	<i>Delta Smelt</i>	54
39	<i>Longfin smelt</i>	61
40	<i>Striped bass</i>	69
41	<i>Threadfin shad</i>	75
42	The Pelagic Organism Decline: A Historical Perspective	79
43	Why a historical perspective?	79
44	<i>Change in Ecosystems</i>	80
45	<i>Environmental History: Four Eras</i>	81
46	<i>Human History – Then and Now</i>	87
47	A Tale of Two Estuaries	88
48	Why a new conceptual model?	88
49	<i>Ecological regime shift</i>	89
50	<i>POD – a regime shift</i>	90
51	The new pelagic regime in the Sacramento-San Joaquin Delta.....	97
52	The new benthic regime in the Sacramento-San Joaquin Delta	98
53	The new littoral regime in the Sacramento-San Joaquin Delta.....	101
54	2010 POD Program.....	104
55	References Cited	104
56	Tables.....	128
57	Figures.....	135
58	Appendix 1.....	176

60 List of Acronyms

Acronym	Definition
BDCP	Bay-Delta Conservation Plan
BEM	bioenergetic modeling
CCF	Clifton Court Forebay
CDFG	California Department of Fish and Game
CESA	California Endangered Species Act

CPUE	catch per unit effort
CSTARS	Center for Spatial Technology and Remote Sensing
CVP	Central Valley Project
CVRWQCB	Central Valley Regional Water Quality Control Board
DDT	dichlorodiphenyltrichloroethane
DO	dissolved oxygen
DRERIP	Delta Regional Ecosystem Restoration Implementation Plan
DSP	Delta Science Program
DWR	California Department of Water Resources
E/I	export to inflow
EDC	endocrine disrupting compounds
EQ	environmental quality
FESA	Federal Endangered Species Act
FMWT	Fall Midwater Trawl
IEP	Interagency Ecology Program
LC	Lethal Concentration
NCEAS	National Center for Ecological Analysis and Synthesis
OMR	Old and Middle rivers
OP	organophosphate
PBDEs	polybrominated diphenyl ethers
PBO	piperonyl butoxide
PCBs	Polychlorinated biphenyl
PCR	polymerase chain reaction
POD	Pelagic Organism Decline
POD-MT	POD Management Team
SAV	submerged aquatic vegetation
SFF	State Skinner Fish Protective Facility
SL	Standard Length
SPMDS	semi-permeable membrane devices
SRWTP	Sacramento Regional Wastewater Treatment Plant
SWP	State Water Project
SWRCB	California State Water Resources Control Board
TFCF	federal Tracy Fish Collection Facility
TL	total length
TMDL	Total Maximum Daily Load
TNS	Summer Towntnet Survey
USBR	US Bureau of Reclamation
USGS	US Geological Survey
VAMP	Vernalis Adaptive Management Plan

62

63

64 **List of Tables**

65

66 Table 1. Linkages of POD study elements to the drivers of the basic conceptual model, species-
67 specific conceptual models, season when results of the element apply, and contribution to
68 understanding the regime shift conceptual model.

69

70 Table 2. Costs and funding sources for individual POD study elements.

71

72 **List of Figures**

73

74 1. Map of the (A) the entire San Francisco Estuary and (B) the Sacramento-San Joaquin Delta.

75

76 2. Trends in abundance indices for four pelagic fishes from 1967 to 2009 based on the Fall
77 Midwater Trawl, a California Department of Fish and Game survey that samples the upper San
78 Francisco Estuary. No sampling occurred in 1974 or 1979 and no index was calculated for 1976.
79 Development of abundance indices from catch data is described by Stevens and Miller (1983).
80 Note that the y-axis for longfin smelt represents only the lower 25% of its abundance range to
81 more clearly portray the lower abundance range.

82

83 3. The basic conceptual model for the pelagic organism decline (updated from Sommer et al.
84 2007).

85

86 4. Delta smelt species model.

87

88 5. Longfin smelt species model. The dotted line indicates that the importance of a stock
89 recruitment relationship is unclear. The stage recruitment loop illustrates that both survival from
90 age-0 to age-1 and from age-1 to age-2 are important.

91

92 6. Striped bass species model. The dotted line indicates that the form of the stock recruitment
93 relationship has changed and the present stock-recruitment relationship is unclear.

94

95 7. Threadfin shad species model. The dotted line indicates that the importance of a stock
96 recruitment relationship is unclear.

97

98 8. The ecological regime shift in the Delta results from changes in (slow) environmental drivers
99 that lead to profoundly altered biological communities and, as soon as an unstable threshold
100 region is passed, a new relatively stable ecosystem regime.

101

102 9. Peterson population estimates of the abundance of adult (3+) striped bass > 460 mm total
103 length from 1969 to 2008, with 95% confidence limits (California Department of Fish and Game,
104 unpublished data). Striped bass were only tagged during even years from 1994 to 2002, so no
105 estimates are available for odd years during that period.

106

107 10. Number of male and female striped bass collected and corresponding female:male ratio from
108 the Adult Striped Bass Tagging Program, 1969 – 2008. No estimates available for odd years
109 between 1995 and 2001 or 2006.

- 110
111 11. Log-log relationships between Fall Midwater Trawl abundance and delta outflow for longfin
112 smelt (all ages) and age-0 striped bass. Delta outflow (m^3/s) values represent mean levels during
113 January through June for longfin smelt and during April through July for age-0 striped bass .
114 The data are compared for pre-*Corbula* invasion years (1967–1987; white circles), post-*Corbula*
115 invasion (1988–2000; filled circles), and during the POD years (2001–2009 ; triangles). Lines
116 depict significant linear regression relationships ($p < 0.05$).
117
- 118 12. Scatterplots and LOWESS splines depicting the relationship between a) the Fall Midwater
119 Trawl index of delta smelt relative abundance (FMWT) versus the following calendar year's
120 Summer Towntnet Survey index of delta smelt relative abundance (TNS) for FMWT years 1968–
121 2007; b) the FMWT versus the following calendar year's TNS for FMWT years 1978–2007; c)
122 the TNS and the subsequent FMWT for 1969–2008; and d) the TNS and the subsequent FMWT
123 for 1978–2008.
124
- 125 13. Ln–ln relationships between juvenile and adult lifestages of delta smelt since 1969. The
126 Towntnet Survey is a measure of summer juvenile abundance. The Fall Midwater Trawl is a
127 measure of fall pre-spawning adult abundance. The blue circles represent the data from the first
128 2 Towntnet Surveys which typically begins in June or July and are used to calculate the Towntnet
129 Survey delta smelt index. The red squares represent data from July only and represent an
130 alternate index. Regression equations and coefficients are given in blue font for the full Towntnet
131 Survey data and in red font for the July Towntnet Survey data. Data from 2005 to 2009 are
132 contained in the green ellipse.
133
- 134 14. Log–log relationships between Fall Midwater Trawl abundance and Fall Midwater Trawl
135 abundance in the previous year. The arrows connect consecutive years from 2000 to 2009.
136
- 137 15. Abundance of age-1 and age-2+ striped bass in midwater trawls in A) San Francisco Bay
138 based on the California Department of Fish and Game Bay study (Bay Study) and B) in the Delta
139 from the Fall Midwater Trawl. No Fall Midwater Trawl indices were calculated for years 1974,
140 1976 and 1979.
141
- 142 16. Average catch per unit effort (CPUE) for largemouth bass, bluegill, and redear sunfish from
143 successive electrofishing efforts in the Delta from 1980 to 2010. Data from 1980–2003 is from
144 CDFG (unpublished data) and data from 2008–2010 is from L. Conrad (DWR, unpublished
145 data).
146
- 147 17. Annual salvage density (fish per 10,000 acre feet) of largemouth bass at the CVP and SWP
148 combined from 1979 to 2009 (California Department of Fish and Game, unpublished data).
149
- 150 18. Winter salvage data for striped bass, delta smelt, longfin smelt, and threadfin shad for the
151 federal Central Valley Project (Federal) and State Water Project (State) from 1981 to 2009.
152 Salvage for delta smelt and longfin smelt before 1993 should be interpreted with caution because
153 of variable degrees of training among personnel identifying fishes.
154

- 155 19. Winter salvage for largemouth bass, inland silversides, bluegill, and redear sunfish for the
156 federal Central Valley Project (Federal) and State Water Project (State) from 1981 to 2009.
157
- 158 20. Relationship of mean combined salvage of delta smelt, longfin smelt, and striped bass at the
159 State Water Project (SWP) and Central Valley Project (CVP) to combined Old and Middle rivers
160 flow (cubic feet per second). Open symbols denote pre-POD years (1993–1999) and filled
161 symbols represent post-POD years (2000–2005) (Grimaldo et al. 2009).
162
- 163 21. Delta outflow (m^3/s) averaged over water years (top) and export flow (m^3/s) averaged over
164 seasons (bottom). Water years begin on 1 October of the previous calendar year. Seasons are in
165 3-month increments starting in October. Export flows are the sum of diversions to the federal
166 Central Valley Project and State Water Project pumping plants. The outflow and export data are
167 from California Department of Water Resources (<http://iep.water.ca.gov/dayflow>).
168
- 169 22. Range in primary production in Suisun Bay and the Delta since 1975 plotted on the
170 relationship of fishery yield to primary production from other estuaries around the world
171 (modified from Nixon 1988, using results in Jassby 2008 and Jassby et al 2002 and data provided
172 by James Cloern, U.S. Geological Survey).
173
- 174 23. Summer to fall survival index of delta smelt in relation to zooplankton biomass in the low
175 salinity zone (0.15 – 2.09 psu) of the estuary. The survival index is the log ratio of the Fall
176 Midwater Trawl index to the Summer Towntnet Survey index. The line is the geometric mean
177 regression for log(10)-transformed data, $y = 2.48x - 0.36$. The correlation coefficient for the
178 log-transformed data is 0.58 with a 95% confidence interval of (0.26, 0.78) (Kimmerer, 2008).
179
- 180 24. Prey volume in guts of delta smelt collected during summer 2005 and 2006. Sample size
181 appears in parentheses (S. Slater, California Department of Fish and Game, unpublished data).
182
- 183 25. Water temperature influences on growth in the juvenile stage (i.e. specific growth) from
184 studies by Baskerville et al. (2004) during aquaculture. A cubic spline model was fitted to
185 approximate the functional form of specific growth with respect to water temperature (from
186 Bennett et al. 2008).
187
- 188 26. Seasonal means of X2, the location of the 2‰ bottom isohaline along the axis of the estuary
189 (distance from the Golden Gate in km) from 1967–2009. Symbols indicate water year types for
190 the Sacramento Valley (W: Wet, AN: Above Normal, BN: Below Normal, D: Dry, C: Critically
191 Dry). Map (modified from Jassby et al 1995): Suisun Bay and the western Delta. Insert shows
192 lines positioned along nominal distances (in km) from the Golden Gate along the axis of the
193 estuary.
194
- 195 27. Longfin smelt outflow abundance relationships based on December through May mean
196 outflow (cfs) and (a) Fall Midwater Trawl annual abundance, 1967–2009, all ages, (b) Bay Study
197 Age-0 midwater trawl abundance, 1980–2009, and (c) Bay Study age-0 otter trawl abundance,
198 1980–2009. Abundance data are compared for pre-*Corbula* invasion years (survey start – 1987,
199 open circles), post-*Corbula* invasion (1988–2000, filled circles) and POD years. Fitted lines
200 indicate linear regression relationships that are significant at the $p < 0.05$ level.

- 201
202 28. Ratio of annual longfin smelt catch (all ages) in the Bay Study midwater (MWT) and otter
203 (OT) trawls (log 10 transformed), 1980–2009. Catch data are from all months of the year
204 sampled, which varied by year, and all 35 core stations where valid tows were completed for
205 both nets.
206
- 207 29. Total June through November longfin smelt catch (all ages) in the Bay Study midwater
208 (MWT) and otter (OT) trawls, 1980–2000 for 35 core stations where valid tows were completed
209 for (a) both the nets, and (b) the same data for POD years 2001–2009.
210
- 211 30. Harvest and CPUE for Striped Bass by Commercial Passenger Fishing Vessels in the San
212 Francisco Bay and Delta, 1980–2009.
213
- 214 31. San Francisco Bay Study age-0 striped bass catch May–October. The blue line and circles
215 represent the proportion of the total catch of the combined midwater trawl (MWT) and otter
216 trawl catch (OT) that was taken in the midwater trawl; data for 1994 are incomplete and were not
217 plotted. The pink line and solid squares represent the proportion of the total otter trawl catch
218 taken at shoal stations (<7m deep).
219
- 220 32. Male and female age distribution by decade for striped bass ages 3 – 8+ in fyke net
221 collections. Note that age distribution is the same for gill-net collected fish (graphic not
222 displayed).
223
- 224 33. Abundance estimate (in thousands) of male and female striped bass $\geq 3 - 8+$ collected in
225 the Adult Striped Bass Tagging Program, 1969 – 2005.
226
- 227 34. Apparent growth of male and female striped bass ages 3 – 7, calculated from data collected
228 from the Adult Striped Bass Tagging Program, 1970 – 2005.
229
- 230 35. Estimates of harvest rate (1960 – 2007), and natural and total annual mortality rate (1969 –
231 2007) of legal-sized striped bass $\geq 3 - 8+$, Adult Striped Bass Tagging Program.
232
- 233 36. Four models of ecological change. The gradual change model shows a gradual linear change
234 from regime A to regime B in response to a gradually changing environmental driver. The
235 transition can be reversed by reversing the driver. The threshold model shows a nonlinear abrupt
236 change from regime A to regime B occurring in response to a relatively small change in the
237 environmental driver. Reducing the environmental driver to a value below the threshold results
238 in a change back to the original regime. The hysteresis model describes a nonlinear response to
239 the environmental driver but the threshold for recovery differs from the original threshold
240 causing the collapse. The fourth model describes an irreversible change (adapted from Davis et
241 al. 2010).
242
- 243 37. Two-dimensional ball-in-cup diagrams showing (left) the way in which a shift in state
244 variables causes the ball to move, and (right) the way a shift in parameters causes the landscape
245 itself to change, resulting in movement of the ball (from Beisner et al. 2003).
246

247 38. Annual average salinity by season (spring, summer, fall) and subregion (see Fig. 1) in the
248 upper San Francisco Estuary between 1972 and 2008. Top panels represent data from stations in
249 the “suisun” subregion, which include Suisun Bay and the western Delta. Bottom panels
250 represent data from stations in the “delta” region, which includes the remaining areas of the
251 Delta. Blue line displays a loess fit \pm standard error (shaded area) (from Winder and Jassby
252 2010).

253

254

255 **Executive Summary**

256
257 Abundance indices for four pelagic fishes in the upper San Francisco Estuary (the Delta and
258 Suisun Bay) rapidly declined to record low levels starting in 2002. These fishes include native
259 delta smelt (listed under federal and California Endangered Species acts) and longfin smelt
260 (listed under the State Endangered Species Act) as well as introduced threadfin shad and juvenile
261 (age-0) striped bass. Three of these species have also experienced more gradual long-term
262 declines, but the recent rapid collapse of all four species to persistently low levels was
263 unexpected given the relatively moderate hydrological conditions in the first half of this decade.
264 In 2005, the IEP formed a Pelagic Organism Decline Management Team (POD-MT) to evaluate
265 the potential causes of the declines.

266
267 The POD-MT has developed several conceptual models to guide work plan development and
268 synthesize results. In this report we update previously developed conceptual models with new
269 results and introduce a new conceptual model:

- 270 1. The “basic POD conceptual model” was introduced in 2006 and groups the effects of
271 potential drivers of the POD into four categories (*previous abundance, habitat, top-down effects,*
272 *and bottom-up effects*);
- 273 2. “Species-specific conceptual models” were introduced in 2008 and show how key
274 population drivers presently affect each of the four POD fish species in each season;
- 275 3. A new conceptual model posits that the POD represents a rapid ecological “regime shift”
276 that followed a longer-term erosion of ecological resilience. We present this conceptual model as
277 a working hypothesis for future investigations.

278
279 Much has been learned about individual drivers and their effects on the POD species over the
280 course of the POD investigation. An initial “triage” approach seeking to rule out individual
281 drivers was unsuccessful – we now have evidence that all investigated drivers may have played a
282 role in the POD. As in previous reports, we summarize new evidence for the effects of individual
283 drivers and some of their interactions in the context of the basic POD conceptual model.

284
285 Results of the POD investigation discussed in the report, organized by the four components of
286 the basic POD conceptual model, include:

- 287
288 1. ***Previous abundance.*** IEP fish monitoring provides population indices, not actual
289 population size estimates. This complicates evaluation of the effects of previous
290 abundance, i.e. stock-recruitment relationships. The delta smelt stock-recruitment
291 relationship appears to be density-independent, particularly since the late 1970s. The fall
292 abundance index for pre-spawning adults explains about half of the variation in the
293 juvenile abundance index for the next summer over the period from 1978 to present. The
294 juvenile to adult transition is influenced by density-dependent mechanisms and declining
295 carrying capacity, especially during the POD years. For longfin smelt, preliminary
296 analyses support a stock-recruitment relationship between adults approaching their
297 second birthday and age-0 fall recruits. A significant stage-recruitment relationship (fall
298 age-0 to fall age-1 abundance) also exists. There is little evidence for strong stock-
299 recruitment or stage-recruitment relationships for threadfin shad. The adult striped bass

300 stock is currently not particularly low and total stock size is not likely a mechanism
301 contributing to recent very low age-0 striped bass abundance. However, the effective
302 spawning stock may have decreased due to a decline in number of females relative to
303 males in the spawning runs.
304

- 305 2. **Habitat.** Changes in habitat drivers affect the amount and suitability of habitat available
306 to the POD fishes. Habitat changes not only affect pelagic fishes, but also their predators
307 and prey, which, in turn, can also have effects on the habitat they occupy. Changes in
308 habitat suitability appear to be important in the long term declines of all of the POD
309 species, but particularly delta smelt and age-0 striped bass. Key drivers are decreasing
310 turbidity and changes in the salinity field. While these changes negatively affect the POD
311 fishes, they appear to benefit other organisms such as the invasive aquatic weed *Egeria*
312 *densa* and harmful bloom-forming algae such as *Microcystis aeruginosa*. Toxic
313 chemicals are more likely to cause chronic effects than acute effects in the POD fishes,
314 except in long-lived striped bass. Large female adult striped bass can transfer
315 bioaccumulated compounds to eggs with negative effects on survival of embryos and
316 larvae. Increasing total ammonia concentrations and *Microcystis* blooms are more likely
317 to have effects through the food web than via direct toxicity to fish (see Bottom-up).
318 Pyrethroid toxicity to invertebrates and endocrine disrupting compounds are of increasing
319 concern in the estuary.
320
- 321 3. **Top-down.** In the basic POD conceptual model, top-down effects refer to mortality from
322 predation and entrainment into water diversions. Piscivorous predators in the Delta
323 include native pikeminnows as well as introduced largemouth bass, striped bass, and
324 Mississippi silversides. Striped bass prey on all four POD species. While increasing in
325 abundance, largemouth bass primarily consume littoral, not pelagic prey. New genetic
326 evidence suggests that Mississippi silversides prey on larval delta smelt. Decreasing
327 turbidity may be increasing the vulnerability of pelagic prey to predators. Mortality
328 associated with the State Water Project (SWP) and Central Valley Project (CVP) water
329 diversions is well-documented in the San Francisco estuary. However, mortality
330 estimates based on fish caught in fish screens at these diversions (salvage) are
331 underestimates because small larval fish are not collected at all, other small fish are
332 caught inefficiently, and entrainment-associated mortality that occurs before fish are
333 collected in the screens (pre-screen loss) is not regularly assessed. Shifting of more
334 exports to winter has been accompanied by increased salvage for POD species and other
335 Delta fishes. However, the population-level effects of increased entrainment remain
336 unclear and may vary greatly within and among years and between species. Greater net
337 flow through Old and Middle Rivers toward the SWP and CVP diversions rather than
338 seaward is associated with greater salvage of adult delta smelt, longfin smelt, and striped
339 bass. Overall, entrainment can affect multiple life stages of the POD fishes and often
340 interacts with other drivers affecting the behavior and spawning success of the POD
341 fishes.
342
- 343 4. **Bottom-up.** Compared to other estuaries, phytoplankton primary productivity in the San
344 Francisco estuary is low and has experienced a long-term decline. The long-term decline
345 has been linked to grazing by invasive clams and to shifts in nutrient ratios and

346 concentrations, especially increasing ammonium concentrations. It has led to a decline in
347 overall food availability for pelagic fishes. In addition, there have been substantial
348 changes in phytoplankton and zooplankton community composition which led to changes
349 in food quality. The now common cyanobacteria *Microcystis aeruginosa* are of
350 particularly low nutritional quality for Delta zooplankton species, while nutritionally
351 valuable diatoms have declined. Similarly, a nutritionally inferior, small copepod species
352 is now the dominant zooplankton species in Suisun Bay and the western Delta. Overall,
353 bottom-up food limitation is likely an important driver influencing long-term fish trends
354 in the upper estuary. However, it is likely not the sole driver behind the recent POD
355 decline for several reasons. First, phytoplankton and zooplankton declines preceded the
356 POD. Second, while high, current clam abundance and biomass is not unprecedented.
357 Finally, while phytoplankton production fuels the Delta food web, new research shows
358 that many zooplankton species currently present in the estuary are omnivorous and can
359 consume microbes utilizing dissolved and particulate organic carbon. Shifts in
360 community composition at the base of the food web may be as important as declines in
361 overall productivity, or perhaps even more important. The current composition may favor
362 non-native over native consumers.

363
364 The emerging conclusion is that the POD was caused by multiple and often interacting drivers.
365 The multi-driver origin of the POD is an important insight. However, it is not particularly helpful
366 to policy makers and managers seeking guidance for management strategies aimed at reversing
367 the POD declines. The POD-MT has thus used two other approaches to evaluating drivers.

368
369 The first approach focuses on how the major drivers differ for each of the four POD fish species,
370 and how they differ in relative importance during different life history stages or seasons. The
371 results are summarized in the species-specific conceptual models and may help policy makers
372 and managers identify targeted management actions and guide further research studies for
373 individual POD species.

374
375 The second approach seeks to understand the POD in the context of an ecological “regime shift”
376 affecting the entire estuarine ecosystem and explores the effects of changing drivers through
377 several historical periods leading up to the POD. In this conceptual model, drivers are
378 distinguished based on their approximate rate of change and their importance to ecological
379 resilience. We hypothesize that drivers that changed slowly over decades (“slow drivers”)
380 contributed to the slow erosion of ecological resilience of the system. This made the system more
381 vulnerable to the effects of drivers that changed more rapidly around the time of the POD (see
382 basic POD conceptual model) and/or have greater species specificity (see species-specific
383 conceptual models). In order of their hypothesized importance to the resilience of the system and
384 approximate rate of change, the slow drivers we propose for the POD regime shift are 1) outflow,
385 2) salinity, 3) landscape, 4) temperature, 5) turbidity, 6) nutrients, 7) contaminants, 8) harvest. In
386 this report, we summarize changes in these environmental drivers and provide hypotheses about
387 their individual and combined effects on the biota and importance to ecological resilience and the
388 POD regime. We also briefly mention other potential slow drivers. Finally, we describe the
389 resulting new ecosystem regime for the pelagic, benthic, and littoral zones of the Delta and
390 Suisun Bay regions. As it becomes more fully developed, the POD regime shift story may inform

391 adaptive management strategies aimed at shifting the ecosystem into a more desirable state and
392 improving long-term ecosystem resilience and adaptive capacity in the face of future threats.

393

394 The number of peer-reviewed POD publications has increased rapidly since 2005. However,
395 many studies that will provide important POD information are still in progress. The 2010 POD
396 work plan includes 39 continuing study elements and 32 new elements. The 2010 POD program
397 is funded by DWR, USBR, and SWRCB (approximately \$7,547,000). POD-related Delta
398 (previously CALFED) Science Program grants are estimated at \$3,019,000. We provide some
399 preliminary results from these studies, but peer-reviewed products may not be available for some
400 time. This report considers available information through about August 2010. We intend to
401 publish a fully peer reviewed, final POD synthesis report in 2012–2013.

402

403 **Introduction**

404

405 The San Francisco Estuary (Figure 1) has been an area of importance to humans starting with
406 Native Americans and continuing to the present day. Beginning with the California Gold Rush in
407 the mid-1800s and subsequent rapid population growth, exploitation of resources broadened and
408 accelerated. The resulting changes in land use from natural landscapes to agriculture and
409 urbanization combined with development of an extensive water management infrastructure have
410 been accompanied by declines in nearly all species of native fish (Moyle 2002; Brown and
411 Moyle 2005). The construction and operation of two large water projects, the federal Central
412 Valley Project (CVP), and the State Water Project (SWP) have been especially important.
413 Mitigation for the potential effects of these projects on important fisheries has included salmon
414 hatcheries below some dams and extensive fish screening facilities at the pumping plants in the
415 Delta (Brown et al. 1996). Despite these efforts, populations of important fish species have
416 declined in the San Francisco Estuary and its watershed, culminating in listings, under the
417 Federal and California Endangered Species Acts (FESA and CESA), of a growing number of
418 species since the 1990s, including winter-run and spring-run Chinook salmon *Oncorhynchus*
419 *tshawytscha*, Central Valley steelhead *O. mykiss*, delta smelt *Hypomesus transpacificus*,
420 Sacramento splittail *Pogonichthys macrolepidotus* (subsequently removed from listed status, but
421 currently undergoing another status review), longfin smelt *Spirinchus thaleichthys* (only under
422 CESA) and green sturgeon *Acipenser medirostris*. The petition to list the San Francisco Estuary
423 longfin smelt population as a distinct population segment under FESA was denied, but a range-
424 wide status review was initiated.

425

426 The four primary pelagic fishes of the upper estuary (delta smelt, longfin smelt, age-0 striped
427 bass *Morone saxatilis*, threadfin shad *Dorosoma petenense*), have shown substantial variability
428 in their populations, with evidence of long-term declines for the first three of these species
429 (Kimmerer et al. 2000, Bennett 2005, Rosenfeld and Baxter 2007, Thomson et al. 2010).
430 Against this backdrop of long-term decline, Fall Midwater Trawl (FMWT) abundance indices for
431 these four pelagic fishes (Figure 2) appeared to decline sharply around 2000, despite hydrology
432 expected to support at least modest fish production based on previous relationships (Sommer et
433 al. 2007). Subsequent statistical analyses have supported the existence of “change points” in the
434 early 2000s (Thomson et al. 2010). These species have remained at low levels through 2009

435 with record low values for delta smelt in 2009, age-0 striped bass in 2004, longfin smelt in 2007,
436 and threadfin shad in 2009. These declines are collectively known as the Pelagic Organism
437 Decline (POD). The POD is now a major policy and management issue in the context of
438 California water resources (Sommer et al. 2007).

439
440 The POD Management Team (POD-MT) was established in 2005 by the Interagency Ecology
441 Program (IEP) to evaluate the potential causes of the species declines (Sommer et al. 2007). The
442 POD-MT organized an interdisciplinary, multi-agency research effort including staff and
443 researchers from California Department of Fish and Game (CDFG), California Department of
444 Water Resources (DWR), California State Water Resources Control Board (SWRCB) and
445 Central Valley Regional Water Quality Control Board (CVRWQCB), U.S. Bureau of
446 Reclamation (USBR), U.S. Fish and Wildlife Service (USFWS), U.S. Environmental Protection
447 Agency (USEPA), U.S. Geological Survey (USGS), and the Delta Science Program (DSP)
448 (formerly CALFED Science Program). Many POD research activities were also carried out by
449 academic, non-profit, and private scientists.

450
451 At the core of the POD investigation has been a series of evolving, non-quantitative conceptual
452 models about the declines. These conceptual models were developed to help organize,
453 synthesize, and communicate existing and emerging information about the declines and to
454 identify data gaps for the initiation of new studies. Particularly useful has been what has come to
455 be known as the “basic POD conceptual model” (Figure 3, updated from Sommer et al. 2007).
456 In addition to the basic model, the POD-MT has developed and refined species-specific
457 conceptual models focusing on seasonal drivers of population dynamics of the four POD fish
458 species (Figures 4–7, first introduced in Baxter et al. 2008) and a conceptual model placing the
459 POD in the context of an ecological regime shift (Figure 8, the model is explained in later
460 sections).

461
462 The POD investigation has taken place in a rapidly evolving water resource management, policy,
463 and science landscape. This rapid change has been caused by growing concerns about the ability
464 to meet the dual goals of water supply reliability for Californians and a functioning Delta
465 ecosystem. These concerns grew even more pressing during the 2007–2009 drought.
466 Additionally, California is currently facing a major economic crisis. Some of the major
467 developments since the last comprehensive POD report (Baxter et al. 2008) include: (1) new
468 Biological Opinions for the operation of the State and Federal Water Projects; (2) a new Federal
469 California Bay-Delta Memorandum of Understanding and formation of a Federal Bay-Delta
470 Leadership Committee; (3) a comprehensive water bill package by the California legislature
471 establishing a Delta Water Master, Delta Stewardship Council, and a Delta Conservancy; (4) a
472 new National Research Council committee on “Sustainable Water and Environmental
473 Management in the California Bay-Delta;” and (5) intense work on a new Bay-Delta
474 Conservation Plan (BDCP). Sound scientific information is regarded as a key ingredient to the
475 success of these efforts. The IEP is one of the main providers of such information for the estuary
476 and has been responding to the new demands by creating a new IEP Lead Scientist position,
477 forming new technical work teams, and focusing new POD studies on emerging information
478 needs. Future changes to the IEP may include new formal partnerships and initiatives that are
479 currently under development by the IEP Coordinators.

480

481 **Purpose**

482

483 This report has three objectives:

484

- 485 1. Synthesize the information collected by the POD investigation through approximately
- 486 August 2010,
- 487 2. Provide a basis for future syntheses and work plans, and
- 488 3. Present the 2010 POD work plan.

489

490 In previous reports, our approach was to synthesize this information in the context of our basic
491 conceptual model (Figure 3; Sommer et al. 2007). In this report, we continue to use the basic
492 conceptual model as an organizing principle, but we also present two additional types of
493 conceptual models. The first type is the “species-specific” conceptual models (Figures 4–7) first
494 introduced in Baxter et al. (2008). This type of conceptual model narrows the basic POD
495 conceptual model to the level of individual POD fish species and shows how key population
496 drivers presently affect the POD fish species in each season. We believe this approach increases
497 the clarity of the presentation because not all new information necessarily applies to all POD
498 species. The second type of conceptual model is new. It broadens the basic POD conceptual
499 model to the ecosystem level and places it in a historical ecological regime shift context. In the
500 regime shift conceptual model, the POD is part of a larger, fundamental change in structure and
501 functions of the San Francisco Estuary ecosystem brought about by changes in multiple and
502 often interacting environmental drivers. Some of these changes occurred slowly or decades ago,
503 while others were more rapid or recent. In this report we describe our current understanding of
504 the present state of the system, the changes that led to this state, and what this may mean for the
505 resilience of the system in the face of future challenges such as global climate change and
506 additional species invasions.

507

508 From the beginning, the POD investigation has been designed to assess the role of multiple
509 drivers individually and in combination. Similar to Baxter et al. (2008) we rely on a *weight of*
510 *evidence* approach (Burkhardt-Holm and Scheurer 2007, Linkov et al. 2009) to synthesize and
511 interpret the many individual and in some cases conflicting lines of evidence emerging from the
512 POD studies. In particular, we examined the multiple types of evidence to develop plausible
513 linkages within our conceptual models. As a result, this report focuses on the linkages within the
514 conceptual models and their contributions to the individual “fish species stories,” the overall
515 “POD story,” and the emerging “regime shift story” provided in the final synthesis, rather than
516 providing in-depth examinations of relationships between individual drivers and fish abundance.

517

518 As for previous reports, many of the POD study elements have not been completed or fully
519 evaluated and we wish to emphasize that our conceptual models represent works in progress.
520 We envision delivering a final comprehensive POD report in 2012–2013. The final POD report
521 will synthesize all prior POD results, including the final analyses and insights resulting from an
522 ongoing collaboration of the IEP with the University of California-Santa Barbara, initially with
523 the National Center for Ecological Analysis and Synthesis (NCEAS). The present progress report
524 was written by members of the IEP POD-MT. It has been reviewed by the IEP Agency
525 Coordinators, but has not undergone any external, independent peer review.

526

527 This report is divided into several major sections. We first provide some basic background on
528 the POD fishes. Second, we present the progression of scientific knowledge on specific topics
529 within the context of the basic conceptual model for the POD that was developed in previous
530 reports and publications (Sommer et al. 2007). Third, we refine the more detailed conceptual
531 models for each of the POD fishes that were first presented in Baxter et al. (2008). Fourth, we
532 present a new conceptual model about ecological regime shift that places the information
533 presented in the general and individual species models in a broader ecosystem context. Finally,
534 we describe how the 2010 work plan addresses information gaps associated with the three
535 conceptual model types and what the next phase of this work might hold. Some material is
536 repeated in several sections of the report to minimize the need for readers to search the entire
537 report for pertinent supporting information.

538
539 This report has the same limitations as earlier reports from the POD-MT. Specifically, many
540 studies initiated by the POD or initiated by others that will provide important POD information
541 are still in progress. As explained in the 2009 Addendum to the 2008 POD Work Plan (Baxter et
542 al. 2009), many POD study elements have also been delayed due to a December 2008 “Stop
543 Work Order” for State bond funded projects that lasted eight months. In addition, there have
544 been two to three mandatory furlough days for State employees which started in February 2009
545 and continue to the present, as of the writing of this report. Preliminary results from studies that
546 are still in progress are provided whenever possible, but peer-reviewed products from these
547 studies may not be available for some time to come. We rely as much as possible on peer-
548 reviewed published literature. When such literature is not available we utilized agency reports
549 and reports to the POD management team regarding POD funded research that are available to
550 the public (POD study reports are available at <http://www.science.calwater.ca.gov/pod>). In some
551 cases, we also include information from articles that have been submitted to scientific journals
552 but that have not yet been accepted for publication; we cite them as “submitted.” All other
553 information (e.g., posters, abstracts) is cited as unpublished data or personal communication with
554 the exception of the IEP Newsletter. The IEP Newsletter is cited as a peer-reviewed publication,
555 even though it is not rigorously peer-reviewed. Articles generally receive informal review before
556 submission and are reviewed by an editor with local experience with San Francisco Estuary
557 issues. Also, the IEP Newsletter publishes articles of intense local interest that may not be of
558 interest to journal publishers. We encourage readers to be cautious when evaluating the relative
559 importance of the hypotheses presented in this report. Hypotheses not based on peer-reviewed
560 literature should be viewed with more skepticism. We present them because they represent the
561 newest thinking on POD issues and may stimulate productive discussions and new research.
562

563 **Species Background**

564
565 The apparently simultaneous declines of four pelagic Delta fish species in the early 2000s
566 (Sommer et al. 2007) were surprising because of the differences in their life histories and
567 differences in how each species utilizes Delta habitats. These differences suggested one or more
568 Delta-wide factors to be important in the declines. Thorough descriptions of the POD
569 phenomenon and the early stages of the POD investigation are available in Sommer et al. (2007).
570 Here, we briefly review the general life history of each of the four POD fish species based on
571 descriptions in Moyle (2002) as ecological background for understanding the remainder of the

572 report. Readers are referred to the papers cited in the remainder of this report and the species-
573 specific conceptual models for additional details.

574
575 **Delta smelt** is a slender-bodied fish typically reaching 60–70 mm standard length (SL) with a
576 maximum size of about 120 mm SL. Delta smelt is endemic to the upper San Francisco Estuary,
577 primarily the Delta and Suisun Bay. Delta smelt is generally associated with the low salinity
578 zone locally indexed by X2, which is the distance (in km) along the axis of the estuary from the
579 Golden Gate to the 2 psu isohaline measured near the bottom of the water column (Jassby et al.
580 1995). Delta smelt feed primarily on planktonic copepods, cladocerans, and amphipods. Delta
581 smelt is basically an annual species and spawns in freshwater in the Delta. Upstream migration
582 of maturing adults generally occurs in the late fall or early winter with most spawning taking
583 place from early April through mid-May (Bennett 2005). Larval delta smelt move downstream
584 with the tides until they reach favorable rearing habitat in the low salinity zone. Some
585 apparently remain in upstream reaches including the Cache Slough-Sacramento deepwater ship
586 channel region and the central Delta region year-round (Julio Adib-Samii, CDFG, and J. Hobbs,
587 UCD, unpublished data). A very small percentage of delta smelt is believed to live 2 years and
588 spawn in one or both years (Bennett 2005). Delta smelt was listed as a threatened species by
589 both the federal and state governments in 1993. Its status was changed to state endangered in
590 2009. A similar change to federal endangered status was recently determined to be “warranted
591 but precluded” (USFWS 2010).

592
593 **Longfin smelt** typically reach 90–110 mm SL with a maximum size of 120–150 mm SL. In the
594 San Francisco Estuary, longfin smelt generally occur in Suisun, San Pablo, and San Francisco
595 bays as well as in the Gulf of the Farallones, just outside San Francisco Bay. Longfin smelt is
596 anadromous and spawns at 2-years of age in freshwater portions of the Delta. A few fish over
597 110 mm SL are captured annually during spawning migrations, but it’s uncertain whether these
598 are third-year fish or fast growing fish in their second year. Most spawning takes place from
599 December through March. Incubation takes about a month and larvae are buoyant at hatching.
600 Larval longfin smelt are transported downstream with outflow-generated surface currents. As
601 they develop and can control their depth distribution, larvae move lower in the water column and
602 probably remain in favorable rearing habitat in the low salinity zone and farther downstream. As
603 juveniles grow and become more mobile they disperse downstream into Suisun, San Pablo and
604 San Francisco bays. Longfin smelt feed primarily on mysid shrimp including the non-native
605 *Hyperacanthomysis longirostris* and the native *Neomysis mercedis*. Copepods and other
606 crustaceans can also be important food items, especially for smaller fish. Although other
607 populations of longfin smelt occur on the Pacific Coast, the San Francisco Estuary population is
608 the southern-most reproducing population and was recently proposed for listing under CESA,
609 and as a distinct population segment of the species under FESA (The Bay Institute et al. 2007a,
610 b). In April 2010, the California Fish and Game Commission listed longfin smelt as threatened.
611 The USFWS determined that the longfin smelt in the San Francisco estuary did not meet the
612 definition of a distinct population segment and the species was not listed (USFWS 2009).

613
614 **Striped bass** is native to the Atlantic Coast of North America. It was introduced to California in
615 1879. Striped bass is a large (> 1 m), long-lived (> 10 years) species. Striped bass juveniles and
616 adults are widespread in the San Francisco Estuary watershed. The species can be found in the
617 larger river systems downstream of impassable dams and in the coastal ocean. Striped bass is a

618 generalist predator. Larval and small post-larval bass feed mainly on copepods. As the fish
619 grow they consume larger invertebrates and fishes. Striped bass is an anadromous mass
620 spawner. Females begin spawning at 4–6 years of age and can spawn every year depending on
621 environmental conditions. In the Sacramento River, spawning occurs in the upper parts of the
622 tidal portion of the river and lower portions of the riverine portion of the river. A small
623 proportion of spawning occurs in the San Joaquin River within the Delta. Embryos and larvae
624 drift into the Delta and larval striped bass are associated with the low salinity zone. As the
625 juveniles grow, they disperse throughout the Delta and San Francisco Bay. Although an invasive
626 species, striped bass supports a popular and economically important recreational fishery.

627
628 **Threadfin shad** was introduced to California in 1954 to provide forage for predatory fish in
629 reservoirs. It was planted in the San Francisco Estuary watershed in 1959 and rapidly colonized
630 all available habitat. Adult threadfin shad are typically less than 100 mm total length and
631 primarily inhabit freshwater, where they are omnivorous filter feeders on phytoplankton,
632 zooplankton, and detritus. They can also selectively sight-feed on individual organisms,
633 primarily zooplankton. Most threadfin shad spawn as 2-year olds, although some may spawn at
634 the end of their first year. Spawning occurs from April to August, but most occurs in June and
635 July. Larval and juvenile threadfin shad are mainly found in the freshwater portions of the Delta.
636 Threadfin shad is the most abundant pelagic fish in the upper estuary and is exceptionally
637 important as prey for piscivorous species.

638
639 The apparently simultaneous declines of these four species of fish were surprising because of the
640 differences in their life histories and differences in how each species utilizes Delta habitats.
641 These differences suggested one or more Delta-wide factors to be important in the declines.
642

643 **Conceptual Models**

644
645 All conceptual models developed during the POD investigation revolve around natural and
646 anthropogenic drivers that affect ecological change such as the observed pelagic fish declines.
647 Drivers and stressors are events or processes involving environmental (natural or human)
648 variables and the terms are often used interchangeably or with fluid boundaries. Some authors
649 define drivers as larger-scale and/or external influences on ecosystems, while stressors are
650 defined as smaller-scale, internal changes brought about by the larger-scale drivers (e.g. Ogden
651 et al 2005). Others define stressors as drivers that “exceed the range of variation beyond which
652 the current biological communities can survive (commonly taken as exceeding the long-term or
653 reference range of variation” (Miller et al. 2010). Here, we make the somewhat arbitrary decision
654 to refer only to drivers because the term “stressor” implies an adverse or “undesirable” changes
655 in ecosystem structure and functions that may be directly responsible for the POD phenomenon.
656 Given that two species might respond in opposite ways to the same stressor, the meaning of the
657 term becomes confusing. The distinction between drivers and stressors has become more
658 important in the most recent POD work on regime shifts and ecological thresholds, as will be
659 described in later sections of this report.

660
661 Based on the observation that fish abundance indices declined abruptly after 2000, the POD-MT
662 developed an initial conceptual model about drivers potentially responsible for the declines (IEP

663 2005). After the completion of the first set of studies in late 2005, alternative models were
664 developed based on the results available at that time and the consensus of professional judgment
665 of the POD-MT regarding the extent to which individual drivers were likely to have affected
666 each species-life stage during recent years. The nine drivers evaluated were: (1) mismatch of
667 larvae and food; (2) reduced habitat space; (3) adverse water movement/transport; (4)
668 entrainment; (5) toxic effects on fish; (6) toxic effects on fish food items; (7) harmful
669 *Microcystis aeruginosa* blooms; (8) *Corbula amurensis* effects on food availability; and (9)
670 disease and parasites.

671
672 These earlier conceptual models provided a useful way to: (1) summarize understanding of
673 factors that may have contributed to the POD and (2) design the initial suite of research studies;
674 however, they had several shortcomings. They did not adequately reflect spatial and temporal
675 variation in the nine drivers evaluated, new data showed several assumptions to be incorrect, and
676 the initial models were relatively cumbersome.

677
678 The basic conceptual model developed in 2006 (Sommer et al. 2007) and species-specific models
679 developed in 2007 (Baxter et al. 2008) represented an improvement over the earlier efforts, but
680 still had numerous limitations. Many of the results and inferences were preliminary at the time
681 they were developed and had not been peer-reviewed, so the models were considered
682 preliminary. Moreover, the models may have been influenced by potential biases in the
683 sampling programs. Changes in the size and distribution of the target species have the potential
684 to change our perception of trends in abundances and distributions. This, in turn, would affect
685 our conceptual models. Throughout this discussion we use indices of fish abundance such as the
686 Fall Midwater Trawl and Summer Towntnet survey (TNS) indices or catch per unit effort (CPUE)
687 as estimates of abundance. The relationship between these indices of abundance and the actual
688 population size of any species are generally not known, but they are presumed to increase
689 monotonically with each other. Modeling efforts along with sampling gear evaluation studies are
690 underway to better understand these relationships (e.g., Newman 2008). The same caveats
691 continue to apply to the refined and new conceptual models presented in this report.

692
693 We also recognize that the recent decline in pelagic fish species is superimposed over long-term
694 declines for several of these species and their long-term relationships with other environmental
695 variables. Initial change-point analyses (Manly and Chotkowski 2006) suggested that distinctly
696 different statistical models might be appropriate for different time periods. A clear line divides
697 the POD era from the years preceding it for some species. Bayesian change-point modeling
698 indicated that there were long-term declines in all 4 POD species abundances related to long
699 term changes in some environmental variables, such as water clarity. However, the step decline
700 in abundance of the species in the early 2000s (Thomson et al. 2010) could not be explained by
701 changes in those same environmental variables or any of the other variables considered in the
702 analysis. There also appear to be multiple periods of decline within the historical record
703 preceding the POD and the periods are not always the same across species. The basic and
704 species-specific conceptual models explicitly focus on mechanisms that might have contributed
705 to the decline of pelagic fishes during the POD era; however, the historical antecedents to the
706 POD are a crucial part of the story. These historical changes are more explicitly acknowledged
707 and explored in the new regime shift model; however, this model is still in the early stages of
708 development and contains many uncertainties.

709
710 *Note that none of the conceptual models are intended to exclude other explanations for the*
711 *observed changes in fish abundance, nor are they designed to set priorities for resource*
712 *management.* Instead, they are intended as examples of how different drivers may be linked and
713 produce ecological responses. Moreover, no single model component can explain the declines of
714 all four species (Thomson et al. 2010, Mac Nally et al. 2010). We will continue to further
715 develop and refine the models as additional data become available. In this report, as in previous
716 reports, we use the basic conceptual model as an organizing principle for an examination of the
717 historical development of thought on various topics we believe important to POD species. We
718 use the individual species models to present the suite of drivers we think is important for that
719 particular species. We use the regime shift model to put the POD into a more historical
720 ecosystem context and set the stage for a new phase of the POD investigation.
721

722 **Basic Conceptual Model**

723
724 The basic conceptual model (Figure 3) is rooted in classical food web and fisheries ecology and
725 contains four major components: (1) prior fish abundance, which posits that continued low
726 abundance of adults leads to low juvenile production (i.e., stock-recruitment effects); (2) habitat,
727 which posits that the amount of water (volume or surface area) with suitable conditions for a
728 species has changed because changes in estuarine water quality variables, disease, and toxic algal
729 blooms in the estuary affect survival and reproduction; (3) top-down effects, which posits that
730 predation and water project entrainment affect mortality rates; and (4) bottom-up effects, which
731 posits that consumable resources and food web interactions affect survival and reproduction.
732 Each model component contains one or more potential drivers affecting the POD fishes. The
733 overlap of the four model components indicates that they can affect the fish species
734 simultaneously and can interact to produce synergistic or antagonistic effects. For each model
735 component, our working hypotheses during development of the model were: (1) one or more
736 drivers associated with the component were responsible for an adverse change at the time of the
737 POD; and (2) this change resulted in a population-level effect. However, the emphasis of the
738 POD effort has evolved from concern with changes that took place at the time of the POD to
739 understanding drivers and drivers affecting the abundance of POD species, regardless of time
740 frame.
741

742 In the following sections, we present a synopsis of how thinking has evolved on each driver
743 within the four major components and the current thinking of the POD-MT about how that topic
744 relates to POD species. The intent of this approach is to provide readers with a perspective on
745 the evolution of scientific knowledge on specific topics and how the POD-MT has utilized
746 historical and emerging knowledge to guide the POD investigations.
747

748 *Previous Abundance*

749
750 The relationship between numbers of spawning fish and the numbers of young subsequently
751 recruiting to the adult population is known as a stock-recruitment relationship. Stock-
752 recruitment relationships have been described for many species and are a central part of the

753 management of commercially and recreationally fished species (Myers et al. 1995). Different
754 forms of stock-recruitment relationships are possible, including density-independent, density-
755 dependent, and density-vague types. The latter refers to situations where there is not a
756 statistically demonstrable stock-recruitment relationship observable in available data.
757

758 Unfortunately, none of the POD species were of sufficient interest as commercial or recreational
759 species to warrant development of stock-recruitment models until the 1970s. There was a
760 commercial fishery for striped bass from 1888 until 1935, when it was closed and only
761 recreational fishing was allowed (Dill and Cordone 1997). The sport fishery was subsequently
762 managed through regulations alone with no sampling or population modeling. Regulations
763 became more restrictive as catch per angler indicated that the population was declining (Skinner
764 1962). Studies initiated in 1959 in connection with the CVP and SWP led to the first estimates
765 of population size. The population was estimated at 2–3 million legal-sized adult fish in the
766 early 1960s (Herbold et al. 1992) with a plateau of an estimated 1.5–1.9 million legal-sized adult
767 fish from 1969–1976 (Herbold et al. 1992, Kohlhorst 1999). The other species were never
768 extensively harvested, though longfin smelt was once part of the “whitebait” catch from San
769 Francisco Bay (Skinner 1963), or records were not kept if they were harvested.
770

771 The initiation of fisheries studies in 1959 in association with the planning and operation of the
772 CVP and SWP provided the first opportunity to develop stock-recruitment type models for
773 striped bass, longfin smelt, and delta smelt. We deliberately say “stock-recruitment type
774 models” because classical stock-recruitment models in fisheries management are based on and
775 produce estimates of the actual population size of a species as either numbers or biomass. The
776 IEP monitoring was not designed to produce an actual population estimate (with the exception of
777 Petersen tag estimates for striped bass and sturgeon), but to provide information on trends based
778 on a population index or CPUE (e.g., number per trawl). Efforts are currently under way to
779 develop methods to calculate population sizes of delta smelt and longfin smelt from currently
780 collected data (e.g. Newman 2008).
781

782 Early models related the success of spawning and juvenile recruitment of striped bass with Delta
783 outflow using simple regression models (Stevens and Miller 1983, Stevens et al. 1985). These
784 models were not true stock-recruitment models (nor intended as such) because they assumed that
785 the environmental variable flow was the primary control on recruit abundance. In other words,
786 stock is always sufficient to produce high recruitment given good environmental conditions and
787 good environmental conditions are provided by high flow. These early models began to break
788 down in the late 1970s (Stevens et al. 1985) leading to new work on striped bass and eventually
789 on the other POD species.
790

791 Kimmerer et al. (2000) found evidence for a density-dependent stock-recruitment relationship in
792 San Francisco Estuary striped bass. However, the adult striped bass stock is currently not
793 particularly low (Figure 9), so stock size is not likely a mechanism contributing to recent very
794 low age-0 striped bass abundance. In other words, there appear to be enough adults in the
795 system to produce sufficient young for the population to recover. Recent analyses have revealed
796 a change in sex ratio for adults on their spawning migration. The proportion of female striped
797 bass in the spawning population appears to have been declining in recent years (Figure 10, T.
798 Sommer, DWR, unpublished data). The causes for this change are unknown (discussed below)

799 as are the consequences of the change for the population biology of the species, but at the very
800 least the effective spawning stock has been limited by the low female numbers in the spawning
801 runs. The decline does not seem to be an artifact of sampling bias because it has been noted in
802 two types of sampling gear (gill nets and fyke traps) in two widely separated locations (the Delta
803 and the Sacramento River near Knights Landing, respectively) (T. Sommer, DWR, unpublished
804 data unpublished data).

805
806 There is little evidence for strong stock-recruitment or stage-recruitment relationships for
807 threadfin shad (Feyrer et al. 2009). For longfin smelt, preliminary analyses support a stock-
808 recruitment relationship between adults approaching their second birthday and age-0 fall recruits
809 (The Bay Institute et al. 2007a). A significant stage-recruitment relationship (fall age-0 to fall
810 age-1 abundance) also exists, but survival declined after 1994 (Rosenfield and Baxter 2007)
811 presumably due to continued food limitation, which in turn may have led to a habitat shift
812 (discussed below). However, current populations of longfin smelt and threadfin shad are similar
813 to low populations observed in previous years (Figure 2). Threadfin shad rebounded fully from
814 previous abundance lows in the 1970s and 1980s. Longfin smelt populations rebounded
815 somewhat in the 1990s following previous lows during the 1987–1992 drought. Recovery of
816 these species is only expected if the factors affecting recruitment have not changed substantially.
817 If the factors affecting survival from egg to adult have changed substantially since the beginning
818 of the POD, then recovery might not occur even though recovery from low abundance occurred
819 in the past. The changes in the statistical relationships between outflow and population
820 abundance indices for longfin smelt and age-0 striped bass (Figure 11) indicate that changes in
821 the drivers of recruitment have occurred. These changes are discussed in more detail in
822 subsequent sections.

823
824 Population size can potentially affect survival at multiple points over the life cycle of a species.
825 Typical examples include stock relationships, where the number of adults influences the number
826 of offspring, and cohort relationships, where the abundance of one life stage affects survival to
827 the following stage. In any form of a stock-recruitment model, there is a point at which low
828 adult stock will result in low juvenile abundance and subsequent low recruitment to future adult
829 stocks. This can occur even under favorable environmental conditions while the stock “rebuilds”
830 itself. From a stock-recruitment perspective, the present low abundance of delta smelt is of
831 particular concern. The current population is smaller than at any time previously in the record
832 (Figure 2). The delta smelt stock-recruitment relationship appears to be density-independent,
833 particularly since the latter 1970s (Figure 12). The Fall Midwater Trawl (FMWT) index (pre-
834 spawning adults) explains about half of the variation in the TNS index (juveniles) over the period
835 of 1978 to present. Feyrer et al. (2007) found that incorporating fall salinity increased the
836 explanatory power of this stock-recruitment relationship during this period. As noted above,
837 transitions between life stages (i.e. cohort relationships) can also influence survival. The life
838 history transition between the TNS index (juveniles) and the following FMWT index (pre-
839 spawning adults) (Figure 12) has been influenced by density-dependent mechanisms and
840 declining carrying capacity (Bennett 2005). In a plot of this relationship, the most recent POD
841 years (2005–2009) plot well below the relationships based on the entire record (Figure 13). A
842 plot of adult production from the adults in the previous year, represents an interaction of the
843 above relationships and any other factors affecting survival. This plot suggests declining
844 survival of delta smelt throughout the POD period (Figure 14).

845
846 The combined effects of stock-recruitment and cohort relationships are somewhat complicated to
847 assess without a proper life cycle model. However, the change in the relationship in summer to
848 fall survival (Figure 13) suggests that the primary factors affecting juvenile survival recently
849 changed and shifted to earlier in the life cycle; however, this would not necessarily affect the
850 stock-recruitment relationship. It would only affect available stock. In other words, an
851 individual adult delta smelt might still produce the same number of young; however, fewer
852 young survive to reproduce as adults. In addition, the fecundity of adult smelt has likely
853 changed. The mean size of adult delta smelt has declined since the early 1990s (Sweetnam
854 1999), possibly due to changes in the food web (see Bottom up section). There may also be
855 selection for late-spawned larvae as a result of water export schedules related to the Vernalis
856 Adaptive Management Plan (VAMP), which began in 2000 (Bennett et al. 2008). These smaller
857 larvae have less opportunity to grow to large adult size. Smaller adults due to reduced food
858 supplies and younger adults due to selection for late-spawns are not mutually exclusive
859 mechanisms and the combination could easily have a nonlinear impact on overall fecundity.

860
861 These observations strongly suggest that recent population trends for delta smelt are outside the
862 historical realm of variability and may be associated with a new state of the system (see Regime
863 Shift Model section). This inference is supported by a recent changepoint analysis, which
864 indicated a decline in abundance in the early 2000s independent from environmental variables
865 that previously explained abundance. Thus, recovery is likely to require changes in the drivers
866 that have produced the current low levels of abundance and perhaps new drivers or previous
867 drivers that have since become more important.

868
869 Given the unprecedented low abundance of delta smelt since 2000 (Figure 2), serious
870 consideration should be given to evaluation of Allee effects. Allee effects occur when
871 reproductive output per fish declines at low population levels (Berec et al. 2006). In other
872 words, below a certain threshold the individuals in a population can no longer reproduce rapidly
873 enough to replace themselves and the population, exhibiting inverse density dependence, spirals
874 to extinction. For delta smelt, possible mechanisms for Allee effects include processes directly
875 related to reproduction and genetic fitness such as difficulty finding mates, genetic drift, and
876 inbreeding (Gascoigne et al. 2009). Other mechanisms related to survival such as increased
877 vulnerability to predation (Gascoigne and Lipcius 2004) are also possible and will be briefly
878 discussed in the Top-Down model component section. While theoretical work suggests that
879 Allee effects might be common in nature, empirical evidence for Allee effects in natural
880 populations remains sparse, possibly because they are often masked by measurement errors
881 (Gregory et al. 2010). In addition, the interactive effects of multiple Allee effects may have
882 important implications for species conservation, but have not yet been well explored in ecology
883 (Berec et al. 2006).

884

885 *Habitat*

886

887 According to Hudson et al. (1992), “habitat is simply the place where an organism lives....
888 Physical, chemical and biological variables (the environment) define the place where an
889 organism lives.” Hayes et al. (1996) explain further that “space is the primary component of fish

890 habitat, and other resources and environmental conditions modify the utility of space.” The
891 maintenance of usable space (i.e. suitable habitat quality) is essential to the long-term health of
892 aquatic resources (Rose 2000, Peterson 2003). For the habitat component of the model, a key
893 point is that *habitat suitability affects all other components of the model*. This is indicated by the
894 overlap of habitat with all other components in Figure 3. Hence, changes in habitat not only
895 affect pelagic fishes, but also their predators and prey, which, in turn, can also have effects on
896 the habitat they occupy. Although not a focus of this report, we expect that the habitats of the
897 POD species are especially vulnerable to future climate change. Thus, policy and management
898 alternatives regarding habitat should consider expected changes in climate as well as other
899 changes in major system drivers such as altered water management.

900
901 *Habitat for pelagic fishes:* Habitat for pelagic fishes in the estuary is open water, largely away
902 from shorelines and vegetated inshore areas except perhaps during spawning. This includes
903 large embayments such as Suisun Bay and the deeper areas of many of the larger channels in the
904 Delta. More specifically, estuarine pelagic fish habitat is water with suitable values for a variety
905 of physical-chemical properties (e.g., salinity, turbidity, and temperature), suitably low levels of
906 contaminants, and suitably high levels of prey production to support growth. A key to
907 understanding pelagic fish habitat in the estuary is recognizing that it is not fixed to a specific
908 geographic location. Freshwater and seawater will always meet somewhere in an estuary and
909 this zone will move within the estuary at time scales ranging from annually and seasonally in
910 response to freshwater outflow to hourly in response to tides. However, the geographic locality
911 where the interface between freshwater and saltwater occurs at any particular time may have
912 implications for the effects of other factors (e.g., exposure to a point source of contaminants).
913 Thus, pelagic fish habitat suitability at any specific geographic point in the estuary can be
914 strongly influenced by variation in freshwater flow (Jassby et al. 1995, Bennett and Moyle 1996,
915 Kimmerer 2004).

916
917 We know that aquatic habitats in the Delta have changed substantially since the mid-1850s. The
918 Delta has been converted from a complex mix of seasonal wetlands, perennial wetlands and
919 riparian habitats with many dead-end waterways to a geographically simplified system of leveed
920 agricultural islands separated by a network of interconnected channels, which are often deeper
921 and more steep-sided than natural channels and have hardened, rip-rapped levee shorelines. It
922 has been estimated that 95% of wetlands in the Delta have been lost (The Bay Institute 1998).
923 Large pulses of sediment associated with hydraulic mining have now largely moved through the
924 estuary and dam construction has prevented recruitment of new sediments from upstream areas.
925 This has contributed to increasing water clarity in recent years (Wright and Schoellhamer 2004).
926 The effects of many of the early habitat changes are largely unknown. There is speculation that
927 the success of the striped bass introduction in 1879 was partially due to their semibouyant
928 embryos and pelagic larvae that were better adapted than those of native species to the high
929 sediment loads carried in the rivers at that time due to hydraulic mining (Moyle 2002). The
930 changes in channel configuration, water storage behind dams, and operations of the CVP, SWP,
931 and other water management projects have resulted in complex hydrodynamic and water quality
932 patterns (Kimmerer 2004, Monsen et al. 2007, Healey et al. 2008, Kimmerer and Nobriga 2008)
933 that are very different from what existed in the past (e.g., Brown and Bauer 2009). However,
934 only recently have these changes become of major management concern (ROD 2000). This
935 concern has resulted in tightened restrictions on water diversions, passage of California water

936 legislation in the 2009, the development of new flow criteria for the Sacramento-San Joaquin
937 Delta ecosystem by the SWRCB, and the Bay-Delta Conservation Plan (BDCP) planning
938 process. Kimmerer (2004) has summarized much of the existing knowledge on factors affecting
939 open water habitats.

940
941 Development of habitat-based ideas related to Delta fish populations basically began with the
942 simple regression models, described in the previous section, which related Delta outflow to
943 recruitment of young pelagic fishes of several species (Stevens and Miller 1982, Stevens et al.
944 1985). The next step was the development of X2 as an index of the response of the estuarine
945 community to net freshwater flow (Jassby et al. 1995). X2 denotes the distance (in km) along
946 the axis of the estuary from the Golden Gate to the 2 psu isohaline measured near the bottom of
947 the water column. X2 has been linked to the success of various species including POD species.
948 X2 is related to flow but provides a more direct link to fish habitat (Kimmerer 2002a, b) and has
949 thus been used as a habitat indicator and adopted for regulatory purposes. Unfortunately, the X2
950 relationships for longfin smelt and age-0 striped bass have changed in association with other
951 drivers (e.g., invasive species). Delta smelt has never shown a strong relationship to X2, and the
952 freshwater threadfin shad has no relationship to X2. As a result of the POD, several
953 investigations were started to define appropriate habitat for various POD species so changes in
954 habitat suitability could be better assessed.

955
956 Several of the POD fishes use tidally-assisted swimming behaviors to maintain themselves
957 within open-water areas where water quality and food resources are favorable (Bennett et al.
958 2002). The four POD fishes also distribute themselves at different values of salinity within the
959 estuarine salinity gradient (e.g., Dege and Brown 2004, Feyrer et al. 2007), so at any point in
960 time, salinity is a major factor affecting their geographic distributions. As mentioned earlier,
961 pelagic habitat suitability in the San Francisco Estuary can be characterized by changes in X2.
962 The abundance of numerous taxa increases in years when flows into the estuary are high and the
963 2 psu isohaline is pushed seaward (Jassby et al. 1995, Kimmerer 2002a, b), implying that the
964 quantity or suitability of estuarine habitat increases when outflows are high. Recent analyses
965 indicated that neither changes in area or volume of low salinity water (habitat) account for this
966 relationship for species showing relationships with X2, except for striped bass and American
967 shad (Kimmerer et al. 2009). This suggests that X2 is indexing other environmental variables or
968 processes rather than simple extent of habitat.

969
970 It is also worth noting that the change from net outflow models to more complex habitat models
971 occurred in concert with a change in thinking about Delta hydrodynamics. Through the 1980s
972 the perception was that river-like net flow through Delta channels was the major feature of flow
973 important to the ecosystem even though it was recognized that the Delta was a tidal system.
974 From the 1990s onward there has been increasing recognition that tidal flows combined with the
975 highly altered nature of the Delta channel network combine to create complex hydrodynamic
976 patterns. These patterns can result in sometimes surprising results regarding transport of weakly
977 swimming fishes, other organisms, and water quality constituents (e.g., salt) through the Delta.
978 For example, Lucas et al. (2002) and Lopez et al. (2006) showed that the production and
979 distribution of phytoplankton biomass can be highly variable within and between nearby habitats
980 of the same type in the Delta due to variations in phytoplankton sources, sinks, and transport.
981 Therefore, superficially similar, geographically proximate habitats can function very differently

982 and assessment of the role and functioning of different habitat types needs to include a regional
983 landscape perspective (Lopez et al. 2006, Cloern 2007). This perspective could be essential in
984 predicting the success of habitat restoration efforts (Lopez et al. 2006). Kimmerer and Nobriga
985 (2008) used a particle tracking model to explore the relationships between hydrodynamics and
986 entrainment risk and found release location and hydrology to be important factors. This latter
987 application is especially relevant to entrainment issues discussed in a later section.

988
989 Based on a 36-year record of concurrent midwater trawl and water quality sampling, there has
990 been a long-term decline in fall habitat suitability for delta smelt and striped bass, but not for
991 threadfin shad (Feyrer et al. 2007). The long-term habitat suitability declines for delta smelt and
992 striped bass are defined by a lowered probability of occurrence in samples based on changes in
993 specific conductance (a surrogate for salinity) and Secchi depth (a measure of water clarity or,
994 conversely, turbidity). Notably, delta smelt and striped bass habitat suitability declined recently
995 coinciding with the POD. The greatest changes in habitat suitability occurred in Suisun Bay and
996 the San Joaquin River upstream of Three Mile Slough and in the southern Delta (Feyrer et al.
997 2007). There is evidence that these habitat changes have had population-level consequences for
998 delta smelt. The inclusion of specific conductance and Secchi depth in the delta smelt stock-
999 recruitment relationship described above improved the fit of the model, suggesting adult
1000 numbers and their habitat conditions exert important influences on recruitment. Subsequently,
1001 Feyrer et al. (2010) developed a model for estimating habitat suitability based on X2.

1002
1003 The importance of salinity in this study was not surprising, given the relationships of population
1004 abundance indices with X2 for many species. Fall X2 has been high during the POD years
1005 despite moderate to high outflow conditions during the previous winter and spring of most years.
1006 Contra Costa Water District (2010) recently reviewed and summarized a wide range of historical
1007 reports and analyses regarding salinity in the western Delta and Suisun Bay. The basic
1008 conclusion was that a variety of human activities including channelization of the Delta,
1009 elimination of tidal marsh, construction of deepwater ship channels and diversions of water have
1010 contributed to increased salinity in the region. Consistent with fall X2 observations, the report
1011 also noted that fall salinity in 21 of the last 25 years has resembled that expected during a
1012 drought, even though half the years have been relatively wet. Although the operations of
1013 reservoirs and water diversions have been able to ameliorate the effects of salinity on water
1014 supply, salinities still exceed pre-1900 levels.

1015
1016 There appeared to be a curious anomaly in the salinity distribution of delta smelt collected during
1017 the September 2007 survey of the FMWT. All seven delta smelt collected during this survey
1018 were captured at statistically significant higher salinities than what would be expected based
1019 upon the relationship generated by Feyrer et al. (2007) ($p=0.0012$, K. Newman, USFWS,
1020 personal communication 2007). There could be any number of reasons why this occurred,
1021 including a substantial *Microcystis* bloom in the western Delta in 2007, which extended further
1022 downstream than previous blooms and may have affected the distribution of biological
1023 organisms.

1024
1025 The importance of Secchi depth in the long-term changes in pelagic fish habitat suitability
1026 (Feyrer et al. 2007) was more surprising. Unlike salinity, interannual variation in water clarity in
1027 the Delta is not primarily a function of flow variation (Jassby et al. 2002). Water clarity in the

1028 Delta has been increasing since routine monitoring began in 1975 (Jassby et al. 2002, Wright and
1029 Schoellhamer 2004, Jassby et al. 2005, Jassby 2008). The primary mechanisms suggested to
1030 explain the increasing water clarity are: (1) reduced sediment supply due to dams in the
1031 watershed (Wright and Schoellhamer 2004); (2) sediment washout from very high inflows during
1032 the 1982–1983 El Nino (Jassby et al. 2005); and (3) biological filtering by submerged aquatic
1033 vegetation (SAV) (Brown and Michniuk 2007, Hestir 2010).

1034
1035 Results from a recently completed POD-funded study (Hestir 2010) indicate that the three
1036 mechanisms likely played sequential roles in the increasing water clarity in the Delta. The initial
1037 increase in clarity was likely brought about by reductions in sediment supplies due to the first
1038 two mechanisms. The resulting clearer water facilitated the rapid spread of invasive submerged
1039 aquatic vegetation which in turn led to further clearing of the water. In lakes, high densities of
1040 the Brazilian waterweed *Egeria densa* and similar plants can mechanically filter suspended
1041 sediments from the water column (Scheffer 1999). Vegetation has also been shown to facilitate
1042 sedimentation in marshes and estuaries (Yang 1998, Braskerud 2001, Pasternack and Brush
1043 2001, Leonard et al. 2002). The most abundant invasive SAV species in the Delta is *E. densa*.
1044 *E. densa* invaded the estuary in the 1980s and rapidly expanded its distribution in the 1990s
1045 (Service 2007). *E. densa* continued to spread by expansion of existing patches and invasion of
1046 new areas during the POD years (Hestir 2010). Areal coverage of *E. densa* increased more than
1047 10% per year from 2004 to 2006; however, *Egeria* may now occupy most of the Delta habitat
1048 suited to its establishment and growth and the rate of expansion may have declined substantially
1049 over the past few years (E. Hestir, UCD, personal communication). The current growth and
1050 spread of SAV in the Delta is likely controlled by water velocities, rather than light availability
1051 (Hestir 2010). In clear water, *E. densa* can grow to depths of 6 m (Anderson and Hoshovsky
1052 2000). Salinity likely limits its spread into the seaward areas of the estuary (Hauenstein and
1053 Ramirez 1986). According to Hestir (2010), Delta SAV grows best at annual water velocities
1054 below 0.49 m/s and suppresses turbidity levels in its vicinity by reducing sediment resuspension.
1055 The expansion of invasive SAV in the Delta can explain 21–71% of the total increasing trend in
1056 water clarity in the Delta from 1975–2008. Although *E. densa* may have reached the current
1057 limits of its distribution in the Delta, it is possible that additional clearing due to further
1058 reductions in sediment supplies and water velocities will allow *E. densa* to spread into
1059 progressively deeper water and contribute to even more clearing.

1060
1061 The mechanisms causing the negative associations between water clarity and delta smelt and
1062 striped bass occurrence are currently under investigation. Based on research in other systems
1063 (e.g. Gregory and Levings 1998), Nobriga et al. (2005) hypothesized that higher water clarity
1064 increased predation risk for delta smelt, young striped bass, and other fishes typically associated
1065 with turbid water. A certain concentration of suspended particles also seems to be necessary for
1066 proper feeding by young delta smelt (Baskerville-Bridges et al. 2002, Mager et al. 2004).
1067 Turbidity, and to a lesser degree salinity, were also found to be extremely important parameters
1068 influencing larval delta smelt survival in laboratory flow-through assays exposing larval delta
1069 smelt to water collected at Delta and Suisun Marsh sites (Werner et al. 2010). Increased turbidity
1070 associated with the “first flush” after winter storm events may also play a role in the upstream
1071 spawning migration of delta smelt in the winter (Grimaldo et al. 2010, Sommer et al. in review).
1072 However, it is currently unknown if it acts as a migration cue, an important habitat attribute

1073 preventing predation during the upstream migration, or is simply correlated to other relevant
1074 variables.

1075
1076 Trends in habitat suitability for delta smelt differ during the summer period compared to the fall.
1077 Specific conductance, Secchi depth, and water temperature all significantly predict delta smelt
1078 occurrence in summer, suggesting they all interact to affect delta smelt distribution (Nobriga et
1079 al. 2008). However, none of the water quality variables were correlated with delta smelt
1080 abundance (as indexed by the TNS) at the scale of the entire estuary (Nobriga et al. 2008).
1081 Based on these habitat variables, Nobriga et al. (2008) identified three distinct geographic
1082 regions that had similar long-term trends in the probability of delta smelt occurrence. The
1083 primary habitat region was centered on the confluence of the Sacramento and San Joaquin rivers
1084 near Sherman Island; delta smelt relative abundance was typically highest in the confluence
1085 region throughout the study period. There were two marginal habitat regions, one centered on
1086 Suisun Bay where specific conductance was highest and delta smelt relative abundance varied
1087 with specific conductance, with lower abundance at higher conductance. The third region was
1088 centered on the San Joaquin River and the southern Delta. The San Joaquin River region had the
1089 warmest water temperatures and the highest water clarity. Water clarity increased strongly in
1090 this region during 1970–2004. This is also the region most heavily invaded by *E. densa* (Hestir
1091 2010). In the San Joaquin River region, delta smelt relative abundance was correlated with water
1092 clarity; catches declined rapidly to zero from 1970–1978 and remained consistently near zero
1093 thereafter. Note that in a year of low outflow, the low salinity zone would be at the confluence
1094 and summer delta smelt habitat might collapse to a restricted area around the confluence because
1095 salinity of the Suisun Bay region might increase to levels unsuitable for delta smelt. These results
1096 support the hypothesis that basic water quality parameters are predictors of summer delta smelt
1097 relative abundance, but only at regional spatial scales. These regional differences are likely due
1098 to variability in habitat rather than differences in delta smelt responses. Water management
1099 operations are targeted on keeping the lower Sacramento and San Joaquin rivers fresh for water
1100 exports so the range in salinity is probably relatively smaller than the range in turbidity. In the
1101 Suisun Bay region, there is a wider range of salinities relative to the other regions, so a response
1102 to that variable is possible.

1103
1104 Although Nobriga et al. (2008) recognized water temperature as an important aspect of delta
1105 smelt summer habitat, specific conductance and water clarity appeared to have more explanatory
1106 value. However, that does not mean that temperature is never important. Bennett et al. (2008)
1107 observed patterns in liver glycogen depletion and single cell necrosis in liver tissue of delta smelt
1108 collected in 2005 that were consistent with fish being stressed by warm water temperatures (22–
1109 23°C). As of yet, there is no strong evidence that delta water temperatures are increasing in
1110 response to climate change. Jassby (2008) detected a slight upward trend (0.2°C) in March–June
1111 water temperature (1996–2005) in the Delta but not in Suisun Bay. There was not a statistically
1112 significant increase in annual water temperatures if a longer time interval was considered (1975–
1113 2005), although there were some significant monthly trends at some locations (e.g., near
1114 Stockton in the summer months). Flow did not have a strong effect on water temperature. Water
1115 temperature in the Delta is mainly driven by air temperature and statistical models have been
1116 constructed for monitoring stations with a sufficient period of record. When these models are
1117 used to estimate water temperatures under various scenarios of climate change, temperatures
1118 stressful or potentially lethal to delta smelt become more common (Wagner et al. submitted).

1119
1120 *Contaminants and Disease:* In addition to habitat changes associated with changes in the
1121 estuarine habitat characteristics discussed above, contaminants can also change habitat suitability
1122 and overall ecosystem functions and productivity through numerous pathways. While there has
1123 been recent progress on assembling and analyzing data on trends in contaminant concentrations
1124 in and loadings to the Delta (Kuivila and Hladik 2008, Johnson 2010), the effects of
1125 contaminants on Delta fishes and other organisms are still not well understood. Evaluation of
1126 direct and indirect toxic effects on the POD fishes of both man-made contaminants and natural
1127 toxins associated with blooms of *M. aeruginosa* (a cyanobacterium or blue-green alga) remains
1128 an important and active component of the POD investigation. The main indirect contaminant
1129 effect under investigation is inhibition of prey production. The current state of knowledge about
1130 effects of contaminants on the POD fishes has recently been summarized by Brooks et al. (in
1131 review). This work suggests that while acute contaminant toxicity is not a likely cause for the
1132 population declines, sublethal stress from metals, nutrient-rich effluents, *M. aeruginosa* blooms,
1133 and pesticides are all potential contributors to, but not the sole cause of, past and ongoing
1134 declines.

1135
1136 Concern over contaminants in the Delta is not new. There are, for example, long standing
1137 concerns related to metals such as mercury and selenium in the watershed, Delta, and Bay (Davis
1138 et al. 2003, Linville et al. 2002).

1139
1140 Mercury contamination is mainly a result of historic gold mining. Contamination occurred as a
1141 result of mining mercury, primarily in the Coast Ranges, and loss of mercury during gold mining
1142 operations in the Sierra Nevada. As a result, mercury is a nearly ubiquitous contaminant
1143 throughout the Bay, Delta, and watershed. A TMDL (total maximum daily load) was recently
1144 promulgated for the Delta by the CVRWQCB. However, despite the concerns for mercury in
1145 general, mercury has not been associated with POD declines.

1146
1147 Selenium (Linville et al. 2002) has had two major sources in the system. Selenium in
1148 agricultural drainage water in the San Joaquin River drainage is the most well known source
1149 because of the bird deformities documented at Kesterson Reservoir at the terminus of the never
1150 completed San Luis Drain (e.g., Ohlendorf 2002). There is no strong evidence that selenium
1151 from agricultural sources in the San Joaquin Valley and transported by the San Joaquin River
1152 have been a major problem for POD species in the estuary; however, future changes in water
1153 management and hydrodynamics could conceivably increase exposure of Delta biota to this
1154 source (R. Stewart, USGS, personal communication, 2009). Other sources of selenium were oil
1155 refineries along the shoreline of San Francisco Bay, primarily San Pablo Bay and Carquinez
1156 Strait. These sources were controlled as part of the Clean Water Act. There is no evidence that
1157 these sources of selenium affected POD species; however there were effects on some benthic-
1158 foraging species, including white sturgeon, Dungeness crab *Cancer magister* and Sacramento
1159 splittail (Stewart et al. 2004).

1160
1161 In the 1960s, municipal and industrial waste disposal was a major issue in the San Francisco
1162 Estuary, as it was elsewhere in the nation. Passage of the Clean Water Act in 1972 eventually
1163 led to upgraded water treatment and improved water quality conditions in the estuary and
1164 recovery of resident biota (e.g., Nichols et al. 1986, Hornberger et al. 2000). However, passage

1165 of the Clean Water Act did not lead to removal of all municipal and industrial contaminants, and
1166 in some cases, unexpected sources and consequences of contamination and interactions of urban
1167 and agricultural contaminants are only now beginning to emerge.

1168
1169 One example of an unexpected consequence of ongoing contamination that is currently receiving
1170 much attention is the newly emerging idea that municipal ammonium pollution can lead to
1171 suppression of phytoplankton production in the estuary. Ammonium is the ionized form of
1172 ammonia gas, which forms when ammonia is dissolved in water according to the following
1173 equilibrium: $\text{NH}_4^+ \leftrightarrow \text{NH}_3 + \text{H}^+$. The equilibrium between ammonium and un-ionized ammonia
1174 depends primarily on pH (more H^+ available at lower pH), and also on temperature and salinity.
1175 Ammonia can cause toxicity to organisms, including fishes. Ammonium is considered a plant
1176 nutrient.

1177
1178 In many ecosystems around the world, increasing nutrient loading from agricultural run-off and
1179 urban waste disposal has led to eutrophication with often catastrophic consequences for many of
1180 the resident biota. In contrast to many other estuaries, the San Francisco Estuary has been
1181 considered relatively resilient to nutrient pollution and has not shown many of the common
1182 symptoms of eutrophication such as enhanced algal growth (Nichols et al. 1986, Cloern 2001,
1183 Kimmerer 2004). From 1975 to 1995, the Delta has actually experienced a long-term decline in
1184 phytoplankton production (Jassby et al. 2002). However, this trend has started to reverse since
1185 the mid-1990s (Jassby 2008) and potentially harmful algal blooms consisting mostly of floating
1186 colonies of the cyanobacteria (blue green algae) *M. aeruginosa* have become a common and
1187 widespread occurrence in the Delta in the late summer and fall (Lehman et al. 2008, see also
1188 sections below for more details). The long-standing and widely accepted explanation for the
1189 overall low levels of water column phytoplankton production and biomass in the estuary has
1190 been the suppression of phytoplankton growth by low water clarity, losses due to intense benthic
1191 grazing pressure, and transport through the Delta to San Francisco Bay (Cloern 2001, Jassby et
1192 al. 2002, Lucas et al. 2009a). The consequences of low phytoplankton production on the Delta
1193 food web are presented in more detail in the Bottom-Up section below. New studies suggest that
1194 phytoplankton growth may at times also be inhibited by high ammonium concentrations in and
1195 upstream of Suisun Bay and that changes in nutrient loadings may have also affected
1196 phytoplankton species composition, with repercussions throughout the food web (Wilkerson et
1197 al. 2006, Dugdale et al. 2007, Jassby 2008, Glibert 2010, Parker et al. in review, R. Dugdale,
1198 CSUSF-RTC, unpublished data). The largest source of ammonium to the system is the
1199 Sacramento Regional Wastewater Treatment Plant (SRWTP) (Jassby 2008). The inhibitory
1200 effect of ammonium on phytoplankton was unexpected by the agencies regulating wastewater
1201 treatment plants and thus was not considered during the development of past permits.

1202
1203 In addition to the unexpected inhibitory effect on phytoplankton, ammonia, the unionized, toxic
1204 gas form of the nutrient ammonium, can also have direct toxic effects on fish and their food
1205 organisms (EPA 2009). Evaluation of possible direct toxicity effects of dissolved ammonia and
1206 ammonium on delta smelt indicated that ambient concentrations in the Sacramento River and
1207 Delta were not high enough to be acutely toxic to 55-day old delta smelt (Werner et al. 2009a,
1208 2010b). However, delta smelt appear to be more sensitive to ammonia than many other fishes
1209 (Werner et al. 2009b). A new molecular tool called a DNA microarray has recently been
1210 developed to evaluate sublethal contaminant effects on delta smelt (Connon et al. 2009), which

1211 might help elucidate chronic effects and causal mechanism of ammonia toxicity for delta smelt.
1212 Mortality and growth of the amphipod *Hyalella azteca* in Delta water samples collected
1213 biweekly over a 4-year period (2006–2010) was significantly and negatively correlated with
1214 ammonia and ammonium concentrations at several sites throughout the Delta and Suisun Marsh
1215 (Werner et al. 2010a). Some important food organisms of delta smelt such as the copepod
1216 *Pseudodiaptomus forbesi* may also be particularly sensitive to ammonia according to preliminary
1217 results from ongoing studies (Werner et al. 2010b). Further results and discussions of ammonium
1218 and ammonia dynamics and effects in the estuary can be found in Ballard et al. (2009), Foe
1219 (2009), Werner et al. (2009a, 2010b) and CVRWQCB (2010).

1220
1221 There is increasing concern regarding endocrine disrupting compounds (EDCs), which are often
1222 but not exclusively associated with wastewater discharges. Like ammonium, advanced
1223 wastewater treatment can lead to substantial removal of at least some classes of EDCs (Huang
1224 and Sedlak 2001). Some pesticides (e.g. synthetic pyrethroids) also have endocrine disrupting
1225 properties (Jin et al. 2010). EDC contamination of the environment can lead to disruption of
1226 sexual functions and changes in population sex ratios. Ongoing research shows a higher
1227 proportion of male Mississippi silversides *Menidia beryllina* at an urban sampling site near the
1228 discharge of a waste water treatment plant compared to an agricultural site in Suisun Marsh (S.
1229 Brander, UCD, personal com.). This study found estrogenic activity (i.e., feminization) at both
1230 sites, but significantly higher androgenic activity (i.e., masculinization) at the urban site. Lavado
1231 et al. (2009) conducted a survey of estrogenic activity in 15 agriculturally impacted waterways of
1232 the California Central Valley including the Delta and found that overall estrogenic activity was
1233 highest in water from their Delta sampling site and in water from a Napa River site. Teh (2007)
1234 found evidence of intersex (ovatestis, ovarian tissue within testes) in 9 of 65 male delta smelt
1235 (14%) collected from Delta and Suisun Marsh sites, but did not measure endocrine disrupting
1236 compounds or their activity in the water. While no firm conclusions about the effects of EDCs on
1237 the POD fishes can be drawn from these few lines of evidence, it seems clear that the role of
1238 EDCs in the Delta deserves further study.

1239
1240 The effects of pesticides, including herbicides and insecticides, are perhaps the largest unknowns
1241 relative to POD. Pesticides can cause acute toxicity (mortality) or chronic, sublethal effects.
1242 Agricultural applications of pesticides are well regulated and tracked in detail by the California
1243 Department of Pesticide Regulation (<http://www.cdpr.ca.gov/docs/pur/purmain.htm>). Urban
1244 applications are less well documented but county-wide sales information is available. However,
1245 there are many different chemicals applied and the relationship between the amount of a
1246 pesticide applied and the transport of that pesticide into waterways is not simple. Furthermore,
1247 there is no established long-term monitoring program for dissolved or sediment-bound pesticides
1248 currently in place in the Bay, Delta, or watershed. The existing evidence is weak, but a recent
1249 review of available information collected before and during the POD investigation (Brooks et al.
1250 in review) suggests that pesticides and other toxicants may affect POD fishes or other parts of the
1251 ecosystem.

1252
1253 Studies conducted before the POD investigations included studies of acute and chronic pesticide
1254 toxicity effects on phytoplankton, invertebrates, and fishes in the Delta and watershed.
1255 Phytoplankton growth rate may have occasionally been inhibited by high concentrations of
1256 herbicides (Edmunds et al. 1999). Toxicity to invertebrates was noted in water and sediments

1257 from the Delta and associated watersheds (e.g., Kuivila and Foe 1995, Giddings 2000, Werner et
1258 al. 2000, Weston et al. 2004). Undiluted drainwater from agricultural drains in the San Joaquin
1259 River watershed can be acutely toxic (quickly lethal) to fish and have chronic (long-term) effects
1260 on growth (Saiki et al. 1992). Evidence for mortality of young striped bass due to discharge of
1261 agricultural drainage water containing rice herbicides into the Sacramento River (Bailey et al.
1262 1994) led to new regulations for discharge of these waters. Sublethal effects of contaminants
1263 were detected in striped bass larvae even after these regulations were in place (Bennett et al.
1264 1995). Bioassays using caged fish (non-POD species) have revealed DNA strand breakage
1265 associated with runoff events in the watershed and Delta (Whitehead et al. 2004). Kuivila and
1266 Moon (2004) found that peak densities of larval and juvenile delta smelt sometimes coincided in
1267 time and space with elevated concentrations of dissolved pesticides in the spring. These periods
1268 of co-occurrence lasted for up to 2–3 weeks, but concentrations of individual pesticides were low
1269 and much less than would be expected to cause acute mortality.

1270
1271 We initiated several studies to address the possible role of contaminants and disease in the POD.
1272 The largest study centered on biweekly monitoring of ambient water toxicity over 4 years in two
1273 phases (2006–2007 and 2008–2009) with standard bioassays using the amphipod *Hyaella azteca*
1274 at 15 to 16 sites in the Delta, Suisun Bay, and the Napa River as well as additional monitoring of
1275 water toxicity to larval delta smelt during the spring months (Werner et al. 2008, Werner et al.
1276 2010a, Werner et al. 2010b). The study also included laboratory investigations on the
1277 comparative sensitivity of important fish and aquatic invertebrate species to chemical
1278 contaminants of concern in the Delta, development and application of *in situ* exposure systems,
1279 and studies to develop biomarker tools for fish species of special interest.

1280
1281 Significant amphipod mortality was observed in 5.6% of 693 ambient water samples collected in
1282 2006–2007 and in 0.5% of 752 samples collected in 2008–2009. The tests also included addition
1283 of the enzyme inhibitor piperonyl butoxide (PBO) which synergizes (increases) pyrethroid
1284 pesticide toxicity, but antagonizes (reduces) organophosphate (OP) insecticide toxicity. PBO
1285 additions significantly affected survival in 1.1% and 0.9% of all samples collected in 2006–2007
1286 and 2008–2009, respectively. Growth was affected by PBO addition in 10.1% and 13.3 % of
1287 ambient samples collected in 2006–2007 and 2008–2009, respectively. Some of the toxic water
1288 samples contained measurable amounts of pyrethroids, such as bifenthrin and esfenvalerate, OPs
1289 such as chlorpyrifos and diazinon, relatively high levels of ammonium and/or ammonia, and, in
1290 two cases, relatively high concentrations of dissolved copper. Numerous samples contained more
1291 than one contaminant. Water from sites in the lower Sacramento River had the highest total
1292 ammonia and ammonium concentrations. Water from the lower Sacramento River region had the
1293 largest number of acutely toxic samples and, a relatively high occurrence of PBO effects on
1294 amphipod growth. Werner et al. (2010a) concluded that ammonia and ammonium and/or other
1295 contaminants occurring in mixture with these, likely contributed to the observed toxicity to the
1296 test invertebrates and that toxicity to invertebrates is most common in the northern Delta.

1297
1298 Laboratory flow-through assays exposing larval delta smelt to ambient water samples were
1299 conducted in 2007–2009. Results from these assays indicated that water quality in the
1300 Sacramento River at Hood, in the northern Delta near Cache and Lindsey Sloughs, and in the
1301 southern Delta in the San Joaquin River near Stockton was at times unfavorable for larval delta
1302 smelt. Delta smelt survival was highest in water from Suisun Slough.

1303
1304 It is important to note that delta smelt and *H. azteca* mortality were not necessarily observed in
1305 the same water samples. While the *H. azteca* tests are very useful for detecting biologically
1306 relevant levels of water column toxicity (especially pyrethroids), interpretation of the *H. azteca*
1307 test results with respect to fish should proceed with great caution. In addition, results from
1308 controlled bioassays conducted with grab samples under laboratory conditions may not be
1309 readily transferrable to variable field conditions.
1310
1311 To address the limitations of laboratory bioassays, Werner et al. (2010b) used a novel *in situ*
1312 exposure system for simultaneous tests of multiple fish and invertebrate species at Hood on the
1313 Sacramento River and in Stockton on the San Joaquin River. The *in situ* devices renewed the test
1314 water continuously, exposing test organisms to the same water quality fluctuations that would be
1315 experienced by stationary organisms in the river. Six *in situ* exposure experiments with larval
1316 delta smelt, larval fathead minnow (*Pimephales promelas*), and the amphipod *H. azteca* were
1317 conducted in spring 2009. No toxicity to *H. azteca* or *P. promelas* was detected. Larval delta
1318 smelt survival was low in the treatment and in the control water. Werner et al. (2010b)
1319 concluded that larval delta smelt may be too sensitive to stresses of transport, salinity, and
1320 temperature to be suitable test organisms for *in situ* tank systems and recommended using
1321 surrogate species, such as rainbow trout, in future tests.
1322
1323 To assess the comparability and relevance to Delta species of test results from bioassays
1324 conducted with different organisms, Werner et al. (2010b) compared the contaminant sensitivity
1325 of standard test organisms, such as larval fathead minnow, the waterflea *Ceriodaphnia dubia* and
1326 the amphipod *H. azteca*, with important Delta species, such as larval delta smelt and the copepod
1327 *Eurytemora affinis*. They exposed these organisms to a series of chemical contaminants
1328 previously detected in Delta water samples, including copper, total ammonia, the OP insecticides
1329 chlorpyrifos and diazinon, and the pyrethroid insecticides bifenthrin, cyfluthrin, esfenvalerate,
1330 and permethrin. Larval delta smelt were 1.8 to >11 times more sensitive than fathead minnow to
1331 ammonia, copper, and all insecticides tested with the exception of permethrin. Invertebrates
1332 were more sensitive than the two fish species. The zooplankton species *E. affinis* and *C. dubia*
1333 were most sensitive to total ammonia, and *C. dubia* was most sensitive to copper. *C. dubia* was
1334 the most sensitive, and *E. affinis* the least sensitive to the tested organophosphates insecticides.
1335 *H. azteca* was the most sensitive to all tested pyrethroid insecticides.
1336
1337 Pyrethroid insecticides are of particular interest in the POD investigation because use of these
1338 insecticides has increased during the most recent decade (Amweg et al. 2005, Oros and Werner
1339 2005) as use of some OP insecticides has declined. Toxicity of sediment-bound pyrethroids to
1340 macroinvertebrates has been observed in watersheds upstream of the Delta (Weston et al. 2004,
1341 2005). In recent toxicity tests, urban source waters to the Delta were almost always toxic to *H.*
1342 *azteca* due to pyrethroid pesticides (Weston and Lydy 2010). Agricultural runoff samples were
1343 toxic less often. Similar to the findings for ammonium (Jassby 2008), the effluent from the
1344 SRWTP was the largest source of pyrethroids to the Delta in this study and *Hyalella azteca*
1345 always experienced high mortality rates and swimming impairment when exposed to effluent
1346 from this treatment plant (Weston and Lydy 2010). This was not the case with effluent from the
1347 Stockton Wastewater Treatment Plant, which utilizes tertiary treatment, suggesting that different
1348 treatment methods may remove or retain pyrethroids differently. Prior to this study, urban inputs

1349 in general and wastewater treatment plants in particular had not been considered a primary
1350 source of insecticides in the Delta. Urban insecticide applications thus represent an unexpected
1351 contaminant source. This is of concern because the estuary as a whole as well the southern and
1352 western subregions can be classified as stressed by urbanization on the basis of impervious cover
1353 due to the rapid growth of the Bay Area, Sacramento, and Stockton metropolitan areas (Stoms
1354 2010).

1355
1356 Brander et al. (2009) showed that mixtures of the two pyrethroid pesticides, permethrin and
1357 cyfluthrin, are lethal to *H. azteca* at concentrations found in Delta water samples. Beggel et al.
1358 (2010) found that commercial formulations of two commonly used insecticides, bifenthrin and
1359 fipronil, were more toxic to fathead minnows than the pure active ingredients. Significant
1360 sublethal effects on fish swimming performance or growth were observed at 10–20% of the
1361 concentrations required to kill 10% of exposed fish (LC10). These results suggest that increased
1362 toxicity due to inert ingredients should be considered in risk assessments and regulation of
1363 insecticides (Beggel et al. 2010).

1364
1365 The sublethal effects measurements conducted by Beggel et al. (2010) represent another major
1366 aspect of the POD contaminants studies, examining sublethal effects. Werner et al. (2008,
1367 2010b) focused primarily on the development and application of biochemical and molecular
1368 biomarkers of sublethal stress responses to contaminants. Geist et al. (2007) investigated the
1369 effects of copper and the pyrethroid insecticide esfenvalerate on survival, growth, swimming
1370 behavior, and expression of stress response genes in juvenile (81–90 d old) striped bass (*M.*
1371 *saxatilis*). They found that the expression of stress response genes was the most sensitive
1372 indicator for copper and esfenvalerate exposures at low concentrations. Expression of stress
1373 response genes was also seen in striped bass exposed to Delta water samples (Werner et al.
1374 2008).

1375
1376 Connon et al. (2009, in press, in review) developed a cDNA microarray for delta smelt to study
1377 the sublethal effects of chemical exposure. Connon et al. (2009) used the microarray to assess the
1378 effects of exposure of larval delta smelt to the pyrethroid esfenvalerate and identified specific
1379 patterns of gene expression related to affected physiological functions. Connon et al. (in press)
1380 then exposed larval delta smelt to sublethal concentrations of esfenvalerate. The microarray
1381 results showed patterns of gene expression indicating neurological impairment, supported by
1382 anomalous behavior, and suggesting effects on gonad and brain development. Connon et al. (in
1383 review) evaluated the effects of copper on survival, swimming velocity and gene expression in
1384 larval and juvenile delta smelt. Swimming velocity declined with increasing copper
1385 concentration. The microarray results indicated significant sublethal effects of copper on nerve
1386 and muscle function, digestion and immune function at approximately 20% of the LC50
1387 concentration. In ammonia exposure experiments, genes predominantly encoding for membrane
1388 bound proteins responded significantly; however, neurological activity and muscular activity
1389 were also impaired (Werner et al. 2010).

1390
1391 Biomarkers were also developed and applied in studies with striped bass and fathead minnows
1392 (Werner et al. 2010). Juvenile striped bass were exposed to extracts from SPMDs (semi-
1393 permeable membrane devices, which collect dissolved lipophilic compounds from water), which
1394 were deployed at several sites in the Delta, Tissue samples from the striped bass were analyzed

1395 for expression of four stress-responsive genes. Results indicated general stress and immune
1396 system responses at Boynton Slough (Suisun Marsh) and Sherman Lake (confluence) sites
1397 (Werner et al. 2010). In fathead minnows, observed gene responses suggested stress-related
1398 cellular effects in fathead minnow larvae exposed to bifenthrin and fipronil (0.07 and 53 µg/L,
1399 respectively) (Beggel et al. 2010, Werner et al. 2010). The concentrations associated with the
1400 gene responses corresponded to the concentrations causing abnormal swimming. Patterns of gene
1401 response in delta smelt exposed to water from different delta sites show great promise for
1402 toxicity monitoring in field surveys (R. Connon, UC Davis, personal communication). Results
1403 from these initial studies support the use of microarrays to assess the effects of contaminants on
1404 aquatic species in field studies. Further method development and application for delta smelt are
1405 in progress.

1406
1407 In addition to the novel methods described above, the POD investigations have included the use
1408 of biomarkers and histological techniques used previously to evaluate toxic effects on POD
1409 fishes (Bennett et al. 1995, Bennett 2005). The results to date have been mixed for the four POD
1410 fish species. Histopathological and viral evaluation of young longfin smelt collected in 2006 and
1411 2007 indicated no histological abnormalities associated with toxic exposure or disease (Foott et
1412 al. 2006, Foott and Stone 2008). There was also no evidence of viral infections or high parasite
1413 loads. Similarly, young threadfin shad showed no histological evidence of contaminant effects
1414 or of viral infections (Foott et al. 2006, Foott and Stone 2008). Parasites were noted in threadfin
1415 shad gills at a high frequency but the infections were not considered severe. Thus, both longfin
1416 smelt and threadfin shad were considered healthy in 2006 and 2007. Adult delta smelt collected
1417 from the Delta during winter 2005 also were considered healthy, showing little histopathological
1418 evidence for starvation or disease (Bennett et al. 2008). However, there was some evidence of
1419 low frequency endocrine disruption. In 2005, 3 of 47 (6%) of adult delta smelt were intersex,
1420 having immature oocytes in their testes (Bennett et al. 2008). This is a lower percentage than
1421 reported by Teh et al. (2007) (9 of 65 male delta smelt, 14%). Juvenile delta smelt exhibited a
1422 high incidence of glycogen depletion (80% of fish) and liver abnormalities (85% of fish),
1423 presumably in response to stressful summer rearing conditions (Bennett et al. 2008).

1424
1425 In contrast, histopathological analyses have found evidence of significant disease in other species
1426 and for POD species collected from other areas of the estuary. Massive intestinal infections with
1427 an unidentified myxosporean were found in yellowfin goby *Acanthogobius flavimanus* collected
1428 from Suisun Marsh (D. Baxa, UCD, unpublished data). Severe viral infection was found in
1429 Mississippi silverside and juvenile delta smelt collected from Suisun Bay during summer 2005
1430 (Baxa et al. in prep.).

1431
1432 Contaminants and disease may severely impair striped bass of all life stages in the estuary
1433 (Ostrach et al. 2008a). Contaminants were identified as having significant effects on larval
1434 striped bass and juvenile striped bass up to 6–8 months old. Juvenile striped bass were found to
1435 be suffering from sublethal contaminant exposure and abnormal disease and parasitism. Adult
1436 striped bass are likely adversely affected by the bioaccumulation of contaminants such as
1437 polybrominated diphenyl ethers (PBDEs). In addition to these contaminant and disease effects,
1438 striped bass may be especially vulnerable to contaminant effects because the long-lived females
1439 can bioaccumulate contaminants over several years and then transfer the contaminants to egg
1440 yolk. This results in maternal transfer of xenobiotics to eggs resulting in severe adverse effects

1441 on larval development and subsequent larval survival that may be contributing to population
1442 level effects (Ostrach et al. 2008b). Contaminants are thus likely a significant driver contributing
1443 to the decline of young striped bass in the San Francisco Estuary (Ostrach et al. 2008a).

1444
1445 Finally, the POD investigation has also included studies of the potentially toxic cyanobacterium
1446 *M. aeruginosa*. Large blooms of *M. aeruginosa* were first noted in the Delta in 1999 and have
1447 since become an annually recurring feature of the Delta plankton community (Lehman et al.
1448 2005, 2008). During the low flow conditions of summer 2007, blooms of this cyanobacterium
1449 spread downstream to the western Delta and beyond (P. Lehman, DWR, unpublished data).
1450 *Microcystis* is more abundant in dry years (2007 and 2008) compared to wet years (2004 and
1451 2005) (P. Lehman, DWR, unpublished data). Genetic analysis determined that the *M.*
1452 *aeruginosa* strain in the Delta is unique (Moisander et al. 2009). The *M. aeruginosa* blooms peak
1453 in the fresh waters of the central Delta during the summer at warm temperatures (20–25°C;
1454 Lehman et al. 2008). Large striped bass and all life stages of threadfin shad occur widely in the
1455 central and southern Delta during summer, and thus may be at the highest risk of exposure to
1456 toxic blooms. Longfin smelt and delta smelt are generally not present in this region of the Delta
1457 during summer (Nobriga et al. 2008, Rosenfield and Baxter 2007) so it is less likely that *M.*
1458 *aeruginosa* toxicity is a factor in their recent decline. However, a recent study by Miller et al.
1459 (2010) documented a complex transfer of toxic microcystins from freshwater, through mollusks
1460 (which bioaccumulated the microcystins), to sea otters, resulting in mortality of sea otters in
1461 coastal Monterey Bay. This study suggests there is some potential that fish residing in brackish
1462 and marine parts of the estuary could be affected by toxic *Microcystis* blooms through trophic
1463 pathways.

1464
1465 Baxa et al. (2010) developed a genetic technique for quickly identifying *Microcystis* abundance
1466 and toxicity and determined that *Microcystis* toxicity is spatially and temporally variable in the
1467 Delta. Elevated toxicity was common in the brackish western Delta near Antioch. This was
1468 unexpected and may indicate the presence of a salinity tolerant strain in this region. This is the
1469 same area where elevated microcystin levels and biomarker scores for juvenile striped bass were
1470 found. A recent laboratory study (Deng et al. 2010) showed that dietary *Microcystis* is toxic to
1471 Medaka fish (*Oryzias latipes*) and has a more adverse impact on male fish. Results suggest that
1472 long-term exposure to microcystins in the Delta may be a health problem for fish.

1473
1474 *Microcystins* probably do not reach concentrations acutely toxic to Delta fishes (Lehman et al.
1475 2008), but during blooms, the microcystin concentrations may be high enough to impair
1476 invertebrates, which could influence prey availability for fishes. Lehman et al. (2010) found that
1477 *Microcystis* may indeed contribute to changes in phytoplankton, zooplankton and fish
1478 populations in the Delta. Ger et al. (2009) determined toxicity of one form of microcystin (LR)
1479 to two species of calanoid copepods, *E. affinis* and *P. forbesi*, which are important as food to
1480 POD species. They found that, although the copepods tested were relatively sensitive to
1481 microcystin-LR compared to other types of zooplankton, ambient concentrations in the Delta
1482 were unlikely to be acutely toxic. However, chronic effects were not determined. Feeding trials
1483 indicated that both species of copepods were very sensitive to *Microcystis* (Ger et al. 2010a).
1484 Both species experienced significant mortality when *Microcystis* was 10% or more of the diet,
1485 whether the *Microcystis* strain produced toxic microcystins or not. This suggests that
1486 consumption of *Microcystis* by wild populations of copepods has the potential to affect food

1487 resources of POD fishes. *P. forbesi* ingested an order of magnitude less than *E. affinis*,
1488 suggesting it is better able to tolerate the presence of *Microcystis* in its diet than *E. affinis* via
1489 more efficient selective feeding on alternative food (Ger et al. 2010b). It also appears that
1490 selective feeding by *P. forbesi* becomes more pronounced during long-term exposure to different
1491 *Microcystis* strains during blooms.

1492
1493 In summary, many different man-made and natural toxins are present in the estuary. The absence
1494 of consistent long-term monitoring of contaminant concentrations does not allow for a
1495 correlative assessment of contaminant effects on Delta biota over time (Johnson 2010) and
1496 prevents contaminants from being included in statistical analyses of the POD declines such as
1497 those conducted by MacNally et al. (2010) and Thomson et al. (2010). While four years of
1498 biweekly water toxicity tests have provided evidence for sporadic direct, acute toxicity, the
1499 majority of contaminant effects are likely sublethal and/or mediated by the food web (Scholz et
1500 al. in review, Brooks et al. in review). Unfortunately, conventional methods of testing pesticides
1501 on a chemical by chemical basis or conducting bioassays with standard test organisms are
1502 unlikely to be useful in assessing the effects of complex mixtures of many chemicals on the
1503 ecosystem, especially when many of the effects are nonlethal and chronic (Scholz et al. in
1504 review). From a scientific perspective, novel biomarkers such as those developed by Connon et
1505 al. (2009) hold great promise for arriving at a more reliable and nuanced understanding of
1506 contaminant effects. From a management perspective, these problems might best be addressed
1507 by reducing transport of all pesticides to the aquatic environment rather than managing
1508 chemicals suspected to be important on a case by case basis (Scholz et al. in review).

1509
1510 *Habitat for Planktonic and Benthic Aquatic Organisms:* Much of the previous discussion about
1511 how physical conditions and water quality affect pelagic fishes is also relevant to other aquatic
1512 organisms including plankton and the benthos. The abundance and distribution of pelagic
1513 organisms in an estuary, particularly plankton, results from the interaction of production,
1514 consumption of the organism, and transport (Lucas et al. 2009b). Factors limiting phytoplankton
1515 growth in the estuary have already been discussed above. The distribution and abundance of
1516 benthos represents an interaction between conditions during the period when species are
1517 recruiting to the bottom and subsequent conditions for survival, growth, and reproduction.

1518
1519 It is important to keep in mind that river flows and exports influence estuarine salinity gradients
1520 and water transport times (i.e., how fast water or transported constituents move through the
1521 system). These factors affect both habitat suitability for benthos and the transport of pelagic
1522 plankton. High delta outflow leads to more rapid transport through the Delta (days), which
1523 generally results in lower plankton biomass (Kimmerer 2004), but also lower cumulative
1524 entrainment effects in the Delta (Kimmerer and Nobriga 2007). In contrast, longer transport
1525 times (a month or more), which result from low delta outflows, may result in higher plankton
1526 biomass, depending on consumption rates by grazers (see Lucas et al. 2009a). This can increase
1527 food availability for planktivorous fishes; however, much of this production may be lost to filter
1528 feeding benthos (e.g., Lucas et al. 2002) or water diversions (Arthur et al. 1996) under low delta
1529 outflow conditions. In the San Joaquin River under extreme low flow conditions, long water
1530 residence times may also promote high biological oxygen demand when abundant phytoplankton
1531 die and decompose (Lehman et al. 2004, Jassby and Van Nieuwenhuyse 2005). Recent particle
1532 tracking modeling results for the Delta show that transport times in the central Delta are highly

1533 variable depending on Delta inflow, exports, and particle release location (Kimmerer and
1534 Nobriga 2007). Very high inflow leads to rapid transport times. The longest transport times
1535 occur in the San Joaquin River near Stockton under conditions of low inflow and low export
1536 flow.

1537
1538 Salinity variation can have a major effect on the benthos, which occupy relatively “fixed”
1539 geographical positions along the gradient of the estuary. While the distributions of the benthos
1540 can undergo seasonal and annual shifts, benthic organisms cannot adjust their locations as
1541 quickly as more mobile pelagic organisms. Analyses of long-term benthic data for four regions
1542 of the upper San Francisco Estuary indicate that two major factors control community
1543 composition: species invasions and salinity (Peterson and Vayssières 2010). Specifically, the
1544 invasion of the clam *C. amurensis* in the late 1980s resulted in a fundamental change in the
1545 benthic community; however, the center of distribution of *C. amurensis* and other benthic
1546 species shifts geographically with the estuarine salinity gradient, as the salinity gradient responds
1547 to outflows. So at any particular location in the estuary, the benthic community can change
1548 substantially from year to year as a result of environmental variation and species invasions.
1549 There was nothing unusual about the composition of benthic assemblages during the POD
1550 period. As will be discussed below, changes in the benthos can have major effects on food
1551 availability to pelagic organisms.

1552
1553 *Climate Change Effects on Habitat:* While climate change did not cause the POD, there are
1554 several reasons we expect climate change will have negative influences on future pelagic habitat
1555 suitability for the POD fishes. First, there has been a trend toward more Sierra Nevada
1556 precipitation falling as rain earlier in the year (Roos 1987, 1991, Knowles and Cayan 2002,
1557 2004). This increases the likelihood of winter floods and may have other effects on the
1558 hydrographs of Central Valley rivers and Delta salinity. Altered hydrographs interfere with
1559 pelagic fish reproduction, which is usually tied to historical runoff patterns (Moyle 2002).

1560
1561 Second, sea level is rising (IPPC 2001). Sea level rise will increase salinity intrusion, moving
1562 X2 landward, unless sufficient freshwater resources are available to repel the seawater. This will
1563 shift fish distributions upstream and possibly further reduce habitat suitability for some species.
1564 Based on the results of Feyrer et al. (2007), Feyrer et al. (2010) developed a model to predict
1565 delta smelt habitat quality in response to X2. Data from several scenarios of climate change
1566 predict declining habitat suitability as climate change proceeds (Feyrer et al. 2010). Sea level
1567 rise will also increase the likelihood of levee failures (Mount and Twiss 2005) and perhaps
1568 reduce the likelihood of repairing such failures, given economic considerations (e.g., Suddeth et
1569 al. 2010). Levee failures and flooding of islands would lead to changes in available aquatic
1570 habitat (Lund et al. 2007, 2008). In addition, currently unleveed wetlands and other habitats will
1571 be susceptible to inundation (Knowles 2010).

1572
1573 Third, climate change models project warmer air temperatures in central California (Dettinger
1574 2005). As stated above, water temperatures do not currently have a strong influence on POD fish
1575 distributions. However, summer water temperatures throughout the upper estuary are fairly high
1576 for delta smelt. Mean July water temperatures in the upper estuary are typically 21–24°C
1577 (Nobriga et al. 2008) and high mortality of delta smelt is expected above about 25°C based on
1578 several lines of evidence. First, the critical thermal maximum of delta smelt acclimated to 17°C

1579 is 25°C (Swanson et al. 2000). Second, delta smelt are rarely captured in field surveys when
1580 water temperatures reach 25°C (Bennett 2005, Nobriga et al. 2008). Third, following
1581 acclimation by gradually increasing temperature to match ambient temperatures and with daily
1582 feeding, juvenile delta smelt can survive short term exposure to ambient temperatures up to
1583 about 27°C for several days (G. Castillo, USFWS, unpublished data). However, continued
1584 exposure to daily peaks in ambient temperatures over 27.0 °C lead to sudden high mortality,
1585 even if subsequent water temperature decline (G. Castillo, USFWS, unpublished data). Models
1586 developed for water temperature in the delta (Wagner et al. submitted) suggest that warmer air
1587 temperatures predicted by climate change will result in summer temperatures in the upper
1588 estuary exceeding 25°C on more days. This will likely affect viability of the delta smelt
1589 population. Temperatures between 20 and 25°C may lead to sublethal stress of delta smelt and
1590 other native species that could reduce growth rates or otherwise affect fish populations (Bennett
1591 et al. 2008). On the other hand, some non-native and nuisance species adapted to warmer
1592 temperatures such as largemouth bass and *M. aeruginosa* will likely increase in abundance with
1593 rising water temperatures (Feyrer and Healey 2003, Paerl and Huisman 2008).

1594
1595 There may be some opportunities to ameliorate or mitigate the effects of increasing water
1596 temperatures. Thermal refugia may persist in deeper channels near the bottom. The role of
1597 shade from riparian vegetation may also be important. Many levees do or can support riparian
1598 trees; however, such vegetation may be removed by the US Army Corp of Engineers to maintain
1599 the structural integrity of the levees. Greenberg (UCD, unpublished data) modeled the relative
1600 difference in incident solar radiation on channels in the Delta under the current conditions and in
1601 a treeless Delta. The model used Lidar data acquired in 2007 to assess the structural conditions
1602 and a solar irradiation model to calculate daily irradiation for summer months. The results
1603 indicated that as much as a 10% increase in solar radiation could occur in a treeless Delta, which
1604 may result in significant increases in water temperature – clearly an undesirable side effect of
1605 removing trees. Increasing shade over current levels might help mitigate the temperature
1606 increases calculated by Wagner et al. (submitted). Tidally driven daily water exchanges between
1607 shallow and deeper water areas can also affect overall water temperatures (C. Enright, DWR,
1608 personal communication), and tidal marsh restoration may provide an opportunity for
1609 counteracting expected increases in water temperature if tidal marshes can cool water
1610 sufficiently.

1611

1612 *Top-Down*

1613

1614 Predation is a common mechanism by which weakened fish are ultimately killed. Thus,
1615 increased predation can be a manifestation of other changes in the ecosystem like decreased
1616 habitat suitability, starvation, and disease. However, in the top-down section of our conceptual
1617 model, we are referring to elevated mortality of healthy individuals due to predation or removal
1618 by water diversions and associated factors. Thus, the top-down effects are predicated on two
1619 hypotheses, which are not mutually exclusive. The first is that consumption or removal of
1620 healthy fish biomass by piscivores (principally striped bass and largemouth bass *Micropterus*
1621 *salmoides*) increased around 2000. The second is that mortality due to water diversions
1622 (SWP/CVP exports; power plant diversions) increased around 2000. This could have occurred if
1623 one or more of the following happened: (1) water diversions and exports increased during

1624 periods the POD fishes were vulnerable to them; (2) piscivorous fishes became more abundant
1625 relative to the POD fishes; (3) pelagic fish distribution shifted to locations with higher predation
1626 risk (e.g. habitat changes); or (4) the POD fishes became more vulnerable to predation as a
1627 consequence of their extremely low population size (i.e., predation could contribute to the Allee
1628 effect hypothesized in the Previous Abundance section) or increases in water clarity.

1629
1630 Predation-driven Allee effects can arise from diminished anti-predator behavior or increased
1631 predator swamping of individuals in smaller prey groups (Berec et al. 2006). They are most
1632 likely to occur with generalist predators in situations where predation is a major source of
1633 mortality, and predation refuges are limited (Gascoigne et al. 2004). In this situation, individuals
1634 of depleted populations continue to be consumed even though they are at low density.
1635 Specialized predators that tend to focus on only a few species at any one time often switch
1636 between abundant prey species and consequently reduce consumption of rare prey species. As
1637 will be described below, the combination of a widely distributed pelagic piscivore (striped bass),
1638 an efficient littoral piscivore (largemouth bass), cumulative entrainment losses of multiple life
1639 stages, and decreased habitat suitability suggest the conditions listed by Gascoigne et al. (2004)
1640 could apply in the Delta.

1641
1642 *Predation Effects:* Predation is a natural process that occurs in almost all ecosystems. However,
1643 problems can occur when established predator-prey relationships are disrupted by environmental
1644 changes or species introductions. Many examples of ecosystem change due to introduced
1645 predators (e.g., Brown and Moyle 1991, Goldschmidt et al. 1993) or changes in established
1646 predator-prey relationships (e.g., Carpenter et al. 2001, Frank et al. 2005) have been documented
1647 in aquatic systems. This hypothesis suggests that predation effects have increased in all water
1648 year types as a result of increased populations of pelagic and inshore piscivores.

1649
1650 Predation processes were almost certainly changed with the introduction of striped bass to the
1651 Delta in 1879. As mentioned earlier, the striped bass was very successful and supported a
1652 commercial fishery within a decade after the introduction. Moyle (2002) observed that the fast
1653 growing, pelagic, schooling striped bass was likely a much more effective predator on pelagic
1654 prey than the native predators including the relatively sluggish thicketail chub (*Gila craussicauda*,
1655 now extinct) and Sacramento perch (*Archoplites interruptus*, extirpated from Delta; Crain and
1656 Moyle, in press), slow growing Sacramento pikeminnow *Ptychocheilus grandis*, and cold water
1657 requiring steelhead rainbow trout *Oncorhynchus mykiss*. In fact, at least the juvenile stages of all
1658 these species are or were likely consumed by large striped bass. Unfortunately, historical data
1659 are not available to determine if changes in predation rates have occurred.

1660
1661 Continued predation pressure from striped bass on Delta pelagic fishes in recent years is likely
1662 because of the number of predatory juvenile (Figure 15) and adult striped bass (Figure 9) as well
1663 as their spatial and seasonal distribution, food habits (Stevens 1966), and bioenergetics
1664 (Loboschefskey et al. submitted). Although age-0 striped bass themselves are a POD species, the
1665 resiliency of the species is comparatively high and demonstrated in part by the fact that juvenile
1666 age-1 and age-2 striped bass have declined more slowly than age-0 striped bass (compare Figure
1667 9 with Figure 15, CDFG, unpublished data). However, a paucity of properly designed striped
1668 bass food habit studies has precluded the direct estimation of the number of delta smelt, longfin
1669 smelt, threadfin shad, and juvenile striped bass consumed by striped bass during the POD years.

1670
1671 Largemouth bass abundance has increased in the Delta over the past few decades (Brown and
1672 Michniuk 2007). Largemouth bass were introduced to the Central Valley in the mid-1890s (Dill
1673 and Cordone 1997) and were present in the Delta soon after that. Although none of the IEP
1674 surveys adequately tracks largemouth bass population trends, a comparison of abundance
1675 estimates between intermittent surveys conducted in the early 1980s, late 1990s, early 2000s
1676 (CDFG, unpublished data) and from 2009 to 2010 (L. Conrad, DWR, unpublished data) shows
1677 that largemouth bass and sunfish populations more than doubled during the years of the POD
1678 (Figure 16).

1679
1680 Analyses of fish salvage data show an abrupt increase in salvage of young largemouth bass in the
1681 early 1990s, before the POD, with salvage remaining at high levels since then (Figure 17). This
1682 suggests an increase in largemouth bass abundance in the early 1990s. The increase in salvage
1683 of largemouth bass occurred during the time period when *E. densa*, an introduced aquatic
1684 macrophyte was expanding its range in the Delta (Brown and Michniuk 2007). Although the
1685 historic distributions of native species of SAV are not known, it is possible that their coverage
1686 may not have been as extensive or persistent as *E. densa* is today. For example, unlike most
1687 native aquatic macrophytes, *E. densa* has a bimodal growth pattern, with peaks in late spring and
1688 the early fall. The second growth period in late fall may help existing patches persist through the
1689 winter and provide a head start on growth the following spring. These characteristics likely
1690 contributed to the expansion of the distribution of *E. densa* in the Delta and perhaps help *E.*
1691 *densa* compete with other aquatic macrophytes (Santos et al. 2010). The invasion of *E. densa*
1692 has occurred during highly altered environmental conditions compared to historical conditions.
1693 Historical conditions, including dynamic flow and salinity regimes, higher turbidity, and
1694 seasonal (rather than perennial) inundation of large portions of shallow-water habitat would
1695 likely have been less favorable for establishment of *E. densa*. The areal coverage of *E. densa* in
1696 the Delta has fluctuated from 2004 to 2008, suggesting that this habitat may no longer be
1697 expanding (Hestir 2010).

1698
1699 Largemouth bass have a much more limited distribution in the estuary than striped bass, but a
1700 higher per capita impact on small fishes (Nobriga and Feyrer 2007). Conceivably, increases in
1701 largemouth bass may have had a particularly important effect on threadfin shad and striped bass,
1702 whose earlier life stages occur in littoral habitat (Grimaldo et al. 2004, Nobriga and Feyrer
1703 2007). However, ongoing analyses of largemouth bass diet suggest that the largemouth bass are
1704 chiefly consuming common littoral invertebrate and fish species, such as crayfish and juvenile
1705 sunfish (L. Conrad, DWR, unpublished data). To date, over 1400 samples collected from sites
1706 located throughout the Delta have been examined and these have contained only 12 threadfin
1707 shad and 1 juvenile striped bass (L. Conrad, DWR, unpublished data). Furthermore, no salmonid
1708 or osmerid species have been found in largemouth bass stomachs. The zero or low frequencies of
1709 the POD species in bass stomachs may be due largely to limited spatial overlap with largemouth
1710 bass; however, increased abundance of largemouth bass may still impose an important predation
1711 threat in limited instances where they do co-occur.

1712
1713 Predation pressures on the early life stages of POD species are poorly understood but likely have
1714 highly significant impacts on recruitment. Historically, questions involving predation of larval
1715 fish have been difficult to address due to the quick degradation of fragile larvae in predator guts.

1716 However, such studies have become feasible with the development of a genetic assay to look for
1717 the DNA of delta smelt in the guts of predators (M. Baerwald et al., DWR, unpublished data).
1718 Initial application of this assay has focused mainly on Mississippi silversides; largely due to their
1719 increasing abundance in the Delta and their ability to eat delta smelt larvae in captivity (Bennett,
1720 2005). To date, sampling conducted in portions of the Sacramento deepwater ship channel
1721 where larval smelt are relatively abundant, found that of 37 captured silversides, 15 tested
1722 positive for delta smelt DNA in their gut contents (B. Schreier, DWR, unpublished data).
1723

1724 A change in predation pressure may, in part, be an effect of interactions between biotic and
1725 abiotic conditions. Natural, co-evolved piscivore-prey systems typically have an abiotic
1726 production phase and a biotic reduction phase each year (e.g., Rodriguez and Lewis 1994).
1727 Changing the magnitudes and durations of these cycles greatly alters their outcomes (Meffe
1728 1984). Generally, the relative stability of the physical environment affects the length of time
1729 each phase dominates and thus, the importance of each. Biotic interactions like predation will
1730 have a stronger influence on the biotic community in physically stable systems (e.g., lakes).
1731 Historically in the estuary, the period of winter-spring high flow was the abiotic production
1732 phase, when most species reproduced. The biotic reduction phase probably encompassed the
1733 low-flow periods in summer and fall. Multi-year wet cycles probably increased (and still do) the
1734 overall “abiotic-ness” of the estuary, allowing populations of all species to increase. Drought
1735 cycles likely increased the estuary’s “biotic-ness” (Livingston et al. 1997), with low reproductive
1736 output and increased effect of predation on population abundances. Flow management in the
1737 San Francisco Estuary and its watershed has reduced flow variation much of the time and in
1738 some locations more than others (Moyle et al. 2010). This has probably affected the magnitudes
1739 and durations of abiotic and biotic phases (Nobriga et al. 2005). In other words, reduced
1740 variability in environmental conditions of the estuary may have exacerbated predation effects.
1741 However, there is no clear evidence that such changes have been abrupt enough to account for
1742 the POD.
1743

1744 *Entrainment:* The water diversions that are of most concern for fishes in the estuary are the
1745 SWP and CVP export facilities, Antioch and Pittsburg power plants, and within-Delta
1746 agricultural diversions. Of these, the operations of agricultural diversions are the least likely to
1747 have had an effect on the POD species because of the small volumes they divert and because it
1748 does not appear they have changed operations from the 1990s to 2000s. A detailed study of one
1749 of these diversions found evidence that their effects on delta smelt are small (Nobriga et al.
1750 2004). Because agricultural diversions seem an unlikely contributor to the POD declines, we do
1751 not address them further. We do not mean to assert that such diversions have no effect on Delta
1752 fishes. Addressing that issue would require additional analysis and study.
1753

1754 The Antioch and Pittsburgh power plants divert approximately 3200 cubic feet per second for
1755 non-consumptive water use when they are being operated. The power plants can directly entrain
1756 pelagic fishes or affect them indirectly through discharge of heated water (Matica and Sommer,
1757 DWR, unpublished data). Studies at the power plants in the late 1970s indicated that losses of
1758 delta smelt and longfin smelt were on the order of hundreds of thousands of individuals;
1759 however, the plants were operated less frequently during the 2000s, suggesting that the power
1760 plants played a minor, if any, role in the POD (Cavallo et al. 2009).
1761

1762 Entrainment losses at the SWP and CVP represent the largest sources of observable mortality for
1763 many pelagic fishes in the estuary (Brown et al. 1996, Sommer et al. 2007, Kimmerer 2008,
1764 Grimaldo et al. 2009). SWP and CVP entrainment losses are indexed by fish captured (salvaged)
1765 at the state Skinner Fish Protective Facility (SFF) and the federal Tracy Fish Collection Facility
1766 (TFCF). These facilities are located at the intakes to the State and Federal export pumps on Old
1767 River in the south-western Delta. While the TFCF is located directly on Old River, the SFF is
1768 preceded by the large Clifton Court Forebay (CCF), a reservoir that is connected to Old River via
1769 operable radial gates. Subsamples of the fish captured at the fish facilities are identified and
1770 counted and all fish are trucked to several locations in the Delta where they are released back
1771 into the wild.

1772
1773 It is important to note that fish salvage represents only the fraction of the total number of
1774 entrained fish. As many studies have shown, many more fish are entrained, but not salvaged due
1775 to pre-screen losses and capture inefficiencies at the fish facilities (Brown et al. 1996, Gingras
1776 1997, Kimmerer 2008, Clark et al. 2009). While capture inefficiencies occur at both facilities,
1777 pre-screen losses at the SWP are exacerbated by the CCF. CCF is a shallow reservoir with high
1778 numbers of predatory fish, piscivorous birds, and in recent years large beds of SAV requiring
1779 removal (Kano 1990, Brown et al. 1996, Clark et al. 2009). Predation rates likely increase with
1780 increasing residence time in CCF. Recent hydrodynamic and 3-D particle tracking modeling
1781 studies in CCF showed that wind and exports affect residence time in CCF (M. MacWilliams,
1782 River Modeling Environmental Consulting, unpublished data). During periods of high winds and
1783 low exports, a strong counterclockwise circulation gyre in CCF results in significant mixing and
1784 increases the average travel time from the radial gates to the SFF, with a large range of estimated
1785 particle residence times. During low wind and high export conditions, residence times are much
1786 shorter and most particles are transported roughly in a straight line trajectory from the radial
1787 gates to the Banks Pumping Plant.

1788
1789 Results of 11 mark-recapture and telemetry studies conducted in CCF with juvenile hatchery-
1790 raised striped bass and salmonids between 1976 and 2007 showed consistently high pre-screen
1791 losses ranging from 63% to 99% (Gingras 1997, Clark et al. 2009), likely due to high levels of
1792 predation in CCF. In a recent study, thousands of marked juvenile and adult hatchery-raised delta
1793 smelt were released into CCF and recaptured in the SFF (G. Castillo, USFWS, unpublished
1794 data). The mean percent recovery of adult fish released in CCF varied from 3.01% and 0.41% in
1795 February and March, respectively, to 0.03% for juveniles in June. This means that most delta
1796 smelt entrained into CCF may never be captured and accounted for at the SFF, with entrainment
1797 varying 10 to 100-fold between February and June. As this and previous studies have shown,
1798 pre-screen losses are both variable and consistently high. *Salvage is thus not a sensitive index of*
1799 *entrainment*. There has not been any monitoring of actual entrainment and no correction factors
1800 for prescreen losses and capture inefficiencies have been developed or applied to estimate
1801 entrainment. Therefore, salvage numbers have been and continue to be used as a substitute for
1802 entrainment in spite of their well-documented limitations.

1803
1804 During the POD years, summer, fall, and winter exports sharply increased from the 1990s, which
1805 led to increased salvage and thus likely increased entrainment of several pelagic fishes in the
1806 estuary (Sommer et al. 2007, Kimmerer 2008, Grimaldo et al. 2009). Winter exports increased
1807 the most during the POD years, which led to increases in salvage of adult delta smelt, adult

1808 longfin smelt, and threadfin shad (Figure 18). Similar increases in the salvage of littoral species,
1809 including centrarchids and inland silverside, were observed during the same period (Figure 19).
1810 The littoral species are less influenced by flow changes than the POD fishes. The increases in
1811 salvage for centrarchids, including largemouth bass, may be at least partially a result of the range
1812 expansion of *E. densa*, which provides favored habitat. This hypothesis is supported by the
1813 observation that the greatest increases in centrarchid salvage occurred at the CVP. The intake of
1814 the CVP is located in an area with significant areas of *E. densa* nearby. Nonetheless, the
1815 increase in entrainment of both pelagic and littoral fishes suggests a large change in the
1816 hydrodynamic influence of the export diversions during recent winters. Note that winter salvage
1817 of all the POD species declined during the winters of 2005–2006, 2006–2007 and 2008–2009,
1818 possibly due to the drop in their population abundance.

1819
1820 In trying to evaluate the mechanism(s) for increased wintertime salvage, early POD studies
1821 produced three key observations (IEP 2005). First, there was an increase in exports during
1822 winter as compared to previous years. Second, the proportion of tributary inflows shifted.
1823 Specifically, San Joaquin River inflow decreased as a fraction of total inflow around 2000, while
1824 Sacramento River increased. Finally, there was an increase in the duration of the operation of
1825 barriers placed into southern Delta channels during some months. These changes may have
1826 contributed to a shift in Delta hydrodynamics that increased fish entrainment. These
1827 observations led to a hypothesis that the hydrodynamic change could be indexed using net flows
1828 through Old and Middle rivers (OMR) (Figure 1B), which integrate changes in inflow, exports,
1829 and barrier operations (Arthur et al. 1996, Monsen et al. 2007). Net flow refers to the magnitude
1830 and direction (seaward or landward) of the water in OMR with the effects of the semidiurnal tide
1831 removed. Grimaldo et al. (2009) found significant relationships between OMR flows and fish
1832 salvage for delta smelt, longfin smelt, and striped bass (Figure 20). Not only has net OMR flow
1833 been useful to understand entrainment of fishes in the estuary (Kimmerer 2008, Grimaldo et al.
1834 2009), it was adopted as a regulatory tool to manage entrainment risk of delta smelt from the
1835 SWP and CVP exports (USFWS 2008).

1836
1837 Although entrainment into water diversions is well recognized as a source of mortality for
1838 individual fish, there has been debate about whether entrainment has important population-level
1839 effects for the pelagic fishes in the estuary. Kimmerer (2008) estimated that entrainment of delta
1840 smelt at the SWP and CVP accounted for high population losses (up to 50%) during POD years,
1841 suggesting that exports played a major role in the POD decline. However, the population-level
1842 effects of entrainment can also be obscured by interactions between year classes or life stages of
1843 a species. For example, Kimmerer (2008) found that exports explained little variability in fall
1844 abundance of delta smelt because of a 50-fold variation in their summer to fall survival.
1845 Kimmerer concluded, “This is not to dismiss the rather large proportional losses of delta smelt
1846 that occur in some years; rather, it suggests that these losses have effects that are episodic and
1847 that therefore their effects should be calculated rather than inferred from correlative analyses.”
1848 Similarly, Manly and Chotkowi (2006) used log-linear modeling to evaluate environmental
1849 factors that may have affected long-term trends in the FMWT abundance index of delta smelt.
1850 They found that monthly or semi-monthly measures of exports or OMR flow had a statistically
1851 significant effect on delta smelt abundance; however, individually exports and flow explained
1852 only a small portion (no more than a few percent) of the variability in the fall abundance index of

1853 delta smelt across the entire survey area and time period. Hence, there are other factors that
1854 dominate the long-term trends of delta smelt fall abundance.

1855
1856 Thomson et al. (2010) found that turbidity (discussed as water clarity in that paper) and winter
1857 exports were important factors associated with the long-term trend in annual fall abundance of
1858 delta smelt (as measured by FMWT) but could not explain the step decline in fish abundance in
1859 the early 2000s. For longfin smelt, they found that a long-term population trend was associated
1860 with turbidity and spring X2. As for delta smelt, these factors could not account for the POD
1861 decline. A long-term trend in age-0 striped bass was associated with turbidity and an
1862 autocorrelation with abundance in previous years. Again, the POD decline was unexplained by
1863 the variables that accounted for the long-term trend. There was weak evidence for winter and
1864 spring exports and calanoid copepod abundance being important for long-term trends in threadfin
1865 shad abundance but again the POD decline could not be explained. So, in all cases, exports (a
1866 surrogate for entrainment) were not useful in explaining the POD declines (Thomson et al.
1867 2010). In a multiple autoregressive analysis of the same data sets, Mac Nally et al. (2010) found
1868 some evidence for export effects on delta smelt and threadfin shad, however, other factors had
1869 stronger effects on POD species including X2 and water clarity. These results suggest that
1870 exports (i.e., entrainment) did not play a major role in the post-2000 POD, although they may
1871 have played an important role in setting up the POD through their longer-term effects and
1872 through interactions with other drivers.

1873
1874 However, these results do not mean that direct export effects can be dismissed as contributing
1875 causes of the POD. There are two aspects of entrainment that were not addressed by the earlier
1876 analyses (Manly and Chotkowski 2006, Thomson et al. 2010, Mac Nally et al. 2010) and are not
1877 well understood: (1) larval entrainment, and (2) the cumulative effects of entrainment of multiple
1878 life stages. Larval entrainment is unknown because larvae are not sampled effectively at the fish
1879 screening facilities. To address this shortcoming, Kimmerer and Nobriga (2008) coupled a
1880 particle tracking modeling with survey results to estimate larval entrainment. Kimmerer (2008)
1881 used data from several IEP monitoring programs to estimate entrainment of delta smelt. These
1882 approaches suggest that larval delta smelt entrainment losses could exceed 50% of the population
1883 under low inflow and high export conditions. Because there are few reliable larval entrainment
1884 data, it is not possible to directly address the question of how important these losses were
1885 historically. A recent attempt to manage diversions to protect fish, the Environmental Water
1886 Account, proved effective at increasing water supply reliability, but benefits to fish were not
1887 clear (Brown et al. 2009).

1888
1889 Moreover, export effects may be subtle and operate only at specific times or in specific years to
1890 disproportionately affect only one life stage of delta smelt. For example, it has been proposed
1891 that losses of larger females and their larvae may have a disproportionate effect on the delta
1892 smelt population (B. Bennett, UCD, unpublished data). Bennett (unpublished data) proposes that
1893 larger females spawn earlier in the season and produce more eggs, which are of better quality
1894 and have higher probabilities of survival, as has been noted for Atlantic cod and other
1895 commercially harvested species (Marteinsdottir and Steinarsson 1998, Swain et al. 2007). As a
1896 consequence, winter exports, which have increased since exports began (Figure 21), could have
1897 an important effect on reproductive success of early spawning female delta smelt. Bennett
1898 hypothesizes that the observed reduction in the mean size of adult delta smelt in the early 1990s

1899 (Sweetnam 1999) is a result of selective losses of earlier spawning adults and their larvae,
1900 thereby selecting for later spawned offspring (that have less time to reach maturity). Under this
1901 hypothesis, the most important result of the loss of early spawning females would manifest itself
1902 in the year following the loss, and would therefore not necessarily be detected by analyses
1903 relating fall abundance indices to same-year predictors.
1904

1905 Some recent studies suggest possible strategies for minimizing the effects of entrainment. There
1906 may be alternative ways to operate CCF to reduce prescreen losses (G. Castillo, USFWS,
1907 unpublished data). Grimaldo et al. (2009) suggested that salvage of pelagic fishes could be
1908 reduced if exports were reduced during periods when specific species were vulnerable to exports.
1909 For example, Grimaldo et al. (2009) found that adult delta smelt entrainment increased in the
1910 period following the first winter rains (first flush) when turbidity increased. This is the time
1911 period when adult delta smelt migrate upstream towards the interior delta to prepare for
1912 spawning. Grimaldo et al. (2009) suggested that entrainment risk for adult delta smelt during this
1913 time period could be reduced if exports were reduced during and after the first flush. On an
1914 intra-annual scale, adult delta smelt entrainment was related to OMR flow but there was also an
1915 interaction of OMR flow with X2. This suggests that the position of the population in relation to
1916 the salinity field prior to migration is important, but only if OMR flows are negative following
1917 first flush events. For age-0 fish, the only model explaining inter-annual differences in delta
1918 smelt salvage included zooplankton abundance suggesting a food effect. In contrast, age-0
1919 striped bass salvage was best predicted by year class strength, indicating that salvage is based
1920 simply on the number of fish in the system. Within years (intra-annual scale), age-0 delta smelt
1921 salvage was best explained by OMR flows, turbidity, and abundance. High turbidity and
1922 abundance resulted in increased salvage. Relationships between OMR flows and salvage were
1923 not found for any non-POD fishes (Grimaldo et al. 2009).
1924

1925 *Bottom-Up*

1926
1927 In this portion of the conceptual model, we propose that changes in the quality and availability of
1928 food have had important consequences for pelagic fishes. Here, we describe the evidence that
1929 there have been long-term and recent changes in food web structure and function. In the first
1930 section we discuss the availability of overall phytoplankton biomass and changes in the
1931 composition of the phytoplankton community. In the second section we discuss the implications
1932 of phytoplankton availability and species composition for primary and secondary consumers.
1933 Finally we briefly discuss the issue of food co-occurrence with fish consumers.
1934

1935 *Food Availability:* Estuaries are commonly characterized as highly productive nursery areas for
1936 a suite of organisms. Nixon (1988) noted that there actually is a broad continuum of primary
1937 productivity levels in different estuaries, which in turn affects fish yield. Compared to other
1938 estuaries, pelagic phytoplankton biomass and primary productivity in the upper San Francisco
1939 Estuary is poor (Cloern and Jassby 2008) and a low fish yield is expected (Figure 22).
1940 Understanding food webs is difficult. Conceptual models for estuarine food webs have
1941 progressed from simple nutrient-driven models to much more complex models incorporating
1942 nutrient cycling, light conditions, hydrodynamics, and grazing to adequately model primary

1943 production (Cloern 2001, Jassby 2008). Understanding of the Bay-Delta food web and drivers of
1944 the food web have advanced rapidly in the last several years.

1945
1946 Productivity of estuarine ecosystems is broadly believed to be fueled by a detritus-based food
1947 web. In the San Francisco Estuary, much of the community metabolism in pelagic waters does
1948 result from microbial consumption of organic detritus. However, evidence suggests that
1949 metazoan production in pelagic waters is primarily driven by phytoplankton production (Sobczak
1950 et al. 2002, 2005, Mueller-Solger et al. 2002, 2006, Kimmerer et al. 2005). Protists (flagellates
1951 and ciliates) consume both microbial and phytoplankton prey (Murrell and Hollibaugh 1998,
1952 York et al 2010) and are an additional important food source for many copepod species in the
1953 estuary (Rollwagen-Bollens and Penry 2003, Bouley and Kimmerer 2006, Gifford et al. 2007,
1954 MacManus et al. 2008). However, the conversion of dissolved and particulate organic matter to
1955 microbial biomass available to zooplankton is a relatively slow and inefficient process. Thus,
1956 shifts in phytoplankton and microbial food resources for zooplankton might favor different
1957 zooplankton species. The recognition that phytoplankton production might impose limits on
1958 POD species through food availability has led to intense interest in factors affecting
1959 phytoplankton production and species composition.

1960
1961 In the 1970s, highest phytoplankton standing stock and primary production generally occurred in
1962 the Suisun Bay-Honker Bay region in association with the entrainment zone (Ball and Arthur
1963 1979), which is now generally referred to as the low salinity zone. Since those early studies,
1964 there has been a significant long-term decline in phytoplankton biomass (chlorophyll *a*) and
1965 primary productivity (estimated from measurements of chlorophyll *a* and of water column light
1966 utilization efficiency) to very low levels in the Suisun Bay region and the lower Delta (Jassby et
1967 al. 2002). Jassby et al. (2002) detected a 47% decline in June–November chlorophyll *a* and a
1968 36% decline in June–November primary production between the periods 1975–1985 and 1986–
1969 1995. This decline was associated with changes in relationships between X2 and various
1970 metazoan populations (see below for details), including some pelagic fishes; however, the
1971 decline occurred well before the recent POD declines. Jassby (2008) updated the phytoplankton
1972 analysis to include the more recent data (1996–2005) from the Delta and Suisun Bay. Jassby
1973 (2008) confirmed a long-term decline in chlorophyll *a* from 1975 to 2005 but also found that
1974 March–September chlorophyll *a* had an increasing trend in the Delta from 1996 to 2005. Suisun
1975 Bay did not exhibit any trend during 1996–2005. A similar pattern was noted for primary
1976 production in the Delta. These chlorophyll *a* patterns continued to hold through 2008 according
1977 to a more recent study by Winder and Jassby (2010). In summary, phytoplankton biomass and
1978 production in the Delta and Suisun Bay seem to have reached a low point by the end of the
1979 1987–1994 drought. While they recovered somewhat in the Delta, chlorophyll *a* stayed
1980 consistently low in Suisun Bay through the POD years. Jassby (2008) did not analyze primary
1981 production for Suisun Bay because of evidence of inhibition of primary production in Suisun
1982 Bay associated with ammonium (see below for details; Dugdale et al. 2007). Hence, low and
1983 declining primary productivity in the estuary is likely a principal cause for the long-term pattern
1984 of relatively low and declining biomass of pelagic fishes in the estuary but not for the recent
1985 POD declines.

1986
1987 A major reason for the long-term phytoplankton reduction in the upper estuary after 1985 is
1988 benthic grazing by the invasive overbite clam (*C. amurensis*) (Alpine and Cloern 1992), which

1989 became abundant by the late 1980s (Kimmerer 2002). The overbite clam was first reported from
1990 San Francisco Estuary in 1986 and it was well established by 1987 (Carlton et al. 1990). Prior to
1991 the overbite clam invasion, there were periods of relatively low clam biomass in the upper
1992 estuary because the invasive Asiatic freshwater clam (*Corbicula fluminea*) (introduced in the
1993 1940s) colonized Suisun Bay during high flow periods and the native marine clam *Mya arenaria*
1994 (also known as *Macoma balthica*) colonized Suisun Bay during prolonged (> 14 month) low
1995 flow periods (Nichols et al. 1990). Thus, there were periods of relatively low clam grazing rates
1996 while one species was dying back and the other was colonizing. The *C. amurensis* invasion
1997 changed this formerly dynamic clam assemblage because *C. amurensis*, which is tolerant of a
1998 wide range of salinity, can maintain large, permanent populations in the brackish water regions
1999 of the estuary. Petersen and Vayssières (2010) analyzed 27 years of benthic data from Grizzly
2000 Bay and three other long-term monitoring stations and documented the establishment and
2001 expansion of *C. amurensis* during the 1987–1994 drought. The drought provided low-flow/high-
2002 salinity conditions that favored establishment. The population has persisted through subsequent
2003 high-flow/low-salinity years. In addition, the grazing influence of *C. amurensis* extends into the
2004 Delta (Kimmerer and Orsi 1996, Jassby et al. 2002) beyond the clam’s typical range, presumably
2005 due to tidal dispersion of phytoplankton-depleted water.

2006
2007 Shifts in nutrient concentrations and ratios may also contribute to the phytoplankton reduction
2008 and changes in algal species composition in the San Francisco Estuary. Phytoplankton
2009 production in the San Francisco Estuary is generally light limited with nutrient concentrations
2010 exceeding concentrations limiting primary production. Dugdale et al. (2007) and Wilkerson et al.
2011 (2006) found that high ammonium concentrations prevented the formation of diatom blooms but
2012 stimulated flagellate blooms in the lower estuary. This occurs because diatoms preferentially
2013 utilize ammonium in their physiological processes even though it is used less efficiently. Thus,
2014 diatom populations must consume available ammonium before nitrate, which supports higher
2015 growth rates, can be utilized. Glibert (2010) analyzed long-term data (from 1975 or 1979 to
2016 2006 depending on the variable considered) from the Delta and Suisun Bay and related changing
2017 forms and ratios of nutrients, particularly changes in ammonium, to declines in diatoms and
2018 increases in flagellates and cyanobacteria. Similar shifts in species composition were noted by
2019 Brown (2009). More recently, Parker et al. (in review) found that the suppression of algal
2020 blooms extends upstream into the Sacramento River to the SRWTP, the source of the majority of
2021 the ammonium in the river (Jassby 2008). Parker et al. (submitted) found that at high ambient
2022 ammonium concentrations, river phytoplankton cannot efficiently take up any form of nitrogen
2023 including ammonium, leading to often extremely low biomass in the river. A study using
2024 multiple stable isotope tracers (C. Kendall, USGS, personal communication) found that the
2025 cyanobacteria *M. aeruginosa* utilized ammonium, not nitrate, as the primary source of nitrogen
2026 in the central and western Delta. The SRWTP reduced its discharge by 12% starting in May 2009
2027 (S. Dean, Sacramento Regional County Sanitation District, personal communication). This
2028 reduction likely contributed to the relatively low ambient Sacramento River ammonium
2029 concentrations observed in spring 2010 and may have led to subsequent unusually strong spring
2030 diatom blooms in Suisun Bay (R. Dugdale, CSUSF-RTC, personal communication). Only a very
2031 muted summer *Microcystis* bloom occurred in 2010 (C. Mioni, UCSC, personal communication)
2032 which might have been due to lower summer water temperatures than in previous years, but the
2033 causes for this have not yet been fully investigated.

2034

2035 Ammonium concentrations in the Delta and Suisun Bay have significantly increased over the last
2036 few decades, due largely to increased loading from the SRWTP (Jassby 2008). Van
2037 Nieuwenhuysse (2007) found that a rapid reduction in wastewater total phosphorus loads in the
2038 mid-1990s coincided with a similarly rapid drop in phytoplankton biomass at three stations in the
2039 upper estuary. Van Nieuwenhuysse (2007) explored the effects of delta inflow and light but did
2040 not address the possible effects of Delta hydrodynamics or grazing on the observed relationships.
2041 It is also unclear how the three stations analyzed relate to the Delta as a whole.

2042
2043 Jassby (2008) conducted a more comprehensive assessment of factors affecting primary
2044 production and suggested the following comprehensive explanation for his observations.
2045 Phytoplankton production in the lower Delta is associated with flow and residence time;
2046 however, other factors introduce a substantial degree of interannual variability. Benthic grazing
2047 by *C. fluminea* is likely a major factor (Lucas et al. 2002, Lopez et al. 2006) but data are
2048 inadequate for a quantitative evaluation of the hypothesis. In Suisun Bay, benthic grazing by *C.*
2049 *amurensis* is a controlling factor that keeps phytoplankton at low levels. Thus, metazoan
2050 populations in Suisun Bay are dependent on importation of phytoplankton production from the
2051 upstream portions of the Delta. Ammonium concentrations and water clarity have increased;
2052 however, these two factors should have opposing effects on phytoplankton production. These
2053 factors likely also contribute to variability in the interannual pattern but the relative importance
2054 of each is unknown. The interactions among primary production, grazing, and transport time can
2055 be complex (Lucas et al. 2002, 2009a,b)

2056
2057 The invasion and establishment of the overbite clam *C. amurensis* during the 1987–1994 drought
2058 was also accompanied by a series of major changes in consumers. Many of these changes likely
2059 negatively influenced pelagic fish production. Some of these changes may have been directly
2060 caused or at least exacerbated by the clam. For example, a major step-decline was observed in
2061 the abundance of the copepod *E. affinis* possibly due to predation by the overbite clam
2062 (Kimmerer et al. 1994) or indirect effects on copepod food supply. Predation by *C. amurensis*
2063 may also have been important for other zooplankton species (Kimmerer 2008). Northern
2064 anchovy abandoned the low salinity zone coincident with the *C. amurensis* invasion, presumably
2065 because the clam reduced planktonic food abundance to the point that occupation of the low-
2066 salinity waters was no longer energetically efficient for this marine fish (Kimmerer 2006).
2067 Similarly, longfin smelt shifted its distribution toward higher salinity in the early 1990s, also
2068 presumably because of reduced pelagic food in the upper estuary (Fish et al. 2009). There was
2069 also a major step-decline in mysid shrimp in 1987–1988, likely due to competition with the
2070 overbite clam for phytoplankton (Orsi and Mecum 1996). Mysid shrimp had been an extremely
2071 important food item for larger fishes like longfin smelt and juvenile striped bass; its decline
2072 resulted in substantial changes in the diet composition of these and other fishes (Feyrer et al.
2073 2003, Bryant and Arnold 2007). As described above, the population responses of longfin smelt
2074 and juvenile striped bass to winter–spring outflows changed after the *C. amurensis* invasion.
2075 Longfin smelt relative abundance was lower per unit outflow after the overbite clam became
2076 established (Kimmerer 2002b). Young striped bass relative abundance stopped responding to
2077 outflow altogether (Sommer et al. 2007). One hypothesis to explain these changes in fish
2078 population dynamics is that lower prey abundance reduced the system carrying capacity
2079 (Kimmerer et al. 2000, Sommer et al. 2007).

2080

2081 Zooplankton species composition, abundance, and size in the Delta and Suisun Bay have
2082 changed tremendously over the last four decades. Mysid and copepod species present four
2083 decades ago have largely been replaced by newly introduced species. While some of these new
2084 species have reached high abundance levels, overall zooplankton abundance, biomass, and size
2085 have declined markedly and have remained at low levels for the last two decades (Winder and
2086 Jassby 2010). Food limitation and predation pressure by the overbite clam and possibly other
2087 invasive species are likely explanations for these trends (Winder and Jassby 2010).
2088

2089 *P. forbesi*, a calanoid copepod that was first observed in the estuary in the late 1980s, has
2090 replaced *E. affinis* as the most common delta smelt prey during the summer. It may have a
2091 competitive advantage over *E. affinis* due to its more selective feeding ability. Selective feeding
2092 may allow *P. forbesi* to utilize the remaining high-quality algae in the system while avoiding
2093 increasingly more prevalent low-quality and potentially toxic food items such as *M. aeruginosa*
2094 (Mueller-Solger et al. 2006, Ger et al. 2010b). After an initial rapid increase in abundance, *P.*
2095 *forbesi* declined somewhat in abundance from the early 1990s in the Suisun Bay and Suisun
2096 Marsh region but maintained its abundance, with some variability, in the central and southern
2097 Delta (Winder and Jassby 2010). Although substantial uncertainties about mechanisms remain,
2098 the decline in the Suisun region may be related to increasing recruitment failure and mortality in
2099 in this region due to competition and predation by *C. amurensis*, contaminant exposures, and
2100 entrainment of source populations in the Delta (Mueller-Solger et al. 2006, Winder and Jassby
2101 2010, J. Durand, UCD, unpublished data).
2102

2103 The abundance of a more recent invader, the cyclopoid copepod *Limnoithona tetraspina*,
2104 significantly increased in the Suisun Bay region beginning in the mid-1990s. It is now the most
2105 abundant copepod species in the Suisun Bay and confluence region of the estuary (Bouley and
2106 Kimmerer 2006, Winder and Jassby 2010). Gould and Kimmerer (2010) found that it grows
2107 slowly and has low fecundity. Based on these findings they concluded that the population
2108 success of *L. tetraspina* must be due to low mortality and that this small copepod may be able to
2109 avoid visual predation to which larger copepods are more susceptible. It has been hypothesized
2110 that *L. tetraspina* is an inferior food for pelagic fishes including delta smelt because of its small
2111 size, generally sedentary behavior, and ability to detect and avoid predators (Bouley and
2112 Kimmerer 2006, Gould and Kimmerer 2010). Nevertheless, this copepod has been found in the
2113 guts of delta smelt (S. Slater, CDFG, unpublished data). Recent experimental studies addressing
2114 this issue suggest that larval delta smelt will consume and grow on *L. tetraspina*, but growth is
2115 less than with *P. forbesi* (L. Sullivan, SFSU-RTC, unpublished data). It remains unclear if
2116 consuming this small prey is energetically beneficial for delta smelt at all sizes or if there is a
2117 breakpoint above which larger delta smelt receive little benefit from such prey. *Acartiella*
2118 *sinensis*, a calanoid copepod species that invaded at the same time as *L. tetraspina*, also reached
2119 considerable densities in Suisun Bay and the western Delta over the last decade (Hennessy
2120 2010). Its suitability as food for pelagic fish species remains unclear, but is also being
2121 investigated (L. Sullivan, SFSU-RTC, unpublished data).
2122

2123 Preliminary information from studies on pelagic fish growth, condition, and histology provide
2124 additional evidence for food limitation in pelagic fishes in the estuary (IEP 2005). In 1999 and
2125 2004, residual delta smelt growth was low from the Sacramento-San Joaquin confluence through
2126 Suisun Bay relative to other parts of the system. Delta smelt collected in 2005 from the

2127 Sacramento-San Joaquin confluence and Suisun Bay also had high incidence of liver glycogen
2128 depletion, a possible indicator of food limitation (Bennett et al. 2008). As previously noted,
2129 warm water temperatures during the summer period may have exacerbated lack of food by
2130 raising metabolic rate of delta smelt. Based on their entire suite of delta smelt data from 2005
2131 (histopathology, date of birth from otoliths, and growth rates from otoliths). Bennett et al.
2132 proposed a novel strategy for delta smelt survival in 2005. Natural selection appeared to favor
2133 individuals with a specific set of characters, including relatively slow larval development, but
2134 faster than average juvenile growth in July: a period with extremely high water temperatures and
2135 salinity encroachment. Many of these fish surviving into the pre-adult stage had also hatched
2136 earlier in the spawning season (i.e. before May). During 2003 and 2004 striped bass condition
2137 factor decreased in a seaward direction from the Delta through Suisun Bay (Gartz and Vu 2006).

2138
2139 Thus far, there is little evidence that the unusually poor growth rates, health, and condition of
2140 fishes from Suisun Bay and western Delta are due directly to the effects of toxic contaminants or
2141 other adverse chemical or physical habitat conditions. Our working hypothesis is that the poor
2142 fish growth and condition in the upper estuary are due to food limitation. However, data from
2143 Bennett et al. (2008), suggest that stressful environmental conditions may be an important
2144 contributing factor. Pollutants may be contributing to poor phytoplankton growth (Dugdale et al.
2145 2007) and invertebrate mortality (Werner et al 2010a, b), which could exacerbate food limitation.
2146 If fishes are food limited in Suisun Bay and western Delta during larval and/or juvenile
2147 development, then we would expect greater cumulative predation mortality, higher disease
2148 incidence, and consequently low abundance indices at later times.

2149
2150 *Food Quality:* Studies on food quality have been relatively few in the San Francisco Estuary,
2151 with even less information on long-term trends. However, food quality may be another limiting
2152 factor for pelagic zooplankton and their fish predators.

2153
2154 At the base of the pelagic food web, food quality for consumers is determined by the relative
2155 contributions of different phytoplankton and microbial species and detritus to the overall organic
2156 particle pool available to primary consumers. For example, diatoms and cryptophytes are thought
2157 to be good quality food sources for zooplankton, while the nutritional value of cyanobacteria
2158 such as *M. aeruginosa* can be very low (Brett and Müller-Navarra 1997), particularly for toxic
2159 varieties (Rohrlack et al. 2005). Several studies have documented shifts in phytoplankton species
2160 composition in the upper San Francisco Estuary from dominance by larger cells and diatoms to
2161 dominance by smaller cells and flagellates (Lehman 1996, 2000, Brown 2009, Glibert 2010)..
2162 Mueller-Solger et al. (2006) found that in recent years, diatoms were most abundant in the
2163 southern San Joaquin River region of the Delta, and Lehman (2007) found greater diatom and
2164 green algal contributions upstream and greater flagellate biomass downstream along the San
2165 Joaquin River. To date, the *M. aeruginosa* blooms have occurred most intensively in the central
2166 Delta, thus POD species that utilize the central Delta such as threadfin shad, striped bass, and the
2167 poorly monitored centrarchid populations (largemouth bass and sunfish) would be most likely to
2168 suffer any direct adverse effects of these blooms.

2169
2170 *Microcystis* is more generally more abundant in dry years compared to wet years with high
2171 abundances in 2007 and 2008 (P. Lehman, DWR, unpublished data). During the low flow
2172 conditions of summer 2007, blooms of this cyanobacterium spread downstream to the west Delta

2173 and beyond (P. Lehman, DWR, unpublished data). The highest cell densities were observed near
2174 Antioch, considerably west of the previous center of distribution, and may thus have affected fish
2175 in the confluence and Suisun Bay regions of the upper estuary. Although ambient concentrations
2176 of microcystins are unlikely to cause acute toxicity to copepods, chronic or episodic effects on
2177 populations are possible; however, copepods recover rapidly once the microcystins dissipate
2178 (Ger et al. 2009). Consumption of *M. aeruginosa* by copepods can cause mortality, even when a
2179 low percentage of the diet ($\geq 10\%$) and whether or not the strain consumed produced toxic
2180 microcystins (Ger 2010a); however *P. forbesi* appeared better able to selectively avoid *M.*
2181 *aeruginosa* when feeding. Therefore, the *M. aeruginosa* bloom may have indirectly affected the
2182 food supply for POD species, primarily threadfin shad and delta smelt, and other biota.

2183
2184 Other factors besides *M. aeruginosa* can also affect food quality for zooplankton and potentially
2185 affect food quantity and quality for higher trophic levels. In general, phytoplankton carbon
2186 rather than the much more abundant detrital carbon are thought to fuel the food web in the San
2187 Francisco Estuary (Mueller-Solger et al. 2002; Sobczak et al. 2002, 2005); however, that does
2188 not mean the detrital pathways are not significant because many zooplankton are omnivorous
2189 and capable of utilizing both pathways. For example, Rollwagen-Bollens and Penry (2003)
2190 observed that while heterotrophic ciliates and flagellates were the dominant prey of *Acartia* spp.
2191 in the bays of the San Francisco Estuary, diatoms and autotrophic ciliates and flagellates also
2192 formed an important part of their diet during phytoplankton blooms. Calanoid copepod and
2193 cladoceran growth and egg production may often be limited by low levels of phytoplankton
2194 biomass. This appears to be true even for omnivorous calanoids such as *Acartia* spp. Kimmerer
2195 et al. (2005) found a significant relationship between *Acartia* spp. egg production and
2196 chlorophyll *a* concentration in the San Francisco Estuary, suggesting that *Acartia* spp. likely also
2197 derived a large part of carbon and energy from phytoplankton. Bouley and Kimmerer (2006), on
2198 the other hand, reported that egg production rates of the cyclopoid copepod *L. tetraspina* were
2199 unrelated to chlorophyll *a* concentrations in the low salinity region of the San Francisco Estuary.
2200 Gifford et al. (2007) reported that larger zooplankton in the estuary are often omnivorous,
2201 and that smaller zooplankton, especially ciliates, are an important diet component for
2202 mesozooplankton in the estuary. In both studies, *L. tetraspina* clearance rates were highest for
2203 ciliates and flagellates, suggesting a greater importance of the detrital carbon pathway for this
2204 species. The dichotomy between phytoplankton and detrital/microbial energy pathways
2205 supporting zooplankton has probably been applied more stringently than is appropriate. Both are
2206 likely important, with the balance between them in specific areas of the estuary likely having
2207 effects on the success of particular zooplankton species. Additional research into the detrital
2208 pathway, especially the link to zooplankton through ciliates and other microbes, might be useful
2209 in understanding the factors controlling zooplankton populations, which are critical food
2210 resources for pelagic fishes in the estuary. Furthermore, the nutritional effects of an increasingly
2211 detrital-based food web also need to be explored in more depth.

2212
2213 In a study focusing on the nutrition and food quality of the calanoid copepods *E. affinis* and *P.*
2214 *forbesi*, Mueller-Solger et al. (2006) found evidence for “trophic upgrading” of essential fatty
2215 acids by *E. affinis* and *P. forbesi*, confirming their importance as high-quality food for fish. They
2216 also found that *E. affinis* gained the greatest nutritional benefits from varied food sources present
2217 in small tidal sloughs in Suisun Marsh. *P. forbesi*, on the other hand, thrived on riverine
2218 phytoplankton in the southern Delta, especially diatoms. Diatoms are likely also an important

2219 food source for other calanoid copepod species. The relative decrease in diatom contributions to
2220 the phytoplankton community in the central/western Delta and Suisun Bay (Lehman 1996, 2000,
2221 Brown 2010, Glibert 2010) is thus a concern and may help explain the declines in *P. forbesi* and
2222 other calanoid copepods in these areas. Mueller-Solger et al. (2006) concluded that areas rich in
2223 high-quality phytoplankton and other nutritious food sources such as the southern Delta and
2224 small tidal marsh sloughs may be critical “source areas” for important fish prey organisms such
2225 as *P. forbesi* and *E. affinis*. This is consistent with results of Durand (UC Davis, unpublished
2226 data) who showed that transport from upstream was essential for maintaining the *P. forbesi*
2227 population in Suisun Bay. Finally, as already mentioned above, zooplankton size has
2228 significantly decreased in the estuary over the past four decades. This has likely reduced the
2229 catchability of zooplankton by visually oriented planktivorous fish with negative energetic
2230 consequences for the fish (Winder and Jassby 2010).

2231
2232 *Food Co-occurrence:* Recently, interest has focused on determining patterns of co-occurrence of
2233 fish predators and their zooplankton prey. The assumption is that for successful predation,
2234 predators should co-occur with their prey. This idea was first explored by Nobriga (2002) who
2235 showed that delta smelt larvae with food in their guts typically co-occurred with higher calanoid
2236 copepod densities than larvae with empty guts. Kimmerer (2008) showed a positive relationship
2237 between delta smelt survival from summer to fall and zooplankton biomass in the low salinity
2238 zone (Figure 23). Miller and Mongan (unpublished data) concluded that April and July co-
2239 occurrence of delta smelt and copepod prey is a strong predictor of juvenile delta smelt survival.
2240 Mueller-Solger (DSP, unpublished data) defined delta smelt habitat based on the habitat
2241 suitability results of Nobriga et al. (2008) and defined the prey spectrum more broadly (as all
2242 copepods) compared to Miller and Mongan (unpublished data). Using these assumptions,
2243 Mueller-Solger found no long-term decline in the total biomass of copepods potentially available
2244 for consumption by delta smelt in midsummer, although species composition has changed
2245 considerably.

2246
2247 There are two major problems for co-occurrence analyses using available monitoring data. First,
2248 it is difficult to characterize fish prey suitability. For instance, *E. affinis* and *P. forbesi* are
2249 generally believed to be preferred prey items for delta smelt (Nobriga 2002, Miller and Mongan
2250 unpublished data). However, diet data show that delta smelt will actually feed on a wide variety
2251 of prey (Lott 1998, S. Slater, CDFG, unpublished data; Figure 24). Thus, the question of prey
2252 co-occurrence involves questions of prey catchability (e.g., Meng and Orsi 1991) and
2253 profitability (energy per item consumed and nutritional quality of individual prey items). For
2254 example, *L. tetraspina* has a large biomass in the system but individual *L. tetraspina* are smaller
2255 and possibly more evasive than the larger calanoid copepods. The energy used by an individual
2256 delta smelt to harvest a similar biomass of *L. tetraspina* compared to the energy used to harvest a
2257 larger species could be very different, as suggested by optimal foraging theory (e.g., Stephens
2258 and Krebs 1986).

2259
2260 The second major problem is that IEP sampling programs sample fish and zooplankton at larger
2261 spatial and temporal scales than those at which predator-prey interactions occur. Both fish and
2262 copepods are likely to be patchy and the long tows required to collect sufficient numbers of
2263 organisms for counting would homogenize such patch structure. Moreover, it is unlikely that the
2264 (monthly or even biweekly) “snapshot” of fish and prey co-occurrence in specific locations or

2265 even small regions provided by the IEP surveys is representative of feeding conditions actually
2266 experienced by fish on an hourly or daily basis.

2267
2268 *Summary:* The weight of evidence strongly supports bottom-up food limitation as a driver
2269 influencing long-term fish trends in the upper estuary. However, the bottom-up hypothesis is
2270 unlikely as a *single* mechanism for the recent POD decline for several reasons. First,
2271 phytoplankton and zooplankton declines preceded the POD, with lowest abundance and biomass
2272 levels in the 1980s and 1990s (Winder and Jassby 2010). Second, *C. amurensis* levels during the
2273 POD are not unprecedented; they are similar to those found during the 1987–1994 drought years,
2274 so it is unclear if and why benthic grazing would have a greater effect on the Suisun Bay food
2275 web during the POD years than during the earlier drought years. Finally, the hypothesis that the
2276 San Francisco Estuary is driven by phytoplankton production rather than through detrital
2277 pathways (Sobczak et al. 2002, 2005, Mueller-Solger et al. 2002) may have been accepted too
2278 strictly. Many zooplankton are omnivorous and can consume microbes utilizing dissolved and
2279 particulate organic carbon. This has recently been demonstrated for several zooplankton species
2280 in the San Francisco Estuary (Gifford et al. 2007 and references therein). Thus, shifts in
2281 availability of phytoplankton and microbial food resources for zooplankton might favor different
2282 species. It is possible that a better understanding of shifts in phytoplankton and zooplankton
2283 community composition and perhaps related changes in the microbial food web in the Suisun
2284 Bay region could explain these apparent inconsistencies.

2285

2286 **Species-specific Models**

2287
2288 The basic conceptual model provides a useful context for the major drivers likely affecting the
2289 POD species. However, it has limited value for helping policy makers and managers identify
2290 actions or guide research studies for individual POD species because it does not show how the
2291 major drivers differ for each species, and how they differ in relative importance during different
2292 life history stages or seasons. In response to these shortcomings, we developed initial species-
2293 specific models for each of the four POD fishes in 2007 (Baxter et al. 2008) and we update them
2294 here. The degree of detail for each model varies substantially based on the available data, so the
2295 degree of confidence in each is not consistent. Nonetheless, we believe that these models are an
2296 effective way to conceptualize the effects of different drivers and how they interact in time and
2297 space, which helps identify research and management priorities. Like the basic conceptual
2298 model, the species-specific models will continue to evolve as more information becomes
2299 available. We have also attempted to be consistent with the draft DRERIP (Delta Regional
2300 Ecosystem Restoration Implementation Plan; <http://www.dfg.ca.gov/ERP/DRERIP.asp>)
2301 conceptual models for delta smelt and longfin smelt, which are still in revision.

2302

2303 As with the basic conceptual model, we relied on a qualitative *weight of evidence* approach
2304 (Burkhardt-Holm and Scheurer 2007) in constructing the species-specific models. We identified
2305 the most *plausible* linkages between drivers and fish life stages based on our evaluation of all
2306 available POD laboratory results, long-term monitoring data, correlations, models, and our
2307 understanding of how the estuary functions. Here, we present the outcome of this approach. We
2308 do not describe all steps of the process itself or reiterate the description of all factors and results
2309 already covered in the basic conceptual model sections in each of the four species-specific

2310 model. In these models, we also generally deemphasize individual study results, including simple
2311 correlations between fish abundance and individual drivers. As with the basic conceptual model,
2312 we expect additional changes to these models as new information is collected. Some drivers
2313 that are deemphasized in the current species-specific models may play a more prominent role in
2314 future versions of these models.

2315
2316 The graphic representation of each species-specific model consists of four panels, each
2317 representing a portion of the species' life history (Figure 4–7) in relation to season and
2318 approximate center of distribution in the estuary. The arrows show progression through the life
2319 cycle in a clockwise direction. The seasons for each species generally correspond to the
2320 traditional definitions of summer (June – August), fall (September – November), winter
2321 (December – February), and spring (March – May); however, the exact timing of life history
2322 events varies somewhat from year to year, depending on environmental conditions. Within each
2323 panel, major drivers are shown in red boxes and proposed mechanisms by which they may affect
2324 the fish population are shown in yellow boxes. Below, we provide narrative descriptions of
2325 each species-specific model. Each narrative description consists of a general section followed by
2326 narratives for each season. The general section briefly summarizes important aspects of the four
2327 main basic conceptual model components for each species and explores drivers of particular
2328 concern. The narratives for each season highlight drivers particularly relevant to each species in
2329 that season and also include spatial considerations.

2330

2331 *Delta Smelt*

2332

2333 We hypothesize that degradation of habitat is the fundamental cause of delta smelt decline and
2334 that it affects the species mainly through effects on growth and subsequent reproductive potential
2335 rather than immediate mortality. Both abiotic and biotic aspects of habitat suitability have
2336 declined over time. This has led to smaller, less healthy adults, which have lower per capita
2337 fecundity. These ecosystem challenges have probably been exacerbated by periodic high
2338 entrainment loss. We hypothesize that habitat degradation has reduced carrying capacity. Thus,
2339 entrainment losses at historical levels could have increased in importance because the population
2340 is smaller. Large-scale water diversion may also influence delta smelt carrying capacity through
2341 seasonal effects on Delta outflow. The following conceptual model approaches factors
2342 influencing delta smelt season by season (Figure 4).

2343

2344 Delta smelt population dynamics have apparently changed over time with declining estuarine
2345 feeding and habitat conditions. In the early years of IEP sampling, no delta smelt stock-
2346 recruitment relationship was apparent (Moyle et al. 1992). However as sampling continued and
2347 abundance declined, a stock-recruitment relationship has emerged. Over the entire period of IEP
2348 sampling, delta smelt now show a fairly strong and linear (i.e., density-independent) stock-
2349 recruitment relationship (Figure 12). In contrast, the summer survival of delta smelt appears to
2350 have been a density-dependent life stage transition starting early in the record (Bennett 2005).
2351 Food availability is a likely limiting factor for this planktivorous fish as Kimmerer (2008) noted
2352 a statistically significant relationship between juvenile smelt survival and zooplankton biomass
2353 over the long term (Figure 23). Also consistent with the food limitation hypothesis is the decline
2354 in the mean size of adult delta smelt following the introduction of the clam *C. amurensis*

2355 (Sweetnam 1999, Bennett 2005), which caused declines in key zooplankton prey. Bennett et al.
2356 (2008) noted complex interactions between life history traits and environmental conditions with
2357 the summer period being a particularly stressful time for juvenile delta smelt. These
2358 observations are consistent with Nobriga et al. (2008), who noted a strong reduction in
2359 probability of capture at water temperatures that approached the laboratory-derived acute lethal
2360 limit of 25°C (for fish acclimated to 17°C; Swanson et al. 2000).

2361
2362 Statistical analyses of the long-term delta smelt trends confirm that in the early 2000s there was a
2363 step decline in the abundance of delta smelt (Manly and Chotkowski 2006, Thompson et al.
2364 2010). We propose that changes in water project operations and low egg supply resulting
2365 directly from low adult abundance and small size of surviving adults are contributing causes of
2366 this recent decline and a current impediment to rapid recovery. The population is now at such
2367 low levels that recovery is unlikely in a single year but will require several consecutive years
2368 with positive population growth rates. Thomson et al. (2010) concluded that the POD decline
2369 could not be explained by the variables that best explained abundance trends prior to the POD.
2370 Increased water project exports during winter resulted in higher losses of adult smelt, particularly
2371 early spawning fish (and their offspring) that may be proportionally more important to the
2372 population (Bennett et al. 2008). By contrast, reduced exports during spring may have increased
2373 survival of later-spawned larvae. Although these larvae may be saved from exports, their small
2374 size at the beginning of the stressful summer period likely leads to slow growth, reduced
2375 survival, and small size at maturity (i.e., lower fecundity) (Bennett et al. 2008). Reduced spring
2376 exports from the Delta since 2000 have been the result of VAMP, a program designed to benefit
2377 outmigrating juvenile Chinook salmon from the San Joaquin River system.

2378
2379 It is unclear whether the population has dropped below critical levels where Allee effects inhibit
2380 recovery. However, the observation that delta smelt have exhibited a declining trend in
2381 production of adults from the population of adults in the previous year (Figure 14) suggests that
2382 record low abundance levels may now be below the threshold for Allee effects. These concerns
2383 were reinforced again by the 2008 and 2009 FMWT delta smelt abundance indices of 23 and 17,
2384 respectively, which represent the lowest indices on record
2385 (<http://www.delta.dfg.ca.gov/data/mwt>).

2386
2387 The delta smelt has been considered semi-anadromous, but in recent years investigations
2388 centered on its northern Delta spawning and early rearing areas have detected delta smelt year-
2389 round, leading to the idea that these putative “resident” individuals might represent alternate life
2390 history contingents (Sommer et al. 2009, Sommer et al. in review). The southern end of the Yolo
2391 Bypass, including Liberty Island, Cache Slough, and the Sacramento deepwater ship channel are
2392 known to support delta smelt spawning and rearing (see Bennett 2005). During 2003 – 2005 the
2393 USFWS collected delta smelt during monthly sampling activities throughout the year, not just
2394 during spring time, suggesting that delta smelt were using this relatively shallow, flooded island
2395 habitat throughout their entire life cycle (USFWS, unpublished data). Similarly, extensions of
2396 the 20-mm Survey, TNS and FMWT surveys into the Sacramento deepwater ship channel caught
2397 delta smelt consistently from June through October, the warmest months of the year (CDFG
2398 unpublished data). Like the “core” rearing habitat of delta smelt near the Sacramento-San
2399 Joaquin River confluence, Liberty Island and adjacent deeper habitats in the Ship Channel and
2400 Cache Slough are very turbid and have very little SAV (Nobriga et al. 2005, Lehman et al. 2010,

2401 CDFG, unpublished data). However, Liberty Island is somewhat warmer during the summer
2402 than the river confluence (Nobriga et al. 2005) and may prove to be a challenging habitat for
2403 rearing. The following conceptual model applies only to the traditional view of delta smelt as a
2404 semi-anadromous species. We are currently evaluating how to integrate these observations into
2405 our conceptual model (T. Sommer, DWR, unpublished data).
2406

2407 *Summer:* Summer is the season that usually has the highest primary and secondary productivity
2408 in a temperate zone estuary. Given its annual life cycle, summer is also the primary growing
2409 season for delta smelt. We propose that delta smelt growth rates are potentially limited during
2410 summer by high water temperature, low food quality and low food quantity.
2411

2412 Nobriga et al. (2008) found that the catch of delta smelt began decreasing at temperatures above
2413 20° C and became almost zero at 25°C suggesting avoidance of stressful conditions or high
2414 mortality. Temperatures near 25°C are likely to be near the lethal end of delta smelt tolerance
2415 (Swanson et al. 2000) and would certainly affect growth rates and metabolic activities after
2416 prolonged exposure. Jassby (2008) noted a significant temperature increase in the Delta in
2417 recent years, but cautioned that it was not the result of a long-term trend. Using data from 2005
2418 field observations and the culture facility, Bennett et al. (2008) created a specific growth curve
2419 with respect to water temperature (Figure 25), which reinforces the observations of Nobriga et al.
2420 (2008). Bennett et al. (2008) also noted that high July temperatures may have led to poor
2421 growth, cell damage and poor liver function of the juveniles examined.
2422

2423 Fishes require substantial food resources to maintain positive growth when they experience
2424 temperatures near their upper tolerances. However, the trend in the estuary has been toward
2425 lower summertime food quantity and potentially quality. Overbite clam grazing has reduced
2426 zooplankton availability (Kimmerer et al. 1994), but nutrient loading and water export may also
2427 impact zooplankton production. Copepod population dynamics are strongly affected by grazing
2428 pressure from *C. amurensis* (Kimmerer et al. 1994, Winder and Jassby 2010), resulting in fewer
2429 high quality calanoid copepods, the most common prey of juvenile delta smelt (Figure 24; Lott
2430 1998). Moreover, in the decade including the early POD years, there has been a further decline
2431 in the abundance of calanoid copepods in Suisun Bay and the western Delta (Kimmerer et al.
2432 2008). At the same time, these calanoid copepods are being replaced by the much smaller
2433 cyclopoid copepod *L. tetraspina*, which is presumed to be a less suitable prey species (L.
2434 Sullivan, CSUSF-RTC, unpublished data). A laboratory feeding study showed that larval delta
2435 smelt will consume *L. tetraspina*, and when given the choice between *L. tetraspina*, *P. forbesi*,
2436 and *E. affinis*, they will consume each in proportion to its abundance (L. Sullivan, CSUSF-RTC,
2437 unpublished data). *L. tetraspina* is also being consumed in the wild (Figure 24). However,
2438 Sullivan's research indicates that delta smelt grow more slowly when fed *L. tetraspina* compared
2439 to other prey. Thus, the decline in calanoid copepod abundance (and biomass) may not be
2440 compensated for by high abundance of *L. tetraspina*, even if biomass remains roughly the same
2441 (see Baxter et al. 2008; Figure 21) if growth is slowed by a diet of *L. tetraspina*.
2442

2443 Little has been published about the jellyfish in the upper estuary. They may compete with smelt
2444 for food resources (A. Wintzer and P. Moyle, UCD, unpublished data), but it is unlikely they are
2445 a direct source of mortality due to their small size and their absence during the delta smelt larval
2446 period in spring when smelt vulnerability would be highest. Jellyfish abundance in the upper

2447 estuary appears to peak in late summer or early fall and persist for 2–3 months thereafter (CDFG
2448 unpublished data).

2449
2450 The long-term reduction in calanoid copepod availability in the upper estuary has likely resulted
2451 in slower growth rates of delta smelt, as observed in otolith studies by Bennett et al. (2008) and
2452 in part, as a reduction in the mean size of delta smelt in fall (Sweetnam 1999; Bennett 2005).
2453 Baxter et al. (2008) previously hypothesized that over the long term, reduced summer growth
2454 rates due to reduced food availability during the thermally stressful summer period reduced the
2455 summer to fall survival of juvenile delta smelt; however, analysis of available IEP index data do
2456 not support this hypothesis for the initial POD period (2001–2004), but do suggest reduced
2457 summer to fall survival in 2005 and possibly subsequent years (Figure 13). Thus, increased
2458 mortality seems like a periodic or more recent occurrence, and the more substantial effect of
2459 reduced growth is smaller adults, which affects fecundity and reproductive potential, and
2460 possibly aspects of adult survival. These results suggest that summer food limitation remains a
2461 major stressor on delta smelt.

2462
2463 Habitat availability and suitability also may be indirectly influencing food quantity and quality.
2464 Specifically, the summer habitat of juvenile smelt has become restricted due to high water
2465 temperatures and decreased turbidity in the southern and eastern Delta, and saltwater intrusion
2466 in Suisun Bay (Nobriga et al. 2008). This has limited the area available for feeding. By late
2467 summer toxic algae blooms and possibly, competition with jellyfish could be affecting delta
2468 smelt.

2469
2470 When the toxic blue-green alga *M. aeruginosa* blooms during late summer (generally August and
2471 September), it occurs primarily on the San Joaquin River side of the Delta (Lehman et al. 2005)
2472 and most delta smelt occur to the north and west of the bloom's epicenter (Nobriga et al. 2008).
2473 Typically, the bloom only partially overlaps with delta smelt distribution. At concentrations
2474 found in the wild, microcystins have been shown to have only sublethal effects on delta smelt
2475 prey species in laboratory experiments (Ger et al. 2009). However, chronic and episodic effects
2476 on zooplankton remain possible. For example, Ger et al (2010 a) found negative dietary effects
2477 of *M. aeruginosa* cells on copepods. Thus, there may be indirect effects on food resources (see
2478 Food Quality section). In addition, it appears that potentially toxic effects of *M. aeruginosa* may
2479 propagate well beyond the observed bloom region, and particularly downstream. Lehman et al.
2480 (2010) found, for example, that striped bass and Mississippi silversides co-occurring with *M.*
2481 *aeruginosa* exhibited signs of toxic and carcinogenic effects, but so did many from areas
2482 apparently outside the bloom range. Similar results were obtained for some food organisms.
2483 The mechanism for this phenomenon is unknown. The work of Lehman et al. (2010) indicates a
2484 strong likelihood that delta smelt are also exposed to microcystins and the direct effects of *M.*
2485 *aeruginosa*.

2486
2487 Other water quality variables such as contaminants could be important, particularly to early life
2488 stages and when fish are already stressed by other factors (see Anderson et al. 2007). For
2489 example, larval delta smelt were found to be 3–10 times more sensitive than other fish species to
2490 ammonia, copper, and insecticides found in the Delta (Werner 2008). Isolated cases of short-
2491 term exposures may not be lethal, but exposure to such contaminants when fish are already
2492 stressed by high temperatures, osmotic stress, or other factors could result in mortality or

2493 degraded physiological condition (Brooks et al., submitted). Bennett et al. (2008) observed a
2494 small number of fish with ovatestis (intersex fish) during the stressful year of 2005.

2495
2496 In summary, there is evidence of bottom-up and habitat suitability effects on delta smelt during
2497 the summer over the long term, and these interactions may ultimately affect delta smelt
2498 production. The lack of observed salvage during the summer suggests SWP/CVP entrainment
2499 effects are minimal during this period. Improved habitat (including water quality) and food
2500 conditions during the summer would likely improve growth and survival as well as individual
2501 fitness of maturing delta smelt and ultimately their fecundity.

2502
2503 *Fall:* Fall represents the time period when the delta smelt year class completes its somatic
2504 growth and begins gonad development. Some summertime drivers continue to affect delta smelt
2505 into early fall before declining (e.g., high water temperature and *Microcystis* toxicity). Other
2506 drivers such as current water management practices, which lead to chronic low fall Delta outflow
2507 and landward X2 (Figure 26), likely play a key role in reducing delta smelt habitat suitability.
2508 Evidence to date indicates that fall habitat is a significant current (and future) issue affecting the
2509 abundance of delta smelt (Feyrer et al. 2007, 2010). Delta smelt are strongly associated with low
2510 salinity and high turbidity, which have been used to model the availability and suitability of its
2511 habitat (Feyrer et al. 2007, 2010). Fall habitat suitability has shown a long-term decline (Feyrer
2512 et al. 2007, 2010). Because these habitat suitability indices are derived from delta smelt catch
2513 data, they directly reflect population-level effects. In addition, the fall abundance index is the
2514 best single predictor of juvenile production the following year (Figure 12). Thus, the
2515 accumulation of factors that affect fall abundance (FMWT index) influence the abundance of the
2516 next year class.

2517
2518 Reduction of habitat area likely interacts with bottom-up and top-down mortality mechanisms to
2519 affect delta smelt survival. There are several potential mechanisms by which habitat area can
2520 affect delta smelt as described by Feyrer et al. (2010). In general, increased habitat area provides
2521 more space for individuals to safely live and reproduce. More specifically, increased habitat area
2522 presumably reduces the probability of density-dependent effects on the delta smelt population
2523 (e.g., food limitation, disease, and predation). Moreover, increased habitat area also presumably
2524 reduces the probability that stochastic, localized, catastrophic events will affect a sizable portion
2525 of the population. The geographic placement of the remaining habitat area likely also plays an
2526 important role. A key concern for delta smelt is that as habitat suitability declines to low levels,
2527 remaining habitat is centered on the western Delta in closer proximity to anthropogenic sources
2528 of mortality such as water diversions and certain contaminant sources such as agricultural runoff.
2529 Although, direct entrainment is not a major stressor during the fall, shrinking fall habitat that
2530 places delta smelt in closer proximity to the hydrodynamic effects of the export pumps is linked,
2531 along with negative OMR flow, to increased winter salvage (Grimaldo et al. 2009).

2532
2533 In summary, there is good evidence for reduction in habitat availability and suitability during the
2534 fall and a linkage of these reductions with abundance. Slow growth due to food limitation and
2535 physiological stress during summer may affect survival in fall, but the evidence points to poor
2536 growth in summer and fall, which likely contributes to the species decline via reduced size and
2537 fecundity of maturing fish.

2538

2539 *Winter:* Winter represents the main period of adult delta smelt upstream migration, sexual
2540 maturation, and the beginning of spawning. Delta smelt upstream migrations are associated with
2541 winter flow pulses and coincident increases in turbidity (Grimaldo et al. 2009). Conditions
2542 triggering migrations generally follow winter storms starting as early as late December and
2543 extending into February. By March, delta smelt appear to migrate even in absence of such flow-
2544 turbidity events (CDFG unpublished data). This movement is considered an active group
2545 behavior that is likely triggered by olfactory cues associated with changes in water quality or
2546 turbidity (Sommer et al. submitted). Once cues initiate migration, groups of smelt presumably
2547 move quickly and efficiently upstream by swimming with incoming tides until they reach their
2548 holding grounds (Sommer et al. submitted).

2549
2550 We hypothesize that winter entrainment of adults (top-down effects) can affect the delta smelt
2551 population in some years. Salvage and corresponding population loss estimates of delta smelt
2552 increased during the POD years (2000–2005) in comparison to most of the previous decade
2553 (Kimmerer 2008, Grimaldo et al. 2009). Specifically, Kimmerer (2008) estimated that
2554 cumulative proportional population losses were between 3 and 50% for adult delta smelt during
2555 the winter export periods between 2002 and 2006, and concluded that the rather large
2556 proportional losses and their effects occurred episodically and therefore should be calculated
2557 rather than inferred from correlative analysis (e.g., Manly and Chotkowski 2006; Thomson et al.
2558 2010). The recent relatively high adult proportional entrainment in 2002–2004 identified by
2559 Kimmerer (2008) likely negatively affected subsequent recruitment. Moreover, young delta
2560 smelt in the same years also suffered relatively high proportional losses. These repeated, paired,
2561 cross-generational losses, though small individually, could have had a more substantial
2562 cumulative effect on the delta smelt population; they coincide with a sharp decline in delta smelt
2563 abundance (Figure 2). Note that high winter exports may not have demonstrable effect on delta
2564 smelt salvage during periods of high delta outflow because OMR flows are rarely negative and
2565 fewer delta smelt are located within the hydrologic influence of the export facilities during high
2566 outflow periods (Sweetnam et al. 1999, Grimaldo et al. 2009). Similarly, during critical dry
2567 years when the southern Delta has high water transparency, entrainment risks for delta smelt
2568 appear low. There are several possible reasons for this. Under these conditions, most delta smelt
2569 likely migrate up to the northern Delta where turbidity is higher (Sommer et al. submitted).
2570 Also, delta outflow standards (i.e. Water Rights Decision 1641) and limited water in upstream
2571 reservoirs available for release, severely constrain exports so that OMR flows are rarely or only
2572 briefly strongly negative.

2573
2574 There is presently no evidence of habitat constriction or food limitation during this period;
2575 however, no studies have addressed these questions. Contaminant effects are possible during
2576 flow pulses because many agricultural and urban-related contaminants enter waterways with the
2577 first rain events (Kuivila and Foe 1995, Kuivila and Hladik 2008).

2578
2579 *Spring:* Bennett et al. (2008) propose that reduced spring exports resulting from VAMP (mid-
2580 April to mid-May) has selectively enhanced the survival of spring-hatched delta smelt larvae as
2581 compared to those hatching earlier. Initial otolith studies suggest that these spring-hatched fish
2582 dominate subsequent recruitment to adult life stages; by contrast, those delta smelt hatched prior
2583 to the VAMP have been poorly-represented in the adult stock in recent years. Bennett et al.
2584 (2008) further propose that the differential fate of early- and late-hatched cohorts may affect the

2585 sizes of delta smelt in fall because the later cohorts have a shorter period in which to grow. In
2586 addition, Bennett et al. (2008) found that fish in the early-spawned cohorts of 2005 were
2587 generally smaller at age than late-spawned cohorts. The early-spawned fish appeared to partition
2588 their growth during the first few weeks after hatchig. These early-spawned fish developed more
2589 slowly during the larval stage, but were able to maintain much higher growth rates during the
2590 juvenile phase, and in particular, the first two weeks of July. Because the early-spawned cohort
2591 grew slower during spring, they were potentially more vulnerable to predation. Also, because
2592 they reared in the Delta longer, they were vulnerable to exports for a longer period of time and
2593 were also vulnerable before the VAMP-related flow pulses and export decreases took place.
2594 Despite longer exposure times, these slower growing fish actually had higher survival
2595 probabilities once they reached summer and fall. Later-spawned fish hatching immediately
2596 before or during VAMP were usually comparatively protected from entrainment via reduced
2597 exports and more positive flows in southern Delta channels, but because of their high growth
2598 rates and correspondingly higher energy needs, they were less well equipped to handle extreme
2599 summer conditions in 2005, including low food and high temperatures (Bennett et al. 2008).
2600 Thus, current water management favors delta smelt larvae that hatch immediately before or
2601 during the mid-April to mid-May period of reduced exports; however, this advantage can be
2602 severely curtailed if spring conditions abruptly change to raise water temperatures and food
2603 resources simultaneously decline, as they did in 2005.

2604
2605 From the perspective of an individual female, choice of a spawning date is critical to larvae
2606 hatching into favorable biotic and abiotic conditions. Until recently, female delta smelt were
2607 believed to spawn once or repeatedly over a short period of time (Moyle 2002). Thus, the choice
2608 of when to spawn was believed to determine fitness to a large degree. Recently, both the
2609 laboratory and field based observations confirm that delta smelt have the potential to produce
2610 multiple clutches. Lindberg et al. (UCD, unpublished data) documented cases where 1-year-old
2611 wild-origin females produced 3 viable clutches of eggs under ideal (i.e., cold water, unlimited
2612 food) conditions over a period of several months. This information was further corroborated by
2613 S. Teh (UCD, unpublished data) who found histological evidence of repeat spawning by females
2614 collected in the wild. Thus, at least some females can potentially hedge their bets by spawning
2615 repeatedly throughout the spawning season. Repeat spawning likely has several prerequisites: (1)
2616 healthy adults with energy reserves to develop gonads early in the season and survive spawning;
2617 (2) ripe and spent fish are able to locate sufficient food resources to develop additional gametes;
2618 and (3) a protracted period of suitable spawning conditions to allow for the development and
2619 release of multiple sets of gametes. Egg survival is linked to temperature and temperatures
2620 associated with 50% or better survival to hatching range from just below 10°C to about 18°C
2621 (Bennett 2005). Bennett (2005) found that the presence of successive cohorts of post-larvae in
2622 the 20-mm Survey was linked to the duration of a 15–20°C temperature window and suggested
2623 that a similar temperature window for spawning might be the duration of 14–18°C. Thus,
2624 protracted spawning and recruitment windows relate to broader success of both individual
2625 spawning events and provide time necessary for some individuals to undergo several successful
2626 spawnings. Such a spawning strategy would greatly increase the probability of individual
2627 females producing at least one clutch of eggs that hatched under conditions favoring survival of
2628 young delta smelt and is more consistent with life-history theory for annual fishes (Winemiller
2629 and Rose 1992) than the semelparous spawning previously assumed (Moyle 2002).
2630

2631 Until recently, the numbers of larvae entrained in the south Delta export pumps were unknown
2632 because fish less than 20 mm long are not identified in salvage (see Brown et al. 1996).
2633 Kimmerer (2008) estimated that entrainment of larval smelt was greatest during April based on
2634 estimates of abundance from the 20-mm Survey, OMR flows, and risk calculations based on
2635 particle tracking models (PTM). During April, larval entrainment is not observable at the
2636 salvage facilities because fish less than 20 mm SL are not counted at the facilities. Juvenile delta
2637 smelt salvage peaks between May and June of most years (Grimaldo et al. 2009); however, the
2638 magnitude of entrainment is dependent on flow (Kimmerer 2008). Kimmerer (2008) estimated
2639 that larval and juvenile population losses ranged from 0 to about 25% from 1995 to 2006, with
2640 the highest losses occurring in dry years 2001–2004. This period of increased loss overlaps a
2641 similar period of relatively high loss for adults (see winter section above); thus, there was a
2642 period of sequential, intergenerational loss that may have had a cumulative effect on the delta
2643 smelt population. A complicating factor in estimating entrainment are prescreen losses of delta
2644 smelt. The magnitude of these losses is currently unknown, but according to a recent study (G.
2645 Castillo, USFWS, unpublished data) it is likely high. This is consistent with the results of
2646 previous studies of prescreen losses conducted with other species (Clark et al. 2009, Gingras et
2647 al. 1997).

2648
2649 Because of natural variability and delta outflow standards (i.e. Water Rights Decision 1641),
2650 there have been few significant long-term trends in upper estuary spring salinity (Figure 26,
2651 Enright and Culberson 2009). This suggests that it is unlikely that there have been any recent
2652 changes in spring abiotic habitat availability or suitability. Habitat effects based on calanoid
2653 copepod densities, contaminants or disease may have worsened during spring. The dietary
2654 importance of *E. affinis* and calanoid copepods in general suggests that declines in either might
2655 affect delta smelt survival and recruitment. Spring densities of *E. affinis* did not decline
2656 substantially during the POD years and even spiked upward in 2006 and 2008, and neither
2657 *Sinocalanus doerrii* nor *P. forbesi* exhibited consistent low abundance during the POD years
2658 (Hennessy 2008, 2010), so there was little evidence of spring food being limited. There has been
2659 little evidence of direct toxicity to delta smelt larvae, based on limited numbers of bioassays
2660 using water collected during spring from the upper estuary, though a few acute results were
2661 detected using samples from the lower Sacramento River at Hood and near Rio Vista (Werner et
2662 al. 2008, 2010). Lethal and sublethal effects on the invertebrate *Hyallolella azteca* were observed
2663 more often and at even more sites (Werner et al. 2008, 2010) suggesting possible sublethal
2664 effects on fishes, including delta smelt, through the food web, but this was not supported by
2665 calanoid copepod abundances (see Hennessy 2010). Upper estuary habitat does not appear to
2666 have declined based on the factors we were able to assess.

2667

2668 *Longfin smelt*

2669

2670 We hypothesize that winter–spring outflow, adult abundance and food availability most strongly
2671 influenced the long-term pattern of longfin smelt recruitment (Figure 5). X2 (or freshwater
2672 outflow) during the winter–spring spawning and rearing periods continues to exert a significant
2673 positive effect on abundance (year-class strength) (Figure 27; Stevens and Miller 1983, Jassby et
2674 al. 1995; Kimmerer 2002b; Sommer et al. 2007, Kimmerer et al. 2009). The historical
2675 relationship changed subsequent to the establishment of the overbite clam in 1987. The slope of

2676 the relation between X2 and the population remained the same after the introduction, but the
2677 intercept changed indicating that the abundance of fish expected at a specific X2 value declined
2678 significantly (Kimmerer 2002b). Reduced prey availability due to clam grazing is believed to be
2679 the mechanism for the change (Kimmerer 2002b). An additional change in the relationship
2680 occurred after 2002, particularly from 2003 through 2005, when abundance did not increase
2681 when outflow increased (Figure 27 a–c; Sommer et al. 2007, Baxter et al. 2008). After a large,
2682 abundance increase in 2006 in response to substantially higher outflow, the 2007 FMWT annual
2683 index declined to a record low of 13 and has remained low in 2008 and 2009
2684 (<http://www.delta.dfg.ca.gov/data/mwt>). Some decline in 2007 was expected due to low winter–
2685 spring outflows, but the 2007 index fell well below the post-clam FMWT outflow abundance
2686 relationship (Figure 27a) and represented a statistical outlier (Studentized residual = -2.802; R.
2687 Baxter, CDFG, unpublished data). The 2008 and 2009 abundance indices more closely fit the
2688 post-clam outflow abundance relationships (Figure 27 a–c). The mechanism(s) underlying the
2689 2003–2005 lack of longfin smelt abundance response to increased outflow remain(s) unknown as
2690 does the mechanism for the 2007 abundance response, though shifts in distribution away from
2691 habitat sampled by the midwater trawl may have an effect. A similar lack of response was not
2692 apparent in the Bay Study otter trawl relationship (Figure 27c) suggesting that a portion of the
2693 population continued to respond as it had in the recent past.

2695 Preliminary analyses support a stock-recruitment relationship between adults approaching their
2696 second birthday and age-0 fall recruits (The Bay Institute et al. 2007a). Moreover, longfin smelt
2697 abundance in the FMWT exhibits a significant autocorrelation based on a 2-year time lag ($r_{lag 2} =$
2698 0.486, 36 df, $p = 0.002$). Since longfin smelt typically spawn at the end of their second year of
2699 life (Baxter 1999, Moyle 2002, CDFG 2009a), this 2-year lag can be interpreted as additional
2700 evidence of a stock-recruitment relationship. Development of a revised, direct stock-recruitment
2701 relationship is in progress. A significant stage-recruitment relationship (fall age-0 to fall age-1
2702 abundance) also exists, but survival declined after 1994 (Rosenfield and Baxter 2007)
2703 presumably due to continued food limitation. The distributions of the two age classes differ
2704 (Baxter 1999, Rosenfield and Baxter 2007) so discussions of seasonal drivers will include both
2705 young of the year (age 0) and age-1 fish when relevant. Age-2 fish are generally only captured
2706 in the estuary during the winter spawning season, so they will not be discussed in detail.
2707 Declining juvenile recruitment and reduced stage-recruitment survival are important factors in
2708 the declining population trends of longfin smelt, and likely limit its positive response to
2709 favorable environmental conditions.

2710
2711 *Winter:* Upstream migration of mature adults and most spawning occurs in winter with
2712 spawning probably confined to freshwater portions of the estuary (Moyle 2002; Rosenfield and
2713 Baxter 2007). The distribution of longfin smelt has the greatest overlap with the distributions of
2714 the other POD fishes during this season. The geographic and water column distributions of
2715 adults and larvae in winter lead us to hypothesize that entrainment is having an important effect
2716 on the population during this season, particularly during low outflow years when a higher
2717 proportion of the population may spawn farther upstream in the Delta. A CDFG conceptual
2718 model of longfin smelt migration features adults moving up to and congregating in the low
2719 salinity zone (0.5–6 psu) as temperatures decline, starting in late fall (see CDFG 2009a, b).
2720 From here, ripe individuals are believed to make generally short-distance, brief spawning runs
2721 into freshwater where spawning takes place over a sand substrate. The fish are then believed to

2722 return to the low salinity zone if partially spent and farther down estuary when completely spent
2723 (CDFG 2009b). Individual fish are believed to be capable of spawning several times during the
2724 spawning period. When the low salinity zone (indexed by X2) is upstream in the Delta,
2725 increasing numbers of adult longfin smelt probably move into and upstream of the influence of
2726 the export pumps on their spawning migration. This hypothesis is supported by: (1) generally
2727 higher salvage during low outflow years (Sommer et al. 1997, CDFG 2009b); (2) winter catch
2728 density plots showing the population shifting upstream and downstream in concert with shifting
2729 X2 (CDFG 2009b); (3) and increasing winter salvage in relation to fall abundance with
2730 increasing X2 in winter (CDFG 2009b). Increased winter salvage of mostly adult longfin smelt
2731 after 2000 suggests that entrainment levels may have been higher during POD years (Figure 18;
2732 IEP 2005). This entrainment occurred when the adult population was at a fairly low level
2733 (Figure 2). Recent calculations of longfin smelt adult entrainment and loss showing relatively
2734 high adult entrainment and loss during winters of water years 2002–2004 (CDFG 2009b).
2735 However, these losses have yet to be placed into a population context, so we cannot provide
2736 conclusions regarding their recent effect on abundance. Combined winter exports were generally
2737 high from 2000 through 2005 creating a high upstream net flow toward the export pumps in
2738 OMR, known as negative OMR flow (CDFG 2009b). Grimaldo et al. (2009) evaluated the
2739 effects on salvage of a suite of environmental, hydrologic and biological variables, and winter
2740 adult salvage was most parsimoniously attributable to strong negative OMR flows. Thus, X2
2741 and negative OMR flows can be interpreted as distal and proximal factors, respectively,
2742 influencing longfin smelt winter salvage.

2743
2744 Entrainment effects on longfin smelt larvae could be higher than those for adults due to their
2745 predominant surface orientation (Hieb and Baxter 1993, Bennett et al. 2002) and protracted
2746 larval (i.e., weak swimming) period: almost 90 days are required to reach 20 mm FL (fork
2747 length) (J. Hobbs, UCD, personal communication 2008). However, larval entrainment remains
2748 undocumented because larvae are not identified in salvage until they are ≥ 20 mm in length
2749 (Kimmerer 2008). Similar to adults, entrainment of larvae is presumed highest during periods of
2750 low winter–spring outflows when X2 is near or within the Delta because more spawning occurs
2751 above and within the influence of the export pumps, larval downstream transport is reduced and
2752 exports comprise a substantial fraction of inflow and can draw pelagic larvae into the pumps.
2753 CDFG (2009b) assessed loss of larvae to entrainment in south Delta export pumps by temporally
2754 and geographically scaling particle tracking model results for surface oriented particles to
2755 emulate the timing, distribution and behavior of longfin smelt larvae in the Delta. Annual
2756 percent entrainment was then estimated for 3 relatively low outflow years (1992, 2002, 2008),
2757 when spawning was assumed to be predominantly within the Delta. They found that with
2758 strongly negative OMR flows as occurred in 2002, annual combined SWP/CVP particle
2759 entrainment reached almost 15% of the modeled population; under higher outflow conditions and
2760 much reduced exports in 2008, combined particle entrainment dropped to 3.7%. These particle
2761 tracking results suggest that during some years a substantial fraction of longfin smelt larvae may
2762 be entrained, but entrainment of larvae is not likely to exert much of a negative effect on longfin
2763 smelt recruitment during years with modest to high outflows.

2764
2765 Otolith studies by Bennett et al. (2008) suggest that winter-spawned delta smelt have recently
2766 contributed poorly to the adult population as compared to spring-spawned fish, which benefited
2767 from spring export reductions associated with the VAMP. Similarly, early spawning longfin

2768 smelt may be losing higher numbers of larvae to higher winter exports in the early to mid-2000s
2769 and in general because of less restrictive December and January flow criteria that allow export to
2770 inflow (E/I) ratios as high as 65%, whereas E/I ratios for February through June cannot exceed
2771 35% except in critically dry years, and then must remain < 45%. Existing data provide some
2772 support for this hypothesis. Accounting for recent high winter and spring longfin smelt salvage
2773 (i.e., 2001–2004) and estimated loss of larvae (CDFG 2009b) did in some cases greatly reduce
2774 expected recruitment based on post-clam outflow-abundance relationships (Figure 27).
2775 However, in 2002, the high winter adult loss, the almost 15% larval loss from estimated particle
2776 entrainment, and estimated recent peak in spring–summer juvenile loss (CDFG 2009b) were not
2777 sufficient to obviously reduce 2002 juvenile abundance from that expected based on the post-
2778 clam outflow abundance relationship (Figure 27 a–c). This suggests that flow-abundance
2779 relationships already incorporate entrainment effects among the negative factors limiting
2780 recruitment during low outflow years, and that these effects were not substantially higher in 2001
2781 and 2002, but may have been higher in 2003–2005.

2782
2783 Winter habitat for adult and juvenile longfin smelt is broad and non-restrictive, and probably did
2784 not change during the POD years except at the upstream boundary with shifting X2. For larvae,
2785 habitat varies with winter–spring outflow and X2 (Kimmerer et al. 2009), and was probably
2786 reduced compared to the late 1990s because of a general shift in winter X2 location upstream
2787 during the POD years (CDFG 2009b). Adult and juvenile longfin smelt occupy the entire range
2788 of salinities and temperatures available during winter, though ripe adults are believed to seek
2789 freshwater for spawning (Baxter 1999, Moyle 2002, Rosenfield and Baxter 2007). Larvae are
2790 rare at salinities >18 ppt and salinities between 0.1 and 18 ppt have been hypothesized to
2791 represent nursery habitat (Hieb and Baxter 1993, Kimmerer et al. 2009). Estimated nursery
2792 habitat based on salinity varied significantly positively with outflow (negatively with X2),
2793 though the slope of the outflow-habitat relationship was much less than that of the outflow-
2794 abundance relationship (Hieb and Baxter 1993). Similarly, the comparison of slopes of the X2-
2795 habitat and X2-abundance relationships showed a similarly steeper slope for the latter (Kimmerer
2796 et al. 2009). These results suggest that other factors besides salinity that are associated with
2797 longfin smelt larval survival were positively influenced by outflow. The downstream transport
2798 and distribution of larvae within the estuary varies positively with outflow (Baxter 1999, Dege
2799 and Brown 2004). Specifically, larvae disperse farther downstream when X2 is farther
2800 downstream and young juveniles historically remained in the same regions even as X2 recedes
2801 upstream. Thus, in relatively high outflow years, the longfin smelt distribution immediately
2802 begins to diverge from those of other POD species; this is not the case in relatively low outflow
2803 years when most longfin smelt are initially distributed upstream of Carquinez Strait. Turbidity
2804 has recently been linked as a significant component of a longfin smelt larva habitat (Kimmerer et
2805 al. 2009). Increased turbidity, associated with outflow events, may provide a competitive
2806 feeding advantage to longfin smelt larvae or may reduce predation (Stevens and Miller 1983,
2807 Chigbu 2000). Recent histology revealed that longfin smelt larvae and juveniles possess a large,
2808 well developed olfactory system (Scott Foott, USFWS, personal communication 2006, Foot and
2809 Stone 2008), which can be used for food acquisition in a turbid or dark environment; longfin
2810 smelt are known to feed effectively after dark (Dryfoos 1965, Hobbs et al. 2006). We have no
2811 evidence of a change in winter turbidity levels during the POD years.

2812

2813 Contaminant effects have not been evaluated for longfin smelt eggs or winter larvae, but such
2814 effects seem possible given the general contaminant sensitivity of fish eggs and larvae and the
2815 presence of both life stages during winter when both waterborne and sediment borne toxicants
2816 can be high (see Anderson et al. 2007). In particular, life stages present during first-flush events
2817 may be at greater risk. First flush events carried increased pesticide concentrations in the form
2818 of suspended-sediment associated pesticides (Bergamaschi et al. 2001). These authors reported
2819 that such sediments settled out and re-suspend, likely increasing residence time in the Delta. The
2820 proclivity of longfin smelt for spawning on sand and a 20+ day egg incubation period in winter
2821 Delta temperatures (see CDFG 2009a, Tigan and Lindberg, UCD, unpublished data, CDFG
2822 unpublished data) could place them in protracted proximity to a suite of pesticides (see
2823 Bergamaschi et al. 2001 and reference therein). Only limited evidence has been found of recent
2824 winter-time water toxicity to a *H. azteca* (Werner et al. 2008a, b). Nonetheless, longfin smelt
2825 eggs and surface oriented early-stage larvae would be particularly vulnerable to pulse-flow
2826 transported contaminants. Increasing ammonia/ammonium levels from riverine discharges (see
2827 Jassby 2008) represent an emerging issue both in terms of direct toxicity – delta smelt
2828 larvae/juveniles have proven sensitive to ammonia but ambient concentrations are below those
2829 causing acute mortality (Werner et al. 2008a, b) – and indirectly through changes in
2830 phytoplankton community composition (Kimmerer 2005, Lehman et al. 2005). We have no
2831 information on ammonia toxicity for any life stage of longfin smelt and toxicity testing requires
2832 the ability to culture the species. Laboratory spawning and rearing of longfin smelt commenced
2833 during winter 2009 and continued in 2010 (Rettinghouse 2009, 2010).

2834
2835 Food availability for larvae has not been fully evaluated for winter. Limited diet analysis
2836 revealed that longfin smelt larvae feed predominantly on calanoid copepods in general and *E.*
2837 *affinis* in particular (S. Slater, CDFG, unpublished data). Trends in winter abundance indices of
2838 calanoid copepods do not show declines during the POD years (A. Hennessy, CDFG, personal
2839 communication). Recent otolith analysis shows that longfin smelt larvae grow relatively slowly
2840 during winter and early spring, attaining 20 mm in length only after almost 90 days of growth
2841 post hatch (J. Hobbs, UCD, personal communication). Such slow growth could be adaptive for
2842 modest food resources in winter, similar to that observed for delta smelt (cf., Bennett et al.
2843 2008). Age-1 and age-2 longfin smelt most likely feed on mysids when and where available, and
2844 rely on copepods and amphipods otherwise (Feyrer et al. 2003, S. Slater, CDFG, unpublished
2845 data).

2846
2847 *Spring:* Like delta smelt, longfin smelt hatched in spring probably benefited from reductions in
2848 spring exports associated with VAMP (i.e., reduced top-down effects) since 2000. Based on
2849 particle tracking modeling, longfin smelt also likely benefited from additional export restrictions
2850 in place in 2008 to protect delta smelt (CDFG 2009b); similar benefits were achieved in 2009
2851 and 2010. Low winter and spring outflows in 2001–2002, and modest outflow combined with
2852 strongly negative winter and spring OMR flows in 2003 and 2004 likely kept many young
2853 longfin smelt in the Delta resulting in increased juvenile entrainment (CDFG 2009b). Higher
2854 winter and spring flow coupled with less negative OMR flows in 2005 and 2006 (CDFG 2009b)
2855 likely resulted in the transport of many larvae and juveniles to Suisun Bay and farther
2856 downstream (http://www.delta.dfg.ca.gov/data/20mm/CPUE_Map.asp).

2857

2858 In spring, age-1 longfin smelt are broadly dispersed within the estuary (Baxter 1999, Rosenfield
2859 and Baxter 2007) and similar to the winter, not restricted by habitat. Immediately after hatching
2860 buoyant larvae are dispersed downstream such that their mean location is approximately that of
2861 X2 or just upstream (Dege and Brown 2004). Small juveniles then disperse downstream into
2862 more saline habitats (Dege and Brown 2004, Kimmerer et al. 2009). Kimmerer et al. (2009)
2863 found longfin smelt spring habitat based on 20-mm Survey data (larva and small juveniles)
2864 peaked at about 2 ppt and declined rapidly to about 15 ppt, similar to that observed for larvae in
2865 winter (see previous section). The importance of low salinity habitat to the apparent survival of
2866 larvae was examined by Hobbs et al. (2010) who found that larvae surviving to recruit to older
2867 ages (i.e., summer and fall juveniles and fall adults) had primarily reared in low salinity waters
2868 (0.4–3 ppt) as compared to fresh (≤ 0.3 ppt) or more saline water (≥ 4.0 ppt). However, some
2869 fish recruited from all habitats. As observed for winter habitat, turbidity may also be an
2870 important constituent of spring habitat. Kimmerer et al. (2009) found the combination of salinity
2871 and Secchi depth substantially improved the model fits as compared to salinity alone or salinity
2872 and water depth, suggesting that both salinity and reduced water clarity were important
2873 constituents of habitat for young longfin smelt. Kimmerer et al. (2009) also found that longfin
2874 smelt habitat size varied inversely with X2 location. However, X2 location was not consistently
2875 high during spring in POD years (see CDFG 2009b), so a decrease in habitat did not appear well
2876 related to the POD longfin smelt decline.

2877
2878 Age-0 longfin smelt take advantage of seasonally increasing copepod numbers in spring, feeding
2879 particularly strongly on *E. affinis* and switching to mysids as soon as they are capable (CDFG
2880 2009b, S. Slater, CDFG, unpublished data). The dietary importance of *E. affinis* and mysids
2881 suggests that declines in either might affect longfin smelt survival and recruitment. Spring
2882 densities of *E. affinis* have not declined substantially during the POD years, and even spiked
2883 upward in 2006 and 2008 (Hennessy 2008, 2010). Conversely, spring mysid densities did
2884 decline after 2000 and were substantially lower in odd than even years, culminating in extremely
2885 low numbers in 2007 (Hennessy 2008). Although there was a recovery in 2008, spring 2009
2886 mysid numbers were lower than those of 2007 (Hennessy 2010). These mysid declines probably
2887 reduced feeding opportunities for age-1 longfin smelt in the upper estuary, but we have not
2888 evaluated recent survival. Although no histological evidence of food limitation (or contaminant
2889 effects) was found in young longfin smelt collected in spring 2006 (Foott et al. 2006) or 2007
2890 (Foott and Stone 2008), such young affected fish may not survive long enough to be represented
2891 in collections. No viruses were detected in either year. There was also a low incidence of
2892 parasites, inflammation, or other evidence of cell damage (Foott et al. 2006, Foott and Stone
2893 2008). These data suggest that reduced food in spring may have affected older longfin smelt but
2894 not age-0 longfin smelt, and that disease and parasites were not important factors in spring.

2895
2896 *Summer:* By mid-summer entrainment in south Delta export pumps is no longer an issue for
2897 longfin smelt, because like delta smelt, most of the population moves downstream of the zone
2898 affected by water exports. Increasing Delta water temperatures ($>22^{\circ}\text{C}$) are believed to limit
2899 longfin smelt distribution and cue emigration (CDFG 2009a). More highly mobile age-0 longfin
2900 smelt disperse farther downstream and their distribution further diverges from other POD fishes,
2901 now ranging primarily from eastern Suisun Bay to marine waters of central San Francisco Bay
2902 (Baxter 1999). By summer, age-1 longfin smelt have left the Delta and begin a slow migration
2903 toward central San Francisco Bay (Baxter 1999).

2904
2905 As age-0 longfin smelt grow through summer their diet rapidly broadens to include amphipods
2906 and even more mysids (S. Slater, CDFG, unpublished data). In 2005 and 2006, the transition
2907 away from copepods to mysids (and later amphipods) occurred in summer (S. Slater, CDFG,
2908 unpublished data, CDFG 2009b) as *E. affinis* seasonally declined and *P. forbesi* increased
2909 (Hennessy 2008). Summer mysid abundance declined through the POD years, but remained at
2910 $>10\text{ m}^{-3}$ until 2007 (Hennessy 2008). Like spring mysid abundance, summer abundance
2911 rebounded in 2008 and declined again in 2009 (Hennessy 2010). Since the early 1990s the upper
2912 estuary mysid community has been dominated by *H. longirostris*, which is smaller and slimmer
2913 than *Neomysis mercedis*, and thus may not provide similar nutrition. This trend is reflected in
2914 the long term decline in the mean size of mysid in the estuary (Winder and Jassby 2010). In
2915 summers since 1987, *C. amurensis* grazing has reduced calanoid copepod and mysid availability,
2916 and has probably affected age-0 longfin smelt survival to fall in a manner similar to that
2917 observed for young striped bass (Kimmerer 2002 and reference therein). During the period from
2918 1995 through 2004 (including early POD years) there has been a further decline of calanoid
2919 copepods, particularly *P. forbesi*, in Suisun Bay and the western Delta (Baxter et al. 2008),
2920 which represents a sizable portion of the longfin smelt summer distribution. Regionally
2921 diminishing food resources may also be responsible for reduced fall recruitment in 2003–2005
2922 (Sommer et al. 2007) and for reduced post drought survival of longfin smelt from their first
2923 through second falls (Rosenfield and Baxter 2007). Feeding conditions may improve during
2924 high outflow years. *E. affinis* numbers increased sharply and mysid numbers remained stable in
2925 2006 (Hennessy 2008), the only recent high outflow year. Age-0 longfin smelt collected from
2926 San Pablo and Suisun bays in summer 2006 exhibited 13% (n=107) incidence of hepatocyte
2927 vacuoles that contained either fat or glycogen reserves, which is uncommon for rapidly growing
2928 fishes (Foott et al. 2006). Longfin smelt abundance increased substantially during fall 2006,
2929 though only to the low range of their recent outflow abundance relationship (Figure 27, Sommer
2930 et al. 2007). Unfortunately, no data on abundance of longfin smelt food resources are collected
2931 in San Pablo or central San Francisco Bay where most longfin smelt appear to rear in recent
2932 years.

2933
2934 Changes in food availability may also be responsible for historical and recent changes in longfin
2935 smelt distribution as well as abundance. Longfin smelt exhibited a historical shift to higher
2936 salinity soon after the introduction of *C. amurensis* (Fish et al. 2009). This shift, similar to that
2937 of northern anchovy (Kimmerer 2006), was also likely a response to reduced pelagic feeding
2938 opportunities. More recently, while investigating the Bay Study midwater and otter trawl catch
2939 relationships, we observed a general shift in where longfin smelt are captured in the water
2940 column. The ratio of catch in the water column to catch at the bottom declined sharply during
2941 the POD years and has remained low, suggesting a shift in habitat use toward the bottom (Figure
2942 28). Through the entire period of record, summer–fall longfin smelt (mostly age 0) catches in
2943 the midwater trawl generally exceeded those in the otter trawl in Suisun Bay and the west Delta,
2944 whereas from San Pablo Bay downstream the reverse was true (Figure 29a). During the POD
2945 years, coincident with the sharp drop in the midwater to otter trawl catch ratio (Figure 28),
2946 relative otter trawl catches by embayment shifted downstream and the greatest proportion
2947 occurred in central San Francisco Bay (Figure 29b). Thus both historical and recent downstream
2948 shifts in habitat use have occurred, in addition to the recent shift toward the bottom indicated by
2949 the trawl ratio decline. These shifts downstream and toward the bottom further suggest that the

2950 pelagic feeding environment of the upper estuary has declined and that the longfin smelt
2951 response occurred in stages. Also, such shifts undoubtedly affected longfin smelt abundance as
2952 indexed by midwater trawls, and probably contributed in part to the declines observed in
2953 midwater trawl abundance indices.

2954
2955 No direct link has been made between contaminants and longfin smelt. In 2006, invertebrate
2956 toxicity (*H. azteca*) was detected from water samples taken within the range of longfin smelt, in
2957 particular in eastern San Pablo Bay (Werner et al. 2008a), however, histopathological
2958 examination of longfin smelt collected from the same region before and after the water collection
2959 did not reveal evidence of contact with a toxic substance (Foott et al. 2006). Few longfin smelt
2960 were collected in summer 2007 to assess possible contaminant or parasite effects (Foott and
2961 Stone 2008).

2962
2963 *Fall* : Age-0 longfin smelt seek deep water and may be geographically limited by high water
2964 temperatures (>22°C) and possibly food resources in fall. Presumably, seasonally high water
2965 temperatures limit use of habitat in south San Francisco Bay and the shallows of San Pablo Bay
2966 (Rosenfield and Baxter 2007). Upper estuary drivers (e.g., entrainment effects) probably have
2967 little effect on age-1 fish during fall, because they appear to emigrate to central San Francisco
2968 Bay and some leave the estuary and enter the near coastal ocean (Rosenfield and Baxter 2007).
2969 Observational data suggests high Delta water temperatures may also limit their distribution in
2970 early fall, but both age-0 and age-1 longfin smelt reoccupy the Delta as water temperatures drop
2971 in late fall (Baxter 1999, Rosenfield and Baxter 2007). By late fall, age-1 longfin smelt begin to
2972 mature and their movement into the Delta represents the start of their spawning migration
2973 (Baxter 1999; Rosenfield and Baxter 2007).

2974
2975 The fall distribution of longfin smelt has changed in the long and short terms (see discussion in
2976 the Summer section above). Fall copepod and mysid numbers did not exhibit a distinct,
2977 consistent decline after 2000, but since 2005 mysid numbers, primarily *H. longirostris* (formerly
2978 *Acanthomysis bowmani*), declined sharply and remained low through 2009 (Hennessy 2010).
2979 Low fall mysid abundance in the upper estuary undoubtedly had an effect on longfin smelt
2980 distribution, if not abundance.

2981
2982 Contaminant effects on fall fish remain unresolved. Hepatocyte vacuoles were observed in 76%
2983 (16 of 21) of liver sections from fish collected between September and November 2007, as
2984 compared to 25% of 77 longfin smelt collected between July and October of 2006 (Foott and
2985 Stone 2008). In 2006, these vacuoles were attributed to storage of lipoproteins in maturing fish;
2986 however, the fish collected in 2007 were immature age-0 fish, suggesting that other factors may
2987 be at least partially responsible for the vacuoles. Alternately, the vacuoles may have represented
2988 a biomarker of contaminant exposure, but additional biomarkers often present after contaminant
2989 exposure were not present (Foott and Stone 2008). Hinton and Lauren (1990) discuss factors
2990 associated with changes in fat storage in fish hepatocytes, including exposure to toxicants,
2991 nutritional state, and vitellogenesis in maturing females. Foott and Stone (2008) conclude that
2992 before hepatocyte vacuolation can be used as a biomarker for contaminant exposure it will be
2993 necessary to distinguish it from normal developmental processes by examining the seasonal
2994 changes in healthy longfin smelt.

2995

2996 *Summary:* The conclusion of Thomson et al. (2010), that no single or combination of factors
2997 stands out as responsible for the POD decline of longfin smelt, seems to be an appropriate
2998 summation for this current review. Though these authors detected a step decline in the FMWT
2999 abundance of longfin smelt and the other POD fishes, they were not able to attribute it to any of
3000 the suit of variables assessed, including winter and spring exports, spring X2 location, spring and
3001 summer calanoid copepod biomass, and summer mysid biomass. This was not unexpected given
3002 the complex interactions among environmental and biological drivers. In addition to identifying
3003 both group and individual species step declines during the POD years, their results further
3004 supported water clarity and spring X2 (Mar–May, winter not tested) as important correlates to
3005 longfin smelt abundance. The longfin smelt individual species step decline identified by
3006 Thomson et al. (2010) occurred in 2004 after the multi-species step decline in 2002 and
3007 coincident with the 2004–2005 downward deviation in abundance indices from the outflow
3008 abundance regression line (Figure 27a). Currently, longfin smelt abundance trends are best
3009 explained by changes in winter–spring outflow (X2 location) and possibly stock-recruitment
3010 effects (to be further examined). The abundance declines reflected in the deviations observed in
3011 the outflow abundance relationships likely resulted, at least in part, from recent downstream and
3012 vertical shifts in distribution. These distribution shifts may be related to changes in food
3013 resources; however, these linkages are not yet firmly established and need to be investigated
3014 further.

3015

3016 *Striped bass*

3017

3018 The San Francisco Estuary striped bass population has been monitored and researched for many
3019 decades, so some of the drivers influencing its long-term abundance index trend are fairly well-
3020 understood (Figure 6). The age-0 striped bass abundance index has declined steadily since the
3021 latter 1960s (Figure 2), but adult abundance, as indexed by a long-term Petersen mark-recapture
3022 survey (Figure 9) has not. The lack of change in adult abundance is supported by San Francisco
3023 Bay and Delta Commercial Passenger Fishing Vessel data (Figure 30). This is partly attributed
3024 to active management actions (Kohlhorst 1999). The long-term decline in juvenile abundance
3025 was originally attributed largely to entrainment in the SWP and CVP water diversions (Stevens
3026 et al. 1985). River flows and south delta exports historically explained much of the variation in
3027 striped bass year class strength, although the strongest relationships occurred during the summer
3028 (Stevens et al. 1985). Survival from egg to 38 mm larvae in early summer appeared unchanged
3029 through the mid-1990s (Kimmerer et al. 2000). However, recent research suggests that there was
3030 a step-decline around 1977 due to adult mortality and subsequently reduced egg supply
3031 (Kimmerer et al. 2000, 2001). There was another step-change around 1987 coinciding with the
3032 *Corbula* invasion that decoupled age-0 production from spring X2 (Sommer et al. 2007). The
3033 most recent step decline in the early 2000s remains unexplained (Sommer et al. 2007, Thomson
3034 et al. 2010)

3035

3036 Young striped bass have a strong predator-prey association with mysid shrimp (Stevens 1966,
3037 Feyrer et al. 2003). The carrying capacity for age-0 through age-3 striped bass has shown a
3038 long-term decline that is correlated with declining mysid densities (Kimmerer et al. 2000,
3039 Winder and Jassby 2010). Thus, it is likely that the much lower abundance of mysids in the
3040 post-*Corbula* period has strongly and negatively affected juvenile striped bass production. The

3041 adult striped bass population increased in the 1990s and was at about a 30-year peak in the year
3042 2000. This was likely due to a combination of improved survival resulting from successive wet
3043 years in the mid to late 1990s and planting of millions of juvenile striped bass into the estuary
3044 through 2000 (Kohlhorst 1999). Restored populations of striped bass on the east coast have
3045 caused large reductions in populations of their prey and in high incidence of disease in striped
3046 bass (Hartman 2003, Uphoff 2003)

3047
3048 The reasons for the continued decline of the age-0 striped bass abundance index to record lows
3049 during the POD years, despite an increase in the adult abundance index and by extension, egg
3050 supply, is unknown (Thomson et al. 2010). Striped bass appear to show more signs of
3051 contaminant-related health problems than the other POD species (details below), but we do not
3052 know whether this reflects a long-term chronic problem or a recent change. Abiotic habitat
3053 suitability, calculated as a function of Secchi depth (clarity) and specific conductance, for young
3054 striped bass has declined during fall like it has for delta smelt (Feyrer et al. 2007) and the
3055 entrainment of striped bass increased during the early POD years (Figure 18). Thus, it is
3056 possible that direct (entrainment) and indirect (habitat suitability, particularly salinity) effects of
3057 water diversions have exacerbated the longer-term stresses of reduced prey availability and
3058 contaminant effects. Data from two different long-term monitoring studies indicate that age-0
3059 striped bass may be shifting in distribution from channel stations to shoal stations, possibly
3060 contributing to the discrepancy seen in the age-0 and adult abundance indices. The percent of
3061 shoal-caught striped bass is increasing through time relative to channel-caught fish in the San
3062 Francisco Bay Study (Figure 31, T. Sommer, DWR, unpublished data). Likewise, catches of
3063 age-0 and age-1+ striped bass in Suisun Marsh have increased concurrent with decreased catches
3064 in Suisun Bay, which may reflect improved prey availability in the marsh or reduced habitat
3065 quality in Suisun Bay (Schroeter 2008). However, the modest shift in striped bass towards shoal
3066 habitat does not appear to be sufficient to fully explain the extreme decline in age-0 striped bass
3067 abundance (Sommer et al. submitted).

3068
3069 *Spring:* In spring, adult striped bass migrate into the Sacramento and San Joaquin rivers where
3070 they broadcast spawn in currents that will suspend eggs and larvae in the water column. Larvae
3071 require roughly 2 weeks of development before they can maintain their position in the water
3072 column (Moyle 2002). In both rivers, migrating adults and suspended eggs and larvae are
3073 subject to discharges carrying potentially toxic contaminants. Agricultural return water can
3074 transport pesticides and other contaminants (Saiki et al. 1992, Bennett et al. 1995, Weston and
3075 Lydy 2010). Municipal discharges may carry a variety of contaminants including potentially
3076 estrogenic compounds (Huang and Sedlak 2001, Weston and Lydy 2010). Springtime
3077 contaminant effects influence egg and larval survival via two pathways: (1) maternally
3078 accumulated contaminant(s) passed on to eggs and larvae that negatively affect viability and
3079 development; and (2) direct mortality effects of contaminants in the environment on larvae
3080 (Ostrach et al. 2008). Maternal transfer of contaminants was assessed by comparing larvae
3081 produced by domestic (controls) and wild striped bass spawned in the laboratory in 2006 and
3082 2007. Developmental studies showed that wild larvae grew slower than control larvae. Wild
3083 larvae also developed abnormally with some processes accelerated (e.g., development of caudal
3084 fin, and notochord) and other processes retarded (e.g., development of brain and liver from 3 to 5
3085 days posthatch) (Ostrach et al. 2008). Chemical analyses of unfertilized eggs from 21 striped
3086 bass collected from the Sacramento River found biologically significant lipophilic compounds

3087 and known endocrine disruptors such as polychlorinated biphenyls (PCBs), PBDEs, and current-
3088 use and legacy pesticides. These analyses, coupled with the morphometric and histopathological
3089 results from the brain and liver indicate that contaminants are maternally transferred and likely
3090 disrupt early development of river larvae (Ostrach et al. 2008). Striped bass eggs and larvae are
3091 likely exposed to the same suite of contaminants as delta smelt (Kuivila and Moon 2004),
3092 possibly in higher concentrations because striped bass spawn closer to locations of upper river
3093 discharges (e.g., Colusa Basin Drain). During the late 1980s and early 1990s, striped bass larvae
3094 captured in the Sacramento River exhibited liver lesions sufficient to cause mortality (Bennett et
3095 al. 1995).

3096
3097 Rapid loss of large fecund females from the spawning stock could reduce total fecundity, and in
3098 turn cause a decline in juvenile recruitment (Stevens et al. 1985, Kimmerer et al. 2000). Such a
3099 decline occurred between 1976 and 1977 (Stevens et al. 1985, Kimmerer et al. 2000). Bennett
3100 and Howard (1997, 1999) hypothesized that a decline in striped bass fecundity since the 1970s
3101 resulted from warming ocean conditions leading to an improved coastal feeding environment,
3102 which in turn resulted in increased fishing mortality along the coast and increased straying to
3103 other river systems. This scenario fits data from the late 1970s through late 1990s, but ocean
3104 conditions shifted to a cool regime in 1999 and although the adult population size decreased after
3105 2000 it remained at or above early 1990s levels (Figure 9).

3106
3107 A dramatic decrease in the female:male sex ratio of adult striped bass (≥ 3 years old) is apparent
3108 in long-term Petersen mark-recapture data from the estuary (Figure 10; CDFG, unpublished
3109 data). Skewed sex ratios during spawning can result from a difference in the maturity schedule
3110 for male and female striped bass. On average, males mature two years younger than females
3111 and, thus appear on spawning grounds in a much higher abundance. This pattern holds true for
3112 the delta population (Figure 32) and is characteristic of stocks outside the estuary as well (Trent
3113 and Hassler 1968, Hoff et al. 1988). However, the recent change in the ratio indicates that
3114 something has changed. There has been a steady decline in the number of older females (Figure
3115 33) which is troubling given that a wide distribution of age classes and a relatively high
3116 proportion of older females are generally good indications of a healthy stock. Since the early
3117 1980s, the number of females ≥ 7 collected in adult tagging efforts is often zero or one. The
3118 same data indicates that there has been a reduction in apparent growth of males and females,
3119 most noticeable in the upper age classes (Figure 34), an increase in natural and total annual
3120 mortality, and a decrease in harvest rate (Figure 35). The loss of these older age classes coupled
3121 with the reduction in apparent growth can lead to smaller, less fecund females, and consequently
3122 lower juvenile production.

3123
3124 Many factors could be responsible for the change in sex ratios. Hypotheses include change in
3125 sampling methodology, increased mortality of females, changes in distribution, or physiological
3126 effects. Some evidence for each is described below.

3127
3128 A change in location of Sacramento River fyke trap sampling of spawning adults in 1990 may
3129 have introduced a gear bias that could account for some, but not all, of the sex ratio change. Bias
3130 also seems unlikely because a similar shift was also noted in gill net captures from the Delta.
3131 Apparent growth rates of females have declined (T. Sommer, DWR, unpublished data) so

3132 females may be increasing the number of years between spawning migrations, reducing their
3133 capture rates in the fyke traps or gill netting surveys in the Delta.

3134
3135 Changes in natural or fishing mortality could also affect sex ratios. Anglers may exert selective
3136 fishing pressure on females but since the sex ratio of recaptured fish closely approximates the
3137 sex ratio of tagged fish, it is unlikely this is a major factor. The female:male ratio of recaptured
3138 fish is higher than the female:male ratio of tagged fish in all but four of the 34-year period of
3139 record, so an angler effect should not be ruled out entirely. An alternative is that there have been
3140 changes in natural mortality. For example, increased predation of female striped bass by
3141 pinnipeds could be important. No time series for the delta exists, making this difficult to
3142 examine, but it is generally recognized that there has been an increase in pinniped numbers since
3143 the passage of the Marine Mammal Protection Act in 1972. It is estimated that the California sea
3144 lion and Pacific harbor seal populations have been increasing at an annual rate of 5% (NMFS
3145 1997). For example, NMFS (1997) report summer counts of 4900 harbor seals in San Francisco
3146 Bay utilizing 118 different haul out locations.

3147
3148 Another potential mechanism for changes in the sex ratio during the spawning season is that
3149 adult striped bass distribution may have shifted. One possibility is that females may leave the
3150 estuary for the nearshore ocean and not return to the estuary for annual spawning events. Such
3151 departure from the estuary has been observed in the Chesapeake Bay striped bass population,
3152 where as much as half the adult population has been observed in the ocean, far from spawning
3153 habitats (Secor et al. 2001, Secor 2008). It is unclear if this mechanism would result in skewed
3154 sex ratios unless females are more likely to skip spawning. Secor (2008) reports that skipping
3155 rates for Chesapeake Bay striped bass were less than 20%, which may not be able to account for
3156 the currently extreme low percentages of females detected during migration in the San Francisco
3157 Estuary. A related factor is that it is possible that maturation schedules have changed for female
3158 striped bass, which in turn could alter their seasonal distribution. If females are maturing more
3159 slowly, they may remain in downstream areas for longer periods. Males are already known to
3160 migrate upstream at a younger age, so a delay in female maturation would have a strong effect on
3161 sex ratios in young adults (e.g. age-3 and age-4).

3162
3163 A fourth explanation for the shift in sex ratios is that there may be environmental effects that
3164 reduce the number of females in the population. Temperature has been strongly implicated as a
3165 co-factor in sex determination in various members of the family Moronidae (Ospina-Alvarez and
3166 Piferrer 2008; Vandeputte et al. 2007), although temperature effects on sex ratios have not been
3167 documented for striped bass. Chemical inputs can also be a factor. Female adult fish
3168 concentrate lipophilic contaminants in their tissues, many of which can be estrogenic. These
3169 chemicals can adversely affect larval development, causing organ deformities, retarded growth,
3170 reproductive system abnormalities, and hatching difficulties (Ostrach et al. 2008). Based on
3171 research on other species, abnormalities include the increased presence of intersex fish,
3172 feminized males, and an increase in female fish. As evidence that this could apply to Pacific
3173 striped bass, intersex fish have also been observed in high numbers in striped bass populations in
3174 Oregon. However, the high incidence of hermaphroditism may also be a result of insufficient
3175 genetic diversity (Waldman et al. 1998).

3176

3177 Striped bass consume primarily copepods and cladocerans at first feeding (Heubach et al. 1963,
3178 Foss and Miller 2004), and prey density has been positively correlated with larval growth rate
3179 (Heubach et al. 1963, Foss and Miller 2004), which is believed to be inversely related to
3180 mortality. Although a long-term reduction in calanoid copepods has occurred (Orsi and Mecum
3181 1986, Kimmerer et al. 1994, Orsi and Mecum 1996, Winder and Jassby 2010), we have not
3182 found evidence of a recent decline in spring coincident with the striped bass decline. Also,
3183 cladoceran abundance increased steadily during spring of the POD years (Hennessy and Hieb
3184 2007). However, broad annual and seasonal indices of abundance may mask important short
3185 term patterns. For example, in the low salinity zone, early first feeding larvae probably rely on
3186 *E. affinis*, whereas those hatching later will encounter primarily *P. forbesi*. During the spring-
3187 time transition period from *E. affinis* to *P. forbesi* dominance, it is possible that finding sufficient
3188 numbers of either species might be problematic.

3189
3190 *Summer:* Summer is a generally a period of rapid growth for striped bass, but food resources
3191 may not be uniformly adequate throughout their range to support this growth. In particular,
3192 calanoid copepod numbers have been reduced recently in the western Delta and Suisun Bay
3193 (Winder and Jassby 2010). However, striped bass grow rapidly, even with the relatively low
3194 copepod densities found in the upper San Francisco Estuary (Foss and Miller 2004). Success
3195 from first feeding to 25 mm may have influenced recruitment since 2000, but there was no
3196 suggestion of a bottle neck during this period in the past (Kimmerer et al. 2000). By 25 mm in
3197 length, age-0 bass exploit larger mysid and amphipod prey, appearing to seek out mysids (Feyrer
3198 et al. 2003; Bryant and Arnold 2007). The dominant upper estuary mysid, the introduced *H.*
3199 *longirostris*, has slowly declined in summer since 2000, (Hennessy and Hieb 2007) but densities
3200 remain higher than those of the late 1990s, thus are unlikely to have influenced the striped bass
3201 decline since 2000. Long-term diet and growth have not changed substantially since 2000. Diet
3202 data from a long-term study (1973–2002) of striped bass collected from the pelagic environment
3203 during their first summer of life, revealed that the percentage of age-0 striped bass stomachs
3204 with food did not decline in the last two years, nor did the mean ration size, except perhaps in
3205 2002 for fish <25 mm (Bryant and Arnold 2007). In a short-term shore-based study, a
3206 bioenergetic modeling (BEM) approach was used to evaluate food limitation in age-0 striped
3207 bass. The comparison of field collections with the BEM simulations did not provide evidence of
3208 food limitation and in fact, field-collected striped bass grew larger relative to their BEM growth
3209 predictions (Nobriga 2009). Most recently, feeding incidence calculated from age-0 fish
3210 collected from May–September of 2005 and 2006 gave no indication of food limitation. On
3211 average, 84% of the stomachs examined had at least one prey item. Amphipods (*Gammarus* spp.
3212 and *Corophium* spp.) were detected in large amounts during fall, which could be evidence of a
3213 switch to larger demersal prey items, possibly due to low mysid availability (DFG, unpublished
3214 data). Examination of apparent growth patterns (growth determined from length-frequency
3215 data) for striped bass did not reveal a decline subsequent to 2000 (IEP 2006). However, when
3216 fish condition was compared regionally, striped bass collected from western Suisun Bay weighed
3217 significantly less at the same length compared to those collected in regions farther east (IEP
3218 2006).

3219
3220 No long-term studies exist of disease or parasites in young striped bass of the San Francisco
3221 Estuary. The incidence and intensity of tapeworm larvae (plerocercoid) was examined in fish
3222 collected from 1986–1993 (Arnold and Yue 1997). In this study, up to 79% of the larval and

3223 juvenile striped bass examined annually contained encapsulated plerocercoids (cestodes) in
3224 mesenteries and stomachs, but little immune response was detected. Recent sampling revealed
3225 no viral effects in summer 2006. There was a high incidence and intensity of trematode and
3226 flagellated/ciliated gill and mouth parasites in 2005–2007, with importance of the various
3227 parasites varying from year to year (Ostrach et al. 2009). The direct importance of such
3228 infections is unknown but, at a minimum, such infections represent a chronic stress on the
3229 immune systems of individual fish.

3230

3231 There is no direct evidence for the importance of contaminants in recent summers. High
3232 vitellogenin levels in male fish collected from Suisun Slough and elevated levels in fish from
3233 Honker Bay indicated contact with estrogenic compounds (Ostrach et al. 2005, Ostrach et al.
3234 2009). The use of pyrethroid pesticides doubled in the early 2000s as compared to the early
3235 1990s and the main period of application, July through October (Oros and Werner 2005),
3236 corresponds to the striped bass growing period.

3237

3238 Striped bass summer distribution appeared to shift after the invasion of the overbite clam in
3239 1987. The San Francisco Bay Study deploys two nets at each sampling location: an otter trawl,
3240 towed on the bottom, and a midwater trawl towed obliquely through the water column. The
3241 proportion of total age-0 striped bass caught in the midwater trawl has been highly variable, but
3242 with a downward trend and lower peaks over time (Figure 31). Similarly, the proportion of age-
3243 0 striped bass caught by the otter trawl at shoal stations compared to deeper stations increased
3244 and remained consistently high after 1987, and became less variable after the late 1990s (Figure
3245 31). This suggests that age-0 striped bass spent less time in the water column and moved to
3246 relatively shallow water sooner in the years after the clam arrived. Current analyses cannot
3247 discriminate whether these trends became more extreme in the 2000s. If a similar shift also
3248 occurred in the Delta, age-0 striped bass might become more vulnerable to shoreline predators,
3249 such as largemouth bass.

3250

3251 *Fall:* Fall has historically posed a feeding challenge for striped bass, but it is not clear if
3252 available food resources changed substantially after 2000. Seasonal and decadal declines in
3253 mysids (a highly selected diet item) density led to a broadened feeding niche and an early
3254 initiation of piscivory among age-0 striped bass in fall (Stevens 1966, Feyrer et al. 2003). Fall
3255 mysid numbers in the upper estuary declined steadily after 2000 with the exception of 2005.
3256 However, current mysid abundances are greater than densities in the early 1990s when *N.*
3257 *mercedis* declined to very low densities. The most abundant mysid currently is *H. longirostris*
3258 which invaded the estuary in about 1994 (Hennessy and Hieb 2007). Reduced food availability
3259 has been linked to observed declines in carrying capacity (Kimmerer et al. 2000) but recent
3260 studies (described above) do not indicate that age-0 striped bass are experiencing food limitation
3261 during the fall (Nobriga 2009, DFG unpublished data).

3262

3263 Age-0 striped bass environmental quality during the fall can be effectively described by Secchi
3264 depth and specific conductance (Feyrer et al. 2007). Physical habitat is important during this
3265 time since conditions during their first fall play a role in the density-dependent survival exhibited
3266 by striped bass from age-0 to age-3 (Kimmerer et al. 2000). Long-term trends indicate a decline
3267 in environmental quality from the late 1960s to present. Fall environmental quality in the 2000s
3268 declined from levels in the 1990s, and most of the remaining high quality habitat exists in the

3269 lower Sacramento River upstream of the confluence (Feyrer et al. 2007). If the age-0 striped bass
3270 population is also shifting toward the Delta and to a smaller habitat area there could be increased
3271 intra-specific competition, and predation might increase locally with increased densities and the
3272 overlap with additional piscivores, particularly largemouth bass (Nobriga and Feyrer 2007). In
3273 addition, the upstream shift may increase vulnerability to entrainment at the south delta water
3274 export pumps.

3275
3276 *Winter:* We hypothesize a potential winter bottleneck where only the largest and healthiest
3277 individuals survive. This derives in part from analyses by Kimmerer et al. (2000) who found
3278 little relationship between two indices of age-0 abundance and the number of bass entering the
3279 fishery at age 3. This contrasts with the results for age-1 fish, which are correlated with the
3280 number of age-3 fish. Hence, it appears likely that the problem occurs between age-0 and age-1.
3281 We hypothesize that the bottleneck occurs soon after fall, when age-0 abundance is measured. In
3282 winter, young striped bass revert back to feeding on invertebrate prey (Stevens 1966). In
3283 addition to mysids and amphipods they likely prey upon the recently introduced decapod shrimp,
3284 *Exopalaemon modestus*, as they do later in spring (Nobriga and Feyrer 2007). The increased
3285 numbers of *E. modestus* (Hieb 2007) likely improved food resources for winter striped bass.

3286
3287 Similar to other species, young striped bass showed increased winter entrainment as evidenced
3288 by salvage shortly after 2000 (Figure 18; IEP 2005, Grimaldo et al. 2009). This effect occurs
3289 after age-0 trends are assessed by FMWT. Losses indexed by winter salvage from the early
3290 2000s would only begin to be reflected in adult numbers in 2003 and beyond. However, adult
3291 striped bass estimates do not show a substantial drop in age-3 or age-4 abundance in 2003 or
3292 2004 (Figure 33).

3293

3294 *Threadfin shad*

3295
3296 Prior to the POD investigation there was very little information available on the ecology of
3297 threadfin shad in the Delta. Thus, one element of the POD investigation was to compile and
3298 synthesize the available data from IEP monitoring programs and special studies on threadfin
3299 shad. This effort has recently been completed (Feyrer et al. 2009) and forms the basis for the
3300 threadfin shad species conceptual model presented below (Figure 7).

3301
3302 Threadfin shad successfully invaded and has persisted in the Delta because of suitable
3303 environmental conditions. It is widely distributed but is most commonly encountered and most
3304 abundant in the southeastern Delta, especially the San Joaquin River near and just downstream of
3305 Stockton, where suitable abiotic habitat coincides with high prey abundance (Feyrer et al. 2009).
3306 These regions also have a relatively high density of SAV, which provides important spawning
3307 and larval rearing habitat (Grimaldo et al. 2004). Historic studies conducted in 1963–1964
3308 (Turner 1966) and those more recently (Feyrer 2004, Grimaldo et al. 2004) identified a similar
3309 distribution for threadfin shad. Turner (1966) also found that threadfin shad was relatively
3310 abundant in dead-end sloughs of the northeastern Delta, areas which are not sampled by the
3311 current monitoring programs but provide functionally similar habitat.

3312

3313 Threadfin shad appear to grow relatively fast in the Delta and reach 70–90 mm by the onset of
3314 their first winter. This is generally consistent with fish lengths reported for Lake Powell, Utah
3315 and Arizona, U.S.A. (Blommer and Gustaveson 2002), but substantially larger than that observed
3316 in central Arizona reservoirs (Johnson 1970). Apparent growth rates in the Delta during fall
3317 declined with increasing abundance (Feyrer et al. 2009). This suggests density-dependent effects
3318 may be important and is consistent with previous research indicating that intraspecific
3319 competition for food can be a major factor limiting growth of threadfin shad in reservoirs
3320 (Johnson 1970).

3321
3322 It is important to note that threadfin shad has a limited distribution in the Delta. Of the 100 sites
3323 in the FMWT survey, catches at just seven adjacent southern Delta sites in the San Joaquin River
3324 drive the long-term patterns in the FMWT abundance indices (Feyrer et al. 2009). The overall
3325 pattern in interannual abundance has been variable with periods of high and low abundance and
3326 no apparent long-term trend. The recent period of near-record low indices is not unprecedented
3327 but is especially noteworthy because it has persisted. The decline is also apparent in the 20-mm
3328 Survey, salvage density in all seasons, and commercially harvested biomass (Feyrer et al. 2009).
3329 Finally, the decline is coincident with similar declines for other pelagic fishes (Sommer et al.
3330 2007, Thomson et al. 2010). The persistent low abundance of threadfin shad is also noteworthy
3331 because of the documented ability of threadfin shad to rapidly recover from low abundance
3332 levels, so called population explosions, in part because of their synchronous spawning behavior
3333 (Kimsey et al. 1957, McLean et al. 1982). In addition to lower abundance, threadfin shad have
3334 been captured in fewer trawls suggesting that the recent decline in abundance may be driven by
3335 the FMWT encountering fewer and smaller-sized schools of threadfin shad. There have been
3336 similar periods of smaller-sized catches in the past, especially around the mid-1980s. However,
3337 the persistently low fraction of samples with fish present is unprecedented in the time series
3338 (Feyrer et al. 2009).

3339
3340 Threadfin shad are influenced by many of the same drivers affecting the other POD fish species;
3341 however, several drivers do not seem to be important for threadfin shad. Recent studies suggest
3342 that there are no measurable effects of disease on the population (Baxter et al. 2008). There is
3343 also no evidence that abiotic habitat of threadfin shad – measured as the combination of water
3344 temperature, clarity, and salinity – has declined in recent years (Feyrer et al. 2007). There is
3345 little evidence from the data examined for consistent stock-recruitment or stage-recruitment
3346 effects on the population (Feyrer et al. 2009). However, there does appear to be a complete
3347 disconnect between abundance in summer and fall. There are two drivers that are of particular
3348 concern for threadfin shad during this time period, episodes of low dissolved oxygen (DO) and
3349 blooms of the toxic algae *M. aeruginosa*, both of which occur in the center of threadfin shad
3350 distribution. Other drivers such as predation and low water temperatures are also known to affect
3351 threadfin shad populations in other systems (Parsons and Kimsey 1954, Griffith 1978, Blommer
3352 and Gustaveson 2002, McLean et al. 2006) and may be important at times in the Delta.

3353
3354 A final important note is that focused studies and sampling of threadfin shad are lacking and
3355 limit what can currently be concluded about its ecology in the system. Improved field
3356 observations and controlled laboratory studies designed specifically for threadfin shad, which
3357 can then inform modeling studies, are desperately needed to better understand the drivers that
3358 affect threadfin shad population dynamics in the Delta.

3359
3360 *Summer:* Based on preliminary analyses, Baxter et al. (2008) reported that prior fish abundance
3361 appeared to be important because there was a significant stock-recruitment relationship for
3362 threadfin shad. New analyses suggest that, although there is generally some positive response to
3363 prior abundance, there is little evidence for large scale stock- or stage-recruitment effects on the
3364 population (Feyrer et al. 2009). This is not a surprising result as stock-recruitment relationships
3365 are generally poor for opportunistic-type fishes such as threadfin shad (Winemiller 2005).
3366 However, there does appear to be a complete disconnect between abundance in summer and fall,
3367 suggesting that there are important drivers acting on threadfin shad during the summer-to-fall
3368 transition. We hypothesize that two drivers might be particularly important during this period,
3369 episodes of low DO and blooms of the toxic algae *M. aeruginosa*. These conditions occur in the
3370 center of the threadfin shad distribution in the central and southeastern Delta.
3371
3372 Episodes of low DO concentration commonly occur in the San Joaquin River and have been
3373 known to cause die-offs of threadfin shad. Persistent DO sags in the Stockton Deepwater Ship
3374 Channel portion of the San Joaquin River are well documented (Jassby and Van Nieuwenhuysen
3375 2005). It is likely that smaller and shorter-term events occur in smaller Delta channels, but such
3376 events have not been documented.
3377
3378 In recent years, there have been dense blooms of *M. aeruginosa* geographically centered where
3379 threadfin shad are most abundant in the southern Delta (Lehman et al. 2008a). The blooms also
3380 occur during the critical late summer/early fall when newly spawned fish are recruiting to the
3381 population (Lehman et al. 2008a). The effects of *M. aeruginosa* on threadfin shad could be
3382 indirect by affecting food availability or direct by inhibiting feeding and by physiologically
3383 impairing threadfin shad.
3384
3385 For a variety of herbivorous crustacean zooplankton, *M. aeruginosa* can be toxic, non-nutritious,
3386 or inhibited feeding on co-occurring nutritious food (Fulton and Paerl 1987). Ger et al. (2009)
3387 found that ambient concentrations of a *Microcystis* toxin (microcystin LR) in the Delta were
3388 unlikely to be acutely toxic to two Delta copepod species, *E. affinis* and *P. forbesi*, but both
3389 species were very sensitive to *M. aeruginosa* in their diet (Ger et al. 2010a) and experienced
3390 significant mortality when *M. aeruginosa* was 10% or more of the diet, whether the *Microcystis*
3391 strain produced toxic microcystins or not.
3392
3393 A recently concluded POD study (Teh et al. 2010) on *Microcystis* effects on threadfin shad found
3394 that in laboratory feeding experiments, diets spiked with *M. aeruginosa* collected from the Delta
3395 had detrimental effects on the health of threadfin shad, including decreased growth, malnutrition,
3396 severe liver lesions, and increased ovarian degeneration (atresia) which may impair reproduction.
3397 This study also included health evaluations of 296 sub-adult threadfin shad collected from four
3398 Delta sites (Sherman Island, Brannan Island, Mildred Island, and Stockton) in 2007. Threadfin
3399 shad from the two San Joaquin River sites (Brannan Island and Stockton) were of poorer health
3400 than fish caught at the other two sites, but histopathologic analysis revealed that lesions in fish
3401 from the Stockton site were more likely related to effects of anthropogenic contaminants than to
3402 *Microcystis* toxicity. In contrast, the severe intestinal epithelial cell necrosis and the localization
3403 of microcystins in the liver and *Microcystis* in stomachs and intestines of threadfin shad collected
3404 at Brannan Island strongly indicated *Microcystis* toxicity.

3405
3406 In addition to the above drivers, declines in zooplankton abundance in the Delta (Winder and
3407 Jassby 2010) may have also affected threadfin shad. Feyrer et al. (2009) found evidence for
3408 density-dependent effects in threadfin shad that were consistent with previous research in
3409 reservoirs which showed that intraspecific competition for food could limit growth of threadfin
3410 shad (Johnson 1970). Other studies found that the condition of young threadfin shad was
3411 sensitive to prey abundance (Kashuba and Matthews 1984) and that interactions between food
3412 availability and water temperature affected growth rates and cohort survival of young threadfin
3413 shad (Betsill and Van Den Avyle 1997). It is possible that other biotic and abiotic habitat
3414 attributes are also important for threadfin shad in the Delta, but long-term trends in threadfin
3415 shad summer habitat have not yet been systematically and comprehensively examined.
3416
3417 Long-term trends in abiotic habitat during summer have not been examined. However, there is
3418 no evidence that abiotic habitat – measured as the combination of water temperature, clarity, and
3419 salinity – has declined in recent years in fall (Feyrer et al. 2007). Indirectly, habitat conditions
3420 may affect the abundance or survival of young threadfin shad by controlling the density of
3421 suitable prey organisms, particularly cladocerans.
3422
3423 Threadfin shad is a major component of piscivorous fish diets (Stevens 1966, Nobriga and
3424 Feyrer 2007). However, there is insufficient data to determine if predation mortality
3425 significantly affects the population during summer. Threadfin shad is also the most common fish
3426 collected at the export facilities (Brown et al. 1996). However, there are no significant results to
3427 date for an effect of summer exports on the population.
3428
3429 *Fall:* As mentioned above, Feyrer et al. (2009) noted a disconnect between threadfin shad
3430 abundance in summer and fall which may be related to episodes of low DO, toxic *M. aeruginosa*
3431 blooms, and limited food supply in late summer. An analysis of fall abiotic habitat condition for
3432 threadfin shad – measured as the combination of water temperature, clarity, and salinity – found
3433 no trend in the southeastern part of the Delta where threadfin shad are most commonly found
3434 (Feyrer et al. 2007).
3435
3436 Threadfin shad is a major component of piscivorous fish diets (Stevens 1966; Nobriga and
3437 Feyrer 2007). Striped bass, one of the primary predators on threadfin shad, move into the Delta
3438 from downstream bays and the ocean during the fall. This large influx of predators undoubtedly
3439 increases predation pressure on threadfin shad. The other primary predator, largemouth bass,
3440 resides in the Delta all year. However, as mentioned, the effects of predation on the population
3441 have not been studied. Two additional sources of top-down mortality include water exports and
3442 commercial harvest. Seasonal salvage of threadfin shad has been highest and most variable
3443 during fall. Two recent modeling efforts provide limited support for some effect of exports
3444 (Thomson et al. 2010, Mac Nally et al. 2010). Commercial fishing harvest has also been highest
3445 during the fall months (Feyrer et al. 2009). The effects of the loss of these fish from the
3446 population have not been studied.
3447
3448 There is no direct correlational evidence that food densities affect threadfin shad abundance.
3449 However, apparent growth rates during fall have been negatively related to abundance. As
3450 mentioned above, this relationship suggests density-dependent effects may be important.

3451
3452 *Winter:* The abundance of threadfin shad is not measured during winter, thus it is not possible to
3453 determine the potential importance of individual drivers. Long-term trends in habitat conditions
3454 (salinity and temperature) during winter are presumed to be generally similar to the fall, which
3455 would suggest no major impacts on the population. However, low water temperatures are known
3456 to cause massive die offs of threadfin shad and may be important at times in the Delta because
3457 winter temperatures occasionally decline below the minimum tolerances of threadfin shad (6–
3458 8°C; Turner 1966).

3459
3460 The top-down effects of predation are presumed to be similar to the fall; however, there is
3461 insufficient data to determine if predation mortality significantly affects the abundance of
3462 threadfin shad during winter. There was an increase in the salvage of threadfin shad during
3463 winter prior to the onset of the POD, but it has steadily declined since that time.

3464
3465 *Spring:* As mentioned above, there is little evidence for large-scale stock-recruitment effects
3466 (Feyrer et al. 2009). There are years in which spring abundance appears to respond to fall
3467 abundance, but the data are highly variable.

3468
3469 Long-term trends in abiotic habitat conditions during spring have not been studied.
3470 As already mentioned above, indirect evidence suggests that the availability of food resources
3471 may have an effect on the survival and abundance of young threadfin shad during spring and
3472 summer.

3473
3474 Seasonal salvage of threadfin shad is lowest during spring and exports are not likely to affect the
3475 population during this season. As water temperatures increase during spring so does predation
3476 pressure from other species. The extent to which this affects the population has not been studied.
3477

3478 **The Pelagic Organism Decline: A Historical Perspective**

3479

3480 **Why a historical perspective?**

3481
3482 Since its inception in 2005, the POD investigation has largely focused on drivers that currently
3483 affect resources in the San Francisco Estuary. It is, however, important to understand that current
3484 conditions represent the outcome of a continuum of major changes to the system. Some of these
3485 historical changes are likely as important in understanding the POD as some of the more recent
3486 changes. Carpenter and Turner (2000) state this concept as follows: “ecosystem dynamics are
3487 history dependent because of the coupling of events across a range of cycling times” (or time
3488 scales). Note that we do not present this historical perspective to represent historical conditions
3489 as a baseline or target for management; however, it is instructive to compare the functioning of
3490 the current ecosystem with earlier periods. Such comparisons provide useful insights into some
3491 of the structural and functional ecosystem attributes that have been lost, and insight into the
3492 magnitude of changes that might be necessary to recover them.
3493

3494 Ecological resilience is a key concept for understanding how ecosystems respond to natural and
3495 human disturbances. Ecological resilience is the capacity of an ecosystem “to absorb disturbance
3496 and reorganize while undergoing change so as to retain essentially the same function, structure,
3497 identity, and feedbacks” (Folke et al. 2004). In other words a resilient ecosystem will look and
3498 function the same, before and after a disturbance, given adequate time to recover. A resilient
3499 ecosystem has options to adapt to change – it has “adaptive capacity” (Folke et al 2002).
3500 Understanding changes in ecological resilience of the San Francisco Estuary due to long-term
3501 historical changes or relatively recent changes is central to the emerging understanding of the
3502 POD as an ecological regime shift. An ecological regime shift is an abrupt transition from the
3503 previous state (regime) to a fundamentally different state with altered functions, structure,
3504 identity, and feedbacks. These concepts are explored in more detail in the following sections.
3505

3506 *Change in Ecosystems*

3507
3508 Estuaries are highly dynamic ecosystems that constantly undergo change (Healy et al. 2008).
3509 Before summarizing the historical changes in the San Francisco Estuary, we will briefly discuss
3510 drivers of change in ecosystems, and define some terms used in the remainder of the report.
3511

3512 Ecological change is brought about by interlinked drivers operating across a large range of
3513 temporal and spatial scales and often occurs in cycles (Chapin et al. 2009). Adaptive cycles
3514 consist of a slow growth and conservation phase which, after a significant disturbance, is
3515 followed by rapid collapse and release and eventually by reorganization and renewal. The growth
3516 and conservation phases are characterized by increasingly efficient, but also increasingly rigid
3517 use of available resources which erodes the resilience of the system to disturbance. Persistent
3518 small-scale disturbances (e.g. seasonally variable hydrographs, localized floods, small-scale
3519 fires) prevent optimization of resource use, but can help maintain resilience to larger scale
3520 disturbances.

3521
3522 Exogenous drivers, such as climate and geology, generally affect large regions and remain
3523 relatively constant over long periods of time (centuries). They set the context for drivers
3524 operating at the ecosystem scale. At the other extreme are small-scale, rapidly (less than a day to
3525 a few years) acting drivers with more localized, but sometimes large and/or lasting effects. They
3526 include individual events such as fires, floods, droughts, disease outbreaks, or toxic spills as well
3527 as variables exhibiting daily (e.g. tides), seasonal (e.g. seasonal upwelling or oxygen sags), or
3528 multi-year (e.g. strong upwelling) variations and cycles. The operational time scale of these fast
3529 variables is similar to biological response times (e.g., spawning migration and mating,
3530 reproductive cycle, and life span; see Figure 2 in Kimmerer 2004) and human planning horizons.
3531 These coinciding temporal scales may be the reason that fast variables have received a
3532 disproportionate amount of attention and many of them have come to be regarded as undesirable
3533 stressors. From an ecosystem resilience perspective, however, slower ecosystem drivers –
3534 usually just a few per ecosystem – that operate at intermediate scales are most important
3535 (Carpenter and Turner 2000). These “critical slow variables” (Chapin et al. 2006) include
3536 presence and redundancies of particular functional types of organisms, disturbance regimes,
3537 hydrological and nutrient regimes, biogeochemical processes, and biotic and abiotic landscape
3538 features. These variables change relatively slowly (years to decades). They provide important

3539 stabilizing feedbacks that may counteract the effects of more rapid drivers, thus influencing
3540 resilience and the likelihood of regime shifts. More than any other drivers, slow variables
3541 establish the identity of an ecosystem, including its structure (the biological and physical
3542 components) and functions (processes), and longevity as an entity. In spite of their importance,
3543 they have received relatively less attention, as have the linkages between drivers interacting
3544 across temporal and spatial scales (Groffman et al. 2006).

3545
3546 An ecological regime shift is a large change in the identity of an ecosystem such that there is a
3547 transition from one ecological state (regime A) to another (regime B) (Figure 36). When the
3548 ecosystem responds more or less linearly to changes in drivers, the regime change happens
3549 gradually. When there is a non-linear response to a relatively small change in drivers, change
3550 happens more abruptly and is often referred to as a “regime shift” (Scheffer and Carpenter 2003).
3551 If the trend in drivers reverses direction, the system may respond with a transition back to regime
3552 A that closely mirrors its previous transition to regime B; however, a return to regime A may not
3553 occur (Davis et al. 2010). Some change is simply irreversible because key attributes of the
3554 original regime have been irretrievably lost. In other cases, the back-transition may require a
3555 larger change in drivers than expected based on the original transition. This is particularly
3556 common for non-linear changes, where driver thresholds for the back-transition may be different
3557 than the thresholds for the original transition. This pattern is called “hysteresis,” or system
3558 memory. Hysteresis is the result of regime-specific internal feedbacks most often associated
3559 with slow variables. These features of regime shift have many important implications for
3560 ecosystem management and restoration such as surprising collapses, slower than expected
3561 recovery, or overall unexpected outcomes of recovery efforts. Consequently, adaptive
3562 management approaches require flexibility with a strong learning (science) component.

3563
3564 Drivers of ecosystem change include both natural drivers and human activities. In the remainder
3565 of this report, we define natural and anthropogenic drivers that produce adverse or undesirable
3566 changes in ecosystem structure and functions as stressors. Like many other estuaries around the
3567 world, the San Francisco Estuary has been severely altered by humans over the last 150 years
3568 (Lotze et al. 2006). Some of the changes were gradual, while others were more rapid (see
3569 Atwater et al. 1979, Nichols et al. 1986, Lund et al. 2007, Healey et al. 2008, Moyle et al. 2010).
3570

3571 *Environmental History: Four Eras*

3572
3573 Here, we provide a brief summary of four major periods in the environmental history of the San
3574 Francisco estuary: (1) Pre-European; (2) Gold Rush; (3) Post-Reservoir; and (4) POD. We
3575 selected these periods based on our understanding that each transition between these periods
3576 involved an extreme change to the ecosystem with major structural and functional alterations. In
3577 many cases, the transitions were associated with important slow variables.

3578
3579 Our summary of the Post-Reservoir and POD periods relied mostly on actual data from sampling
3580 in the region. By contrast, data are largely lacking for the first two periods. There are some data
3581 on salinity regimes and landscape changes (e.g. Byrne et al. 2001, Atwater et al. 1979), but
3582 almost no quantitative information about the biota. In the absence of these data, much of the
3583 analysis of earlier periods is based on professional judgment and reasonable assumptions about

3584 how the ecosystem likely functioned. As a consequence, the description of the Pre-European and
3585 Gold Rush periods should be interpreted with caution.

3586

3587 *1. Pre-European Period*

3588

3589 Before the arrival of Europeans, the San Francisco estuary was characterized by high landscape
3590 heterogeneity and hydrological variability with strong oscillations at tidal, seasonal, and decadal
3591 scales (Enright and Culbertson 2010). Sediment cores indicate that the three millennia prior to
3592 European colonization had two extended dry periods lasting centuries during which salinities in
3593 the Delta were relatively high (Byrne et al. 2001). Except for these dry periods, however,
3594 salinity in the Delta was generally very low. At the time of European arrival, the estuary was in
3595 an extended wet period that started around 1250 AD and was characterized by high freshwater
3596 inflows (Byrne et al 2001). Seasonally, inflows and outflows were highest in the winter and
3597 spring and lower in the summer and fall. High San Joaquin River inflows extended longer into
3598 the summer than inflows from the Sacramento River due to later snow melt from higher
3599 mountain ranges in the southern Sierra Nevada (Moyle et al 2010). Mean Delta salinities during
3600 this period were overall very low (Byrne et al 2001) and probably similar to those documented in
3601 early technical reports, which placed the boundary between freshwater and saltwater near
3602 Carquinez Straits (Means 1928 cited in Contra Costa Water District 2010). This has since been
3603 confirmed by paleoecological studies (summarized in Contra Costa Water District 2010).
3604 Seasonally, the Delta may have occasionally become somewhat salty in the summer and fall, but
3605 would always return to freshwater in the winter and spring.

3606

3607 The pre-European Delta landscape was not untouched by humans – many native American tribes
3608 lived in and around the Delta. The Delta they knew was characterized by extensive inland tule
3609 marsh, seasonal floodplain, and complex channel geometry including many small tributary
3610 channels and dead-end sloughs as well as larger channels with natural levees of varying sizes
3611 (Moyle et al. 2010, R. Grossinger and A. Whipple, SFEI, unpublished data). This complexity
3612 likely resulted in substantial spatial heterogeneity of various habitat attributes such as water
3613 temperature. On the other hand, small-scale (fast) temporal variability may have been lower than
3614 later on. For example, the extensive tidal marshes (a slow variable) likely muted tidal effects by
3615 dissipating tidal energy, thus reducing short-term variability in salinity and possibly other water
3616 quality variables (Moyle et al. 2010, Enright and Culbertson 2010). Other pre-European
3617 ecosystem attributes are less certain, but may have included higher levels of biological-based
3618 turbidity and lower dissolved nutrient concentrations. This represents the baseline period for our
3619 review, when productivity of native phytoplankton, zooplankton, fishes, and waterfowl
3620 (migratory and resident) were likely at peak levels and invasions by non-native species were
3621 rare. In addition to the native resident and migratory pelagic fishes, the pelagic zone was
3622 inhabited by relatively large invertebrates including native mysid shrimp and calanoid copepods.
3623 Primary producers included the phytoplankton community that likely underwent regular spring
3624 bloom cycles and likely featured abundant nutritious species, as well as benthic algae and
3625 various types of macrophytes. In addition to autochthonous phytoplankton production, young
3626 detritus from adjacent tidal marsh production, easily broken down by microbes, likely subsidized
3627 pelagic production in the summer and fall. Allochthonous riverine and floodplain phytoplankton
3628 and detritus might have been a particularly important subsidy in the spring.

3629

3630 2. *Gold Rush Period*

3631

3632 The Gold Rush Period, starting in the mid-1800s, was characterized by perhaps the most extreme
3633 and rapid changes to the estuary (Atwater et al. 1979). Many of these changes resulted from
3634 human engineering activities such as mining, levee construction, draining of wetlands and
3635 dredging of channels. Mercury was mined in the Coast Ranges for use in gold mining in the
3636 western Sierras. Both the mercury mining and subsequent loss of portions of the mercury used
3637 during gold mining operations contributed to later concerns about bioaccumulation of mercury in
3638 biota. Abandoned hardrock mines for copper, gold and other metals began contributing acid
3639 mine waste containing a variety of metals to streams in the watershed. Upstream hydraulic gold
3640 mining activities resulted in the mobilization of massive quantities of sediment which increased
3641 sediment loading to peak levels in the late 1800s (Shvidchenko et al. 2004, Meade 1982). This
3642 likely resulted in greatly increased non-biogenic turbidity and mineral sediment deposition in the
3643 estuary. It also resulted in filling of river channels with sediments which decreased the capacity
3644 of channels to transport high flows. The resulting increased flooding of surrounding areas and
3645 filling of navigation channels led to a court decision that abruptly curtailed hydraulic mining
3646 activities in 1884. In addition, levees were constructed along channels to support shipping traffic,
3647 reduce flood risks, and flush excess sediments. These actions had the intended effects of
3648 increasing channel capacity and reduce flooding, but also caused substantial losses of marginal
3649 habitat (e.g. tules and riparian areas) and marshes as well as major alterations to channel
3650 geometry. In addition, the levees were used to isolate areas for drainage and conversion into
3651 farmland. Overall, the formerly complex estuarine landscape was simplified into large, diked
3652 polders locally called “islands” intersected by a roughly linear grid of relatively deep, steep-sided
3653 channels. Development activities also likely increased inputs of nutrients and some
3654 contaminants, but loss of wetlands may have reduced inputs of dissolved and particulate organic
3655 carbon. Loss of shading by tules and riparian vegetation in the Delta and upstream in the
3656 watershed may have increased average water temperatures to some degree but more likely
3657 decreased temperature heterogeneity. The simplified habitat geometry from loss of dead-end
3658 sloughs likely resulted in changes in mixing that may also have further decreased temperature
3659 heterogeneity.

3660

3661 Salinity in the Delta appears to have remained low until upstream diversions started significantly
3662 reducing Delta outflows in the 1920s (Contra Costa Water District 2010). Low salinity until the
3663 1920s was associated with high levels of freshwater diatoms in sediment cores (Byrne et al.
3664 2001). However, phytoplankton productivity probably decreased substantially as a result of light
3665 limitation from increased turbidity levels caused by increased suspended sediment. There is little
3666 doubt that the huge input of hydraulic mining sediments radically altered the benthic community.
3667 Benthic microalgae, once likely a major component of the estuarine primary producer
3668 community, were likely subjected to higher rates of deposition of fine materials from upstream
3669 mining activities. Similarly benthic invertebrates would have been subject to higher
3670 sedimentation rates. This may have favored benthic plants and animals with high tolerances for
3671 silt and the ability to avoid smothering, which may have caused changes in species composition,
3672 but there is no historical information to evaluate this hypothesis. A drop in primary productivity
3673 likely cascaded upward to zooplankton. Zooplankton composition may have been altered by the
3674 start of invasions from ballast water, the intentional introduction of sport fishes (and water from

3675 their native habitats), and more frequent salinity intrusions after 1920 with especially high
3676 salinity levels during the 1928–1934 drought.

3677
3678 Following the Gold Rush, Europeans introduced many fishes intended to provide food or sport.
3679 The early introductions included American shad, striped bass, centrarchids (bass and sunfish)
3680 catfish and carp (Dill and Cordone 1997). American shad and striped bass did particularly well,
3681 thriving in the high turbidity estuary. Native species are thought to have declined substantially
3682 during this period because of habitat alteration and competition from introduced species. The
3683 once common native thicktail chub started declining in the late 19th century and finally became
3684 extinct in the late 1950s. Salmonids, particularly Chinook salmon were heavily fished during
3685 this period. Mining activities resulted in sedimentation and loss or reduced quality of spawning
3686 gravels for salmon and all other riffle spawning species. At the same time access to upstream
3687 spawning areas was reduced due to relatively small and sometimes temporary dams and
3688 diversions constructed for mining or agricultural purposes. The construction of larger,
3689 permanent dams accelerated through this period for municipal and agricultural water supply and
3690 flood protection. The area of dam construction reached its peak with the construction of the
3691 SWP and CVP.

3692
3693 *3. Post-Reservoir Period*

3694
3695 The construction of reservoirs and the CVP and SWP started in the first half of the 20th century
3696 and led to additional extreme changes to the estuary. The hydrology of the estuary was altered
3697 by reservoirs that held spring run-off and then released stored water during summer and fall for
3698 municipal and agricultural use (Mount 1995, Moyle et al. 2010). These changes reduced flood
3699 flows, which reduced seasonal inundation of floodplains. The absence of flooding, also
3700 facilitated by levee construction, allowed development of former floodplains for agricultural and
3701 urban uses. The reservoirs also disrupted the sediment transport processes by intercepting and
3702 storing sediments transported down from the Sierra Nevada. Sediments immediately below the
3703 dams that could be transported by the altered flow regime were eventually depleted and could
3704 not be renewed from above the reservoir. The construction of levees, loss of floodplain, and
3705 reduction in flooding, decreased sediment recruitment from the floodplains and banks of the low
3706 elevation portions of the larger rivers.

3707
3708 Water diversions from the Delta increased with increased urban and agricultural demands due to
3709 steady population growth. Year-round water exports from the Delta started in 1968 after the
3710 completion of San Luis Reservoir and eventually led to mostly negative (upstream) river flows in
3711 the OMR channels of the central and southern Delta. Upstream and in-Delta diversions reduced
3712 inflows from the San Joaquin River to a minor proportion of freshwater inflows into the Delta.
3713 After removing seasonal and decadal-scale oscillations, the long-term trend in outflow was
3714 positive from 1929 until 1960 and followed the long-term positive trend in precipitation.
3715 Precipitation continued to increase after 1960, but the long-term trend in outflow turned negative
3716 after 1960. There was overall lower outflow for the period from 1968 to 2006 compared to the
3717 previous four decades. The greatest outflow reductions occurred during the last two decades
3718 (Enright and Culbertson 2010). Reservoir and water project operations also had a dampening
3719 effect on outflow variability; however, both seasonal and annual outflow variability became
3720 overall more, not less, variable in the post-reservoir period due to overall wetter hydrology

3721 during this period. To date, the Mediterranean climate of California remains the most powerful
3722 driver of long-term variability in outflow at the seasonal and interannual scale (Enright and
3723 Culbertson 2010). The spatial extent of salinity intrusions peaked early in the 20th century before
3724 fresh water was drawn from the Sacramento River across the Delta to the CVP and SWP export
3725 facilities. However, while saltwater was kept out of the central Delta by drawing fresh water
3726 across the OMR corridor, the western Delta and Suisun Bay became increasingly salty in the
3727 summer months (Contra Costa Water District 2010) and, similar to outflow, exhibited more
3728 variability in salinity after 1968 than in the previous decades (Enright and Culbertson 2010).
3729 Several multi-year droughts tested the capacity of the system to satisfy the competing human and
3730 environmental needs for water and other ecosystem services. This is in spite of the relatively
3731 stable and moderate climate enjoyed by California over the last 150 years (Malamud-Roam et al.
3732 2007) and higher precipitation than in the previous period (Enright and Culbertson 2009).

3733
3734 Nutrient inputs increased as a result of inputs of fertilizer in agricultural and urban runoff. The
3735 expansion of agricultural and urban land uses in the San Francisco Estuary and watershed also
3736 increased the diversity and amounts of contaminants in the system. Organochlorine insecticides,
3737 including dichlorodiphenyltrichloroethane (DDT), were already being phased out during this
3738 period but were replaced by organophosphate pesticides. Residues of organochlorine compounds
3739 can still be found throughout the system. A wide array of herbicides was used during this time
3740 period. Waste products from industrial processes led to inputs of a wide variety of contaminants.
3741 Urban uses of pesticides, fertilizers and other industrial products resulted in contaminant inputs
3742 through urban runoff and in treated wastewater. Many such inputs were reduced after their
3743 detrimental effects were recognized through implementation of legislation, such as the Clean
3744 Water Act, or other regulations. These reductions resulted in real benefits to the environment
3745 and organisms; however, low-level, sub-lethal inputs continued. New pesticides and other
3746 chemicals continued to be developed, used, and transported into aquatic ecosystems.

3747
3748 Biologically, the post-reservoir period is characterized by an accelerating rate of invasion by
3749 alien species, especially at lower trophic levels (Nichols et al. 1986, Cohen and Carlton 1998).
3750 There were multiple invasions of mysids, copepods, and jellyfish, with an overall drop in body
3751 size of crustacean zooplankton (Winder and Jassby 2010). Some of the most extreme changes
3752 occurred in the benthic community, where successive invasions by the bivalves *C. fluminea* and
3753 *C. amurensis* led to their dominance of this habitat (Peterson and Vayssières 2010). These
3754 bivalves greatly increased benthic grazing pressure on planktonic phytoplankton and
3755 zooplankton (Jassby et al. 2002). The zooplankton community was also almost entirely replaced
3756 by non-native species (Winder and Jassby 2010). The dramatic displacement of indigenous
3757 pelagic and benthic primary consumers by introduced species during this period was
3758 accompanied by changes in primary producers notably: (1) the decline of diatoms; (2) increases
3759 in flagellates; and (3) an overall decline in phytoplankton cell size, biomass and productivity
3760 (Lehman 2000, Jassby et al. 2002). Finally, two exotic aquatic macrophytes also appeared in the
3761 Delta during this period. The floating aquatic macrophyte *Eichhornia crassipes* (water hyacinth)
3762 started proliferating in the early 1980s; however a successful control program has reduced the
3763 ecological effects of this invasion. *E. densa* was introduced in the 1960s, but did not reach
3764 nuisance levels until after the 1987–1992 drought (Jassby and Cloern 2000). In addition to their
3765 direct effects on native species via competition and predation, some of the invasive species,
3766 especially *E. densa*, also act as powerful “ecosystem engineers” (Jones et al. 1994, Breitburg et

3767 al. 2010), adding to the human reengineering of the system during the Gold Rush and Post-
3768 Reservoir periods.

3769
3770 Downward trends in several fish species became apparent in the second half of the 20th century,
3771 resulting in the eventual listing of seven fish species under the CESA or FESA, including two of
3772 the POD species. Declining turbidity was implicated in the long-term declines of delta smelt,
3773 longfin smelt, and age-0 striped bass. Increases in water exports, X2, and changes in the food
3774 web also contributed to varying degrees to the long-term decline in the POD fish species
3775 (Thomson et al 2010, Mac Nally et al. 2010). There were relatively fewer introductions of fishes
3776 during the Post-Reservoir Period. Increasing populations of previously introduced littoral fishes,
3777 particularly largemouth bass and other centrarchids which benefitted from expanding SAV in
3778 shallow water habitat may have affected native species, especially through predation (Dill and
3779 Cordone 1997, Brown and Michniuk 2007, Moyle and Bennett 2008). Similarly, Mississippi
3780 silverside, a habitat generalist, may affect native species through both competition and predation
3781 on egg and larval life stages (Bennett 2005, B. Schreier, CDWR, personal communication).

3782 3783 4. *POD Period*

3784
3785 The POD Period represents a period of continued decline in turbidity as the last of the mining
3786 sediments were transported out of the river and estuarine channels. Careful control of reservoir
3787 operations and exports resulted in a stabilized hydrology (Moyle et al. 2010) and a relatively
3788 upstream location of X2 especially in the fall irrespective of water year type (Figure 26; Feyrer et
3789 al. 2007). Another major change to physical habitat during this period was the rapid expansion
3790 of *E. densa*, which likely benefitted from the more hydrologically stable conditions and reached
3791 peak levels (Brown and Michniuk 2007, Baxter et al. 2008, Hestir 2010). The proportion of San
3792 Joaquin River water in the Delta was low relative to Sacramento River contributions. Water
3793 temperatures were relatively warm as compared to the previous decade. Ammonium levels
3794 reached their highest levels as a result of urban inputs (Jassby 2008), while phosphorus inputs
3795 (total phosphorus and soluble reactive phosphorus) declined due to an abrupt 2-fold decrease in
3796 urban phosphorus loading in the mid-1990s (Van Nieuwenhuysse 2007). Similarly, pyrethroid
3797 pesticides, herbicides and “emerging contaminants” showed increases because of agricultural
3798 and urban inputs, but other pesticides, including Ops, declined (Oros and Werner 2005).

3799
3800 Phytoplankton biomass and primary production in the Delta and Suisun Bay declined to the
3801 lowest levels ever observed in the mid-1990s and recovered somewhat in the Delta during the
3802 POD period; however, there continued to be major changes in phytoplankton species
3803 composition (Baxter et al. 2008). Perhaps the most substantial change in phytoplankton was the
3804 increase in the frequency, duration, and geographical extent of *M. aeruginosa* blooms (Lehman
3805 et al. 2008a). Similarly, there were continued changes in the zooplankton community, notably
3806 the virtual disappearance of native mysids and the increase of the cyclopoid copepod *L.*
3807 *tetraspina* (Orsi and Kimmerer 1986, Bouley and Kimmerer 2006, Winder and Jassby 2010).
3808 Jellyfish adapted for brackish water increased in the western Delta, Suisun Bay and Suisun
3809 Marsh (Schroeter 2008). Pelagic fishes reached historic lows, while invasive species such as
3810 largemouth bass, other centrarchids, and inland silversides reached their highest levels (Moyle
3811 and Bennett 2008).

3812

3813 *Human History – Then and Now*

3814

3815 The four ecological eras of the San Francisco Estuary summarized above were closely linked
3816 with social and scientific developments in California and the nation. Here we briefly touch on a
3817 few of these changes because they pertain to our understanding of the POD in a historical
3818 context.

3819

3820 Many of the human-caused environmental changes over the past 150 years were intentional and
3821 the outcomes were initially considered highly desirable based on the prevailing social values.
3822 Some are still regarded as desirable by some segment of society. Examples include the
3823 “reclamation” of seasonal and tidal wetlands to support agricultural production and urban
3824 development, intentional introduction of exotic species, prevention of seasonal flooding,
3825 diversion of water from rivers to support a growing human population, and the dilution,
3826 transformation and export of wastes in waterways. Other changes that have occurred were
3827 largely unnoticed because they did not appear to directly affect ecosystem services valued by
3828 humans. Examples of these types of changes include changes in water quality that ultimately
3829 affect ecosystem productivity but do not exceed legal thresholds, and changes in sedimentation
3830 processes that slowly affect clarity of downstream waters. The effects of both types of changes
3831 on important slow variables, such as hydrological and landscape heterogeneity, that previously
3832 had defined and maintained the unique identity of the system during the pre-European period
3833 were largely ignored in favor of more immediate gains through control of adverse (to humans)
3834 impacts of fast variables, such as individual flood or drought events and seasonally unreliable
3835 freshwater supplies. It has now been recognized that some of these changes have affected other
3836 segments of society. For example, decreasing fish populations have affected the economic well
3837 being of some recreational and commercial fishing industries. Perceptions of value have also
3838 changed with society placing new emphasis on the recreational, aesthetic and ecological
3839 functions of natural systems, including the San Francisco Estuary and watershed.

3840

3841 New scientific findings about ecosystems, increasing environmental consciousness, and finally
3842 the new environmental laws of the 1960s and 1970s represented an important scientific and
3843 social change. In the Delta and Suisun Bay, this resulted, among other things, in the initiation of
3844 long-term ecological monitoring and studies by the IEP and others in the 1970s. The latest
3845 development in this context was the enactment new water-related legislation by the California
3846 legislature in 2009. This comprehensive legislative package included the creation of a new Delta
3847 Stewardship Council charged with promoting the: “coequal goals of providing a more reliable
3848 water supply for California and protecting, restoring, and enhancing the Delta ecosystem. The
3849 coequal goals shall be achieved in a manner that protects and enhances the unique cultural,
3850 recreational, natural resource, and agricultural values of the Delta as an evolving place”
3851 (California Water Code §85054).

3852

3853 The co-equal goals as well as the creation of the Delta Stewardship Council are in general
3854 agreement with a recent change in natural resource management toward “ecosystem-based
3855 management” (McLeod et al. 2005) and “resilience-based ecosystem stewardship” of “social-
3856 ecological systems” (Chapin et al. 2009), which increases adaptive capacity (Folke et al 2002).
3857 The codification of the need to protect the Delta ecosystem as a whole, not just specific
3858 ecosystem services or beneficial uses, is particularly consistent with these changes in the

3859 approach to resource management. Ideas associated with this new management model are
3860 rapidly gaining momentum (e.g., The Johnson Foundation 2010). This represents a fundamental
3861 shift from the traditional and still common “steady-state management”, which focuses on
3862 preventing change and reducing variability of natural systems, to adaptive management, which
3863 incorporates and addresses change (Folke et al. 2002).

3864

3865 In the traditional management model, natural systems are considered largely separate from
3866 human systems and are assumed to be relatively stable (unchanging) and infinitely resilient
3867 (Folke et al 2004). The traditional management model ignores changes in slow variables (which
3868 are simply considered stable) and focuses mostly on rigidly controlling fast variables perceived
3869 as undesirable disturbances. The model maximizes a few select ecosystem services (e.g.,
3870 provisioning of water or food) while discounting others (e.g., nutrient cycling). New
3871 management models include change as a fundamental property of social-ecological systems.
3872 Instead of suppressing change, management can include variability at various temporal and
3873 spatial scales in ways that sustain all ecosystem services (Chapin et al 2009).

3874

3875 Fundamental to the thinking of these new theorists is that ecological resilience cannot be taken
3876 for granted (Folke et al. 2004). Reduced resilience is related to changes in slow variables and
3877 thereby to recent undesired and unexpected changes in many ecological structures and functions.
3878 These unexpected outcomes are opposite to the intent and assumptions of traditional steady-state
3879 management, which emphasizes stability and reliable provision of certain ecosystem services
3880 (e.g., water supply reliability). Many ecosystems are currently experiencing rates of change
3881 much greater than during the past 10,000 years (the Holocene). These high rates of unexpected
3882 change lead to catastrophic shifts to less desirable, degraded states (Crutzen 2004, Rockstrom et
3883 al. 2009). On a regional scale, reports of major and often rapid changes in ecosystems are
3884 increasingly common and include collapses of important natural resources, from fisheries to
3885 coral reefs to forest systems (Folke et al. 2004). Many of these changes are so dramatic they have
3886 been called regime shifts. We hypothesize that the POD was one such rapid, unexpected
3887 collapse.

3888

3889 **A Tale of Two Estuaries**

3890

3891 **Why a new conceptual model?**

3892

3893 Here, we introduce the idea of an ecological regime shift in the Delta and Suisun regions of the
3894 estuary as a new conceptual model and working hypothesis for the POD. The regime shift
3895 conceptual model expands the POD beyond the original focus on pelagic fishes to the ecosystem
3896 level and places it in a historical context. This allows integration of the POD with other changes
3897 in the system that may not be a clear part of the narrower POD story, but may be essential to
3898 understanding and managing the system in its current state. For example, the new conceptual
3899 model includes changes in population dynamics and abundance of migratory fish species, such as
3900 salmonids and sturgeons, and in littoral and benthic communities. It also includes changes that
3901 happened or began long before the POD. In addition to providing a more inclusive narrative
3902 outline, the regime shift concept focuses attention on the ideas of system resilience and

3903 ecological thresholds, which have implications for choice of management approach. The regime
3904 shift approach provides for comparison of the San Francisco Estuary with other systems
3905 undergoing regime shifts. Thus, lessons learned in other systems about how to avoid or reverse
3906 regime shift can be more directly translated to the Delta (Biggs et al. 2009). Further exploration
3907 of the POD in the regime shift context should provide insights into possible future alternative
3908 states that may arise as a result of management interventions or other changes in environmental
3909 drivers and may usher in a new era of “resilience science” (McLeod and Leslie 2009) in the
3910 estuary and watershed. Ultimately, this new conceptual model, the associated extended POD
3911 narrative, and the new scientific endeavors that might arise from it are aimed at informing new
3912 resilience-based, adaptive ecosystem management and conservation strategies to build or rebuild
3913 adaptive capacity of the estuary and watershed. For the purpose of this report, however, the
3914 regime shift concept serves to integrate history, drivers, and affected ecosystem attributes in a
3915 cohesive narrative with gradual changes in ecosystem resilience, short-term disturbance, and
3916 random effects on populations and communities.

3917

3918 *Ecological regime shift*

3919

3920 As described earlier, ecological regime shifts are rapid, large-scale, lasting changes in
3921 ecosystems from one more-or-less stable regime (or state) to another (Scheffer and Carpenter
3922 2003) caused by a non-linear system response to drivers of change. The term regime shift was
3923 first introduced in the ecological literature to describe the fluctuating stocks of sardines and
3924 anchovies (Lluch-Belda et al. 1989). Regime shifts have since been described worldwide for a
3925 wide variety of ecosystems including oceans, lakes, estuaries, and terrestrial systems (reviewed
3926 in Folke et al. 2004). A number of studies support the idea of a recent regime shift in the upper
3927 San Francisco Estuary associated with the POD (Manly and Chotkowski 2006, Moyle and
3928 Bennett 2008, Mac Nally et al. 2010, Thomson et al. 2010, Moyle et al. 2010).

3929

3930 Regime shifts can occur naturally or in response to human actions, and often come as a surprise.
3931 A popular analogy (Figure 37) uses an image of a ball to represent a community or ecosystem in
3932 a system of valleys, which represent alternative stable states. A regime shift is represented by
3933 the ball rolling from one valley across a ridge (an unstable state or threshold) into an adjacent
3934 valley (or alternative stable state) (Beisner et al. 2003).

3935

3936 If the initial state is considered desirable by society, the regime shift is regarded as undesirable
3937 due to costly consequences to humans and natural resources, and because of difficulties in
3938 reverting back to more desirable stable states. The severity of a regime shift and the potential for
3939 reversal depends on the resilience of the ecosystem. It is important to note that resilience does
3940 not mean that changes do not occur. Rather, the system may absorb disturbance by reorganizing
3941 and adapting such that essential “functions, structure, identity, and feedbacks” of the system are
3942 maintained (Folke et al. 2004).

3943

3944 In theory, regime shifts happen either because drivers push communities beyond the limits of
3945 their resilience and into an alternate stable state within a constant environment (the ball moves
3946 from one valley to another), or because environmental drivers change the environment so the
3947 stability of one state is reduced while an alternate state becomes more stable (the valley and ridge

3948 topography changes) (Beisner et al. 2003). An example of the former mechanism might be a
3949 combination of overfishing and species invasions and associated trophic cascades (Daskalov et
3950 al. 2007). The latter mechanism can include a slow and often imperceptible reduction of
3951 resilience (valley topography) by changes in important slow variables that sustain important
3952 internal feedbacks. At some point a threshold is exceeded and there is a rapid and unexpected
3953 regime shift. In the context of the ball analogy, changes in slow variables can erode resilience by
3954 moving valleys closer together and lowering the height of the ridge around the ball, making a
3955 regime shift more likely given current drivers. As mentioned above, such shifts may not be
3956 easily reversible. This phenomenon is known as hysteresis. In other words, pushing the ball
3957 back uphill is hard and the ball may not actually return to the original valley, which may have
3958 irretrievably disappeared.

3959
3960 Finally, random events (i.e., stochasticity) also play an important role in regime shifts. The final
3961 push into a different state may come from random variability in populations and communities or
3962 the environment. In the ball analogy, the ball does not sit perfectly still but changes position in
3963 response to random variability in environmental drivers, internal feedbacks of the ecosystem or
3964 both. If a random event is strong enough, an unexpected regime shift can occur and can even
3965 lead to rapid extinction (Melbourne and Hastings 2008, Hastings and Wysham 2010). Such a
3966 regime shift would be unexpected because it could occur even though known drivers remained
3967 within the previously observed range of variability. Random effects may also come into play
3968 with respect to Allee effects (Allee 1931, Dennis 1989, Courchamp et al. 1999, Berec et al.
3969 2006). This makes forecasting of regime shifts particularly difficult.
3970

3971 *POD – a regime shift*

3972
3973 As depicted in figure 8, we propose that changes in a suite of mostly abiotic, environmental
3974 variables (drivers) led to changes in biological populations and communities in the system that
3975 recently became so profound that a regime shift might have taken place that affected not just the
3976 four POD fish species, but the entire system. We have started exploring the POD in the context
3977 of regime shift and do not yet know if this new regime represents an alternative stable state with
3978 stabilizing feedbacks. More work is also needed to establish the slow and fast drivers that led to
3979 this regime shift. Most of the environmental drivers in figure 8 have been under investigation by
3980 the POD study and have been represented in the habitat and top-down boxes of the basic POD
3981 conceptual model. However, the regime shift model explicitly considers operation of these and
3982 other drivers over larger geographic scales and longer time scales than previously considered. In
3983 addition, the regime shift model includes the effects of the drivers on all species and processes.
3984 Importantly, recent analyses have not identified environmental drivers that can explain the POD,
3985 but have identified drivers associated with long-term trends in POD fishes (Thomson et al.
3986 2010). Most of these drivers changed gradually before the POD. As a group, the environmental
3987 drivers we discuss in this report may represent at least some of the critical slow variables
3988 determining the resilience and adaptive capacity of the Delta ecosystem. Additional critical slow
3989 variables may emerge as we continue to explore this regime shift.

3990
3991 The environmental, slow drivers we propose for the POD regime shift are (1) outflow, (2)
3992 salinity, (3) landscape, (4) temperature, (5) turbidity, (6) nutrients, (7) contaminants, and (8)

3993 harvest. These drivers are listed in our hypothesized order of their importance to the resilience of
3994 the system and approximate rate of change. This order will likely change as we learn more about
3995 the regime shift and associated drivers. The original, pre-regime shift state of the pelagic
3996 community has been presented in the description of the Pre-European period provided above. As
3997 described in the environmental history section above, the POD shift was preceded by changes in
3998 drivers associated with the Gold Rush, changes in land use, and reservoir construction and
3999 operation. A better understanding of prior changes will likely lead to a better understanding of
4000 the POD shift and the current state of the system. Improved understanding of these changes will
4001 also give insights into which drivers might be managed in a way that would improve ecosystem
4002 resilience to future threats and perhaps lead to a more desirable state. We hope to explore the
4003 POD regime shift and prior changes in drivers and thresholds in more detail in the future.
4004

4005 In the following sections, we will briefly summarize changes in the eight drivers identified above
4006 and provide hypotheses about their individual and combined effects on the biota and importance
4007 to resilience. We will also briefly mention additional slow abiotic drivers we have not yet
4008 explored as part of the POD investigation. Finally, we will describe the resulting new ecosystem
4009 regime for the pelagic, benthic, and littoral zones of the Delta and Suisun Bay regions.
4010

4011 1) Outflow – Native species are adapted to the natural flow regime of the estuary which includes
4012 natural variability in the frequency, timing, duration, and rate of change of flow events, and the
4013 time intervals between major floods and droughts (Moyle et al 2010). Due to its Mediterranean
4014 climate and the complex interplay of geomorphology and hydrology, outflow of freshwater from
4015 the Delta to San Francisco Bay and the ocean varies greatly on tidal, seasonal, and interannual
4016 scales (Enright and Culberson 2010, Moyle et al. 2010, SWRCB 2010). Seasonally, winters are
4017 wet and summers are dry. On a decadal scale, multi-year dry and wet periods are common.
4018 Major shifts from overall dry to overall wet phases occur at the scale of centuries. Currently, the
4019 estuary is in an extended wet period that started around 1250 AD. The decades since the middle
4020 of the 20th century have been particularly wet. In spite of some dampening of variability by
4021 water project operations, natural climate (precipitation) variability is still the dominant driver of
4022 seasonal and interannual outflow variability. Outflow may actually be more variable now than in
4023 pre-European times when flows were dampened by large wetland and floodplain areas (Enright
4024 and Culberson 2010). Outflow volume, however, is now about 15% lower than before year-
4025 round water project operations began in 1968 and the long-term trend in outflow is now
4026 decoupled from the long-term trend in precipitation due to flow regulation and water diversions
4027 (Enright and Culberson 2010). Recent fall outflows have been uniformly low irrespective of
4028 water year type (Feyrer et al. 2007).
4029

4030 We hypothesize that in spite of recent human-caused flow alterations, the remaining strong,
4031 climate-driven variability in outflows may presently act as the most important critical slow
4032 variable that helps maintain the remaining original resilience of the system. We further
4033 hypothesize that outflow volume is another important slow driver of ecological change via its
4034 effects on residence time, the estuarine salinity gradient, and other physical processes. Future
4035 changes in outflow variability and volume due to climate change will affect the adaptive capacity
4036 of the system to respond to other changes.
4037

4038 2) Salinity – The pre-European Delta, Suisun Bay, and Suisun Marsh were predominantly fresh
4039 water environments, while much of San Francisco Bay was brackish (Contra Costa Water
4040 District 2010). Overall, long-term trends in salinity were closely coupled to long-term trends in
4041 climate and hydrology (Enright and Culberson 2010). Starting in 1968, the long-term trend in
4042 salinity in Suisun Bay changed from a previously negative association with precipitation to a
4043 positive one. In other words, similar to the long-term trends in outflow volume, trends in salinity
4044 became decoupled from the previously dominant climate driver (precipitation) (Enright and
4045 Culberson 2010). Early water diversions and wetland losses led to increased salinity intrusions
4046 deep into the Delta. Reservoir operations and year-round water diversions that necessitated
4047 freshwater in the Delta shifted the salinity gradient back toward the west, but Suisun Bay
4048 remained brackish. These changes constricted and stabilized the estuarine salinity gradient and
4049 shifted the distribution of species adapted to marine, brackish, and freshwater conditions in the
4050 estuary from pre-European times. Jassby et al (1995) showed that the location of the 2 psu
4051 bottom isohaline along the axis of the estuary (distance from the Golden Gate in km) had simple
4052 and significant statistical relationships with annual measures of many estuarine resources
4053 including phytoplankton, clams, zooplankton, and fishes. While several of these X2 relationships
4054 changed over time (Kimmerer 2002), many have remained intact. The invasion by the clam
4055 *Corbula amurensis* and a host of other species as well as the change in some of the X2
4056 relationships coincided with the low flow/high salinity conditions during the extended 1987–
4057 1992 drought (Winder and Jassby 2010). The recent decline in fall outflow has led to increased
4058 and interannually less variable fall salinity in the western Delta and Suisun Bay since the 1987–
4059 1992 drought and especially during the POD years (Figure 26 and Figure 38; Winder and Jassby
4060 2010). The constriction in habitat of suitable salinity (and turbidity) may have contributed to the
4061 decline of delta smelt and other pelagic organisms (Feyrer et al. 2010). The more saline
4062 conditions in Suisun Bay likely favored invasive species such as *Corbula* and invasive jellyfish,
4063 while the uniformly freshwater conditions in the Delta contributed to the widespread
4064 colonization by *E. densa*.

4065
4066 We hypothesize that the position, extent, and variability of the estuarine salinity gradient
4067 represents another critical slow variable. A westward position of the salinity gradient coupled
4068 with sufficient variability, including occasional salinity intrusions into the Delta and freshwater
4069 flows reaching San Pablo Bay, favors native species. A more eastward and more constricted and
4070 stable salinity gradient favors more non-native and nuisance species. In spite of remaining
4071 substantial variability in salinity, the long-term increase in salinity in Suisun Bay, and
4072 constriction of the salinity gradient likely contributed significantly to the erosion of the resilience
4073 of the original ecological regime. The current situation contributes to the stabilization of the new
4074 regime.

4075
4076 3) Landscape – As described above, the pre-European estuarine landscape looked and functioned
4077 very differently from today’s landscape. By geological standards, the San Francisco Estuary is
4078 young – only about 6,000 to 10,000 years old – and its newly evolving landscape likely underwent many
4079 substantial transformations. Some of these transformations occurred slowly over centuries, while others
4080 occurred much more rapidly, providing for a heterogeneous, variable mix of habitats. The most rapid and
4081 dramatic transformation occurred, however, over the last 150 years. Many original landscape features
4082 such as the vast wetlands, riparian forests, and floodplain areas were lost, and replaced with new features
4083 including a grid of deep, steep-sided channels, dry polders (islands), and large shallow lakes (flooded
4084 islands). With the exception of the addition of flooded islands and some small wetland restoration areas,

4085 the current landscape has remained relatively unchanged for nearly half a century. The current landscape
4086 is not the landscape in which the native species of this estuary evolved. It likely favors other species
4087 better adapted to the altered conditions. The loss of habitat in combination with predation and
4088 competition from introduced species is thought to be important in the declines of many native
4089 species (Brown and Moyle 2005), including the extinct thickettail chub and Sacramento perch,
4090 which has completely disappeared from the Delta.

4091
4092 We hypothesize that the current rigid, simplified landscape configuration is a critical slow driver
4093 that stabilizes the POD regime. The dramatic landscape transformation that started 150 years ago
4094 was likely the single greatest driver of ecological change in the system and responsible for much
4095 of the erosion in its original ecological resilience. Due to the focus on recent changes, the POD
4096 investigations have not yet included studies of past landscape transformations and their
4097 interactions with other drivers. In addition to better understanding what led to the POD, such
4098 studies would be helpful for guiding restoration planning.

4099
4100 4) Temperature – Water temperature in the Delta was once likely as variable as its hydrology due
4101 to the Mediterranean climate, variable hydrology, and landscape heterogeneity. Winter and
4102 spring flows into the estuary were likely quite cold. Water temperatures during the warmer
4103 months were likely heterogeneous because of the variable, complex landscape with dead-end
4104 sloughs having little exchange with larger downstream channels. The high degree of
4105 connectedness with tidal and non-tidal wetlands likely also played a role. Loss of shading by
4106 tules and riparian vegetation in the Delta and upstream in the watershed after the Gold Rush may
4107 have increased average water temperatures to some degree but more likely decreased
4108 temperature heterogeneity. The simplified habitat geometry from loss of dead-end sloughs likely
4109 resulted in changes in mixing that may also have decreased temperature heterogeneity. The role
4110 of cool groundwater contributions to tributary rivers is also unknown but may have kept rivers
4111 cooler before flowing into the Delta than currently.

4112
4113 We hypothesize that the warmer temperatures and more uniform temperature distributions in the
4114 contemporary Delta act as a slow driver that stabilizes the current regime by favoring species
4115 adapted to higher temperatures and more stable conditions such as largemouth bass and *M.*
4116 *aeruginosa*, but negatively affects species adapted to cooler and more heterogeneous
4117 temperatures such as delta smelt and salmon. Predicted climate change effects on water
4118 temperatures pose additional threats to native species.

4119
4120 5) Turbidity – The turbidity regime in the pre-European wetland-dominated Delta was likely
4121 very different from what it is now. Wetland-derived organic matter likely played a larger role
4122 year-round role. Sediment derived turbidity may have often been higher compared to the present
4123 because sediment transport processes would have been fully functional, recruiting fine sediments
4124 from the Sierra Nevada, lowland valleys, and wetlands. Seasonal variability in the rivers was
4125 likely high; however, wind resuspension of sediments in Suisun Bay and the Delta may have
4126 kept turbidity high when river inputs were low. Overall, sediment loading to the Delta increased
4127 strongly during the Gold Rush era and has since decreased (Shvidchenko et al. 2004, Wright and
4128 Schoellhamer 2004), leading to declining total suspended solid concentrations and associated
4129 declining turbidity at most IEP monitoring stations over the last 40 years (Jassby 2008). In
4130 addition to being trapped behind dams on tributaries, sediments are also eroded and flushed from
4131 areas below the reservoirs during winter storm events, but that source is not renewed from

4132 upstream so it has been diminishing. Wright and Schoellhamer (2004) showed that peak
4133 sediment concentrations in Sacramento River water associated with the particularly strong flood
4134 events of 1964, 1986 and 1997 have been declining from one strong flood event to the next due
4135 to reduced sediment yield from the watershed. In an analysis of total suspended solid data
4136 monitoring data from 1975 to 1995, Jassby et al. (2005) showed that concentrations in the north
4137 Delta dropped sharply toward the end of the 1982–1983 El Niño-Southern Oscillation (ENSO)
4138 event and did not recover afterward, likely due to erosion and flushing of previously stored
4139 sediments from the Delta. Suspended solids further decreased from 1996 to 2005, although not as
4140 strongly as over the previous decades (Jassby 2008). Jassby (2008) did not report the effect of
4141 the 1997–1998 El Niño event on suspended solids, but a more recent analysis (Hestir et al.,
4142 UCD, unpublished data) showed that it contributed to another decline in suspended solids,
4143 flushing sediments from Suisun and San Pablo bays. The narrow, steep-sided, and deep channels
4144 and associated levees of the northern Delta and upstream tributaries were originally designed to
4145 quickly flush the massive influx of sediments from the hydraulic mining activities of the Gold
4146 Rush era, so these effects are not surprising. In addition, the spread of invasive macrophytes,
4147 especially *E. densa*, in the Delta was both facilitated by and contributed substantially to losses in
4148 suspended sediments over the last two decades (Hestir 2010). Sediment loads to the Delta are
4149 now likely at or below pre-European levels. Overall, the increase in sediment loads and
4150 associated turbidity levels in the Delta during the Gold Rush era may have benefitted small fish
4151 like delta smelt and juvenile salmon by providing refuge from visual predators. It also benefitted
4152 alien species adapted to high turbidity such as striped bass. It seems, however, unlikely that such
4153 benefits would outweigh the negative effects of the destruction of riverine and wetland habitats
4154 during this period. In addition, the influx of sediments during this period likely also had strong
4155 negative effects on benthic organisms and primary producers, thus impairing ecosystem
4156 production. The gradual and occasionally steep declines in suspended sediments and associated
4157 turbidity levels in the post-reservoir and POD eras helped explain long-term declines in three out
4158 of four POD fish species (Thomson et al 2010). Turbidity declines were likely detrimental to
4159 smaller fish and possibly also larger and more slowly swimming pelagic invertebrates such as the
4160 native mysids and calanoid copepods because of increased predation risk. In addition, light
4161 availability was clearly the most important limiting factor for phytoplankton production under
4162 more turbid conditions, but as the Delta clears other factors are becoming equally or more
4163 important. Nutrient effects as well as grazing by littoral and pelagic species will likely become
4164 more important, with additional effects on phytoplankton species composition and repercussions
4165 throughout the food web.

4166
4167 We hypothesize that the initial rapid increase followed by the slow decline in sediment loads and
4168 turbidity represents an important driver that both benefitted and disturbed native species. The
4169 most recent sediment-flushing El Niño event of 1997–1998 occurred just before the onset of the
4170 POD and may thus have contributed to the POD regime shift. The current much clearer regime
4171 benefits non-native fish species and amplifies nutrient and grazing effects on phytoplankton.

4172
4173 6) Nutrients – To our knowledge, the pre-European nutrient regime of the estuary has not been
4174 investigated. Early reviews of nitrate and phosphate concentrations in the Delta found that they
4175 were “near the highest concentrations reported from other estuaries” (Kohlhorst 1976), perhaps
4176 especially during dry years (Siegfried et al. 1978). San Joaquin River water generally had much
4177 higher nutrient concentrations than Sacramento River water and nitrate and phosphate

4178 concentrations were lowest during spring phytoplankton blooms. Total nitrogen increased at
4179 most IEP monitoring stations over the last 40 years, while total phosphorus decreased (Jassby
4180 2008). The decline in phosphorus was associated with a rapid decrease in loading from the
4181 SRWTP in the mid-1990s (Van Nieuwenhuysse 2007). The SRWTP became operational in 1982
4182 and uses secondary treatment. It also contributed to the increase in nitrogen, especially the strong
4183 increase in ammonium concentrations in the northern Delta (Jassby 2008). Changes in nutrient
4184 loading did not just affect individual nutrient concentrations, but also nutrient ratios. While
4185 phytoplankton growth is generally much more limited by light availability than nutrients in this
4186 turbid, nutrient rich system, changing nutrient ratios likely affected phytoplankton species
4187 composition, possibly with repercussions throughout the food web (Glibert 2010). In addition,
4188 ammonium in the northern Delta rose to levels that now appear to be suppressing nitrate uptake
4189 (Wilkerson et al. 2006, Dugdale et al. 2007) and perhaps even the uptake of ammonium itself by
4190 sensitive algal groups such as diatoms. This suppresses diatom production (Parker et al. in
4191 review) and likely also affects microbial processes and community composition. New stable
4192 isotope evidence (C. Kendall, USGS, unpublished data) indicates that the toxic *M. aeruginosa*
4193 blooms that have been occurring in the Delta over the last decade may largely be fueled by
4194 ammonium, not nitrate. The increase in ammonium may thus have contributed to the overall
4195 decline in phytoplankton over the last 40 years (Jassby et al. 2002) and long-term shifts in
4196 species composition (Lehman 2000). As has been shown elsewhere, elevated levels of
4197 ammonium and other nutrients may also benefit invasive rooted and floating aquatic plants in the
4198 Delta, such as the water hyacinth (*Eichhornia crassipes*) and the Brazilian waterweed (*Egeria*
4199 *densa*) (Reddy and Tucker 1983, Feijóo et al. 2002). Furthermore, increasing light availability
4200 due to declining turbidity levels (see above) may amplify the effects of the recent changes in
4201 nutrient concentration in the Delta. In summary, total phosphorus levels declined from the
4202 previous to the current regime, while total nitrogen and especially ammonium levels increased.
4203

4204 We hypothesize that in the pre-European period, a high nutrient supply with a low N to P ratio
4205 supported high ecological productivity and contributed substantially to the resilience of the
4206 original regime. After the Gold Rush, high turbidity levels overwhelmed nutrient effects and
4207 suppressed productivity. A long-term (at least since 1970) shift from previously higher to lower
4208 phosphorus levels and previously lower to higher total nitrogen and especially ammonium levels
4209 contributed to the long-term decline in phytoplankton and estuarine food supplies, substantial
4210 species shifts, and the recent establishment of *Microcystis* blooms in the Delta. The previous
4211 nutrient regime likely favored diatoms and other nutritious estuarine species, while the current
4212 regime favors small flagellates and nuisance species such as *Microcystis*, as well as aquatic
4213 weeds such as *Egeria densa* and *Eichhornia crassipes*.
4214

4215 7) Contaminants – While some contaminants decreased after the passage of the Clean Water Act,
4216 there has overall been an increase in amount and diversity of chemical contaminants in the
4217 system (Johnson et al. 2010, Brooks et al. in review) over the four historical eras. The effects on
4218 the biota in the system are difficult to assess due to the paucity of adequate monitoring data, but
4219 are likely as diverse as the contaminants themselves, with more severe effects on more sensitive
4220 species. Diverse sublethal effects are likely more important than acute toxicity, and interactions
4221 among different contaminants and with other variables such as temperature, turbidity, pH, and
4222 salinity are likely also important. Sublethal stress effects can also make organism more
4223 susceptible to predation and food stress.

4224
4225 We hypothesize that contaminants represent a slow variable that has slowly increased in
4226 importance as contaminants have become more abundant and diverse in the environment. They
4227 have likely played an increasing role in eroding the resilience of the original regime. The present
4228 regime increasingly favors species that are less sensitive to contaminants while presenting an
4229 increasingly greater challenge to sensitive organisms such as salmonids, osmerids, and many
4230 invertebrate species.

4231
4232 8) Harvest – Harvest represents losses of aquatic organisms. It includes the top-down box in the
4233 basic POD conceptual model, which includes effects on fishes from physical (abiotic)
4234 entrainment into water diversions as well as biotic variables such as predation and fishing.
4235 Fishing has occurred in the Delta even in pre-European times. Unlimited fishing during and after
4236 the Gold Rush along with destruction of spawning and migration habitat led to strong declines in
4237 salmonids. Recreational fishing has replaced once thriving commercial fishing and is regulated.
4238 There is overall less fishing by humans now than before, and most of the fishing is for non-native
4239 species, especially largemouth bass and striped bass. Recreational fishing for largemouth bass
4240 has dramatically increased in recent years with bass derbies now occurring year-round; however,
4241 most of these fish are released. Entrainment, especially by the CVP and SWP, has been high in
4242 some recent years with maximum salvage numbers for adult delta and longfin smelt during the
4243 POD period in 2003 and 2002, respectively (Grimaldo et al. 2009a). Salvage does not account
4244 for entrainment-associated losses that occur before fishes are collected by the fish screens, such
4245 as predation in CCF. Predation by introduced predatory fish, specifically largemouth bass, on
4246 native species has likely gone up with the more recent proliferation of these predators and
4247 increased visibility due to declining turbidity levels. Overall, harvest of native fish species has
4248 likely increased over the last 50 years or more, leading to increased top-down control of native
4249 fish species.

4250
4251 It is important to note that predation is fundamental to the structure and functioning of
4252 ecosystems. The presence or absence of predators influences energy flow in an ecosystem,
4253 carbon exchange with the atmosphere and nutrient cycling (Sabo et al. 2010). Predators often
4254 help maintain resilience to change, for example by removing weaker prey from the prey
4255 population gene pool. Individual predators can change their behavior quickly. In contrast, the
4256 number of predators and especially the presence or absence of predators in an ecosystem changes
4257 much more slowly (Sabo et al 2010). While predators were likely always an important part of the
4258 San Francisco Estuary, the recent more stable hydrology along with the decline in turbidity likely
4259 improved conditions for predators in general, but in combination with other habitat changes
4260 especially favored invasive predators.

4261
4262 We hypothesize that harvest in the form of human fishing played an important role in decreasing
4263 the resilience of the original system early on, while entrainment and predators favored by
4264 changing conditions became an important slow variable more recently.

4265
4266 *Other drivers* – As our focus shifts to understanding regime change, we will likely develop a
4267 better understanding and appreciation for the effects of slow variables. There are likely other
4268 important slow variables that have not been as thoroughly explored in the POD investigation due
4269 to its limited scope. Such variables might include slower or more distant (in time and space)

4270 system attributes. Examples may include biogeochemical and geomorphological features, such as
4271 carbon cycling or spatial distribution and connections between shallow tidal wetland areas,
4272 deeper waterways, and river floodplains. In the Delta and its upstream tributaries, these features
4273 underwent massive changes during the Gold Rush and Post-Reservoir periods, including
4274 destruction of 95% of the Delta wetlands and associated changes in biogeochemical cycling, and
4275 the many changes to the upstream tributaries including levee and dam construction. In the
4276 marine-dominated bays of the San Francisco Estuary, biota shift in response to climatic variation
4277 associated with the Pacific Decadal Oscillation and North Pacific Gyre Oscillation (Cloern et al.
4278 2010). These biological responses are likely related to fairly rapid climate-driven changes in
4279 oceanic winds and coastal currents, upwelling, and surface water temperature which affect the
4280 recruitment of marine species into the estuary. Similarly, recent steep declines in Sacramento
4281 River fall-run Chinook salmon populations have been linked to rapid changes in ocean
4282 conditions superimposed on the slower, more gradual degradation of riverine and estuarine
4283 habitat conditions (Lindley et al. 2009). Finally, as already mentioned in the preceding discussion
4284 of some of the individual drivers, there are likely many interactions among these environmental
4285 drivers that produce additive, synergistic, or compensatory effects and complex feedbacks that
4286 cross temporal and spatial scales.

4287
4288 In addition to the mostly abiotic drivers discussed above, the biological communities themselves
4289 can also act as slow drivers affecting ecological resilience. The importance of the presence or
4290 absence of (top) predators has already been mentioned above (see Harvest). The presence or
4291 absence of other functional groups is similarly important. For example, invasive benthic
4292 consumers such as the clam *C. amurensis* have been associated with strong declines in
4293 phytoplankton biomass and productivity, with substantial bottom-up repercussions throughout
4294 the food web and probably stabilizing the current POD regime in Suisun Bay and the Delta
4295 (Jassby et al. 2002, Winder and Jassby 2010). In general, functional diversity acts to stabilize a
4296 given regime, especially if each functional group consists of several species with different
4297 responses to environmental drivers (Folke et al. 2002). Ecological diversity is the biological
4298 counterpart to landscape heterogeneity. Both are key components of the adaptive capacity and
4299 resilience of ecosystems

4300

4301 **The new pelagic regime in the Sacramento-San Joaquin Delta**

4302
4303 The new pelagic regime has already been described in detail in previous chapters – it is the POD
4304 regime that is the topic of the basic POD conceptual model and the species-specific models.
4305 Briefly, it is characterized by lower outflows, a shifted and constricted salinity gradient, a
4306 simplified, rigid landscape, warmer temperatures, lower than previous turbidity, higher
4307 ammonium concentrations especially in the northern part of the system, lower phosphorus levels,
4308 higher contaminant loads and concentrations, and higher harvest including entrainment and
4309 predation.

4310
4311 Pelagic biota include an altered phytoplankton community with a smaller proportion of large
4312 diatoms and a larger proportion of small and motile species as well as recurring *M. aeruginosa*
4313 blooms in the summer and fall. Overall, phytoplankton biomass and production declined through
4314 the early 1990s and then remained fairly stable. It may now rebound to some degree, albeit with

4315 different species than before. The native zooplankton community has been largely replaced by
4316 non-native species. Copepod biomass remained relatively unchanged over the last 40 years, but
4317 now includes much greater numbers of small, evasive cyclopoids than before, especially in the
4318 western part of the Delta. Rotifers, cladocerans, and especially mysids have all strongly declined
4319 and are now at much lower biomass levels than before. Non-native jellyfish started strongly
4320 increasing in numbers in the recent decade in the western Delta and Suisun Bay and Suisun
4321 Marsh. Some invasive fish species that can use pelagic habitat but are more closely associated
4322 with nearshore habitat (e.g., Mississippi silverside) are doing well, while the POD fish species as
4323 well as salmon which use estuarine pelagic and littoral habitat during their migrations are doing
4324 very poorly.

4325
4326 Reduced turbidity gives visual predators such as adult striped bass and largemouth bass an
4327 advantage over small prey species, including introduced species such as threadfin shad or young
4328 striped bass, or native species such as delta smelt or longfin smelt or juvenile salmon. Reduced
4329 turbidity also benefits primary producers, including native, non-native, and nuisance species.
4330 Proliferating beds of the invasive aquatic weed *E. densa* are encroaching on pelagic habitat and
4331 are ideal habitat for largemouth bass and many other non-native species. Finally, the area and
4332 volume of pelagic habitat has probably been more stable in the Delta since pre-European times
4333 than benthic, littoral, and especially tidal and seasonal wetland habitat. However, the suitability
4334 (quality) of this habitat for native species has been severely degraded due to the changes in
4335 abiotic and biotic drivers described above. In some ways the open water areas of the Delta now
4336 resemble those of a eutrophic, shallow lake that is full of aquatic weeds and cyanobacteria such
4337 as Clear Lake, California.

4338

4339 **The new benthic regime in the Sacramento-San Joaquin Delta**

4340

4341 Relatively little is known about the pre-European benthic community of the San Francisco
4342 Estuary. Native freshwater mussels were likely an important component of the Delta community,
4343 but few persist in the Delta or elsewhere in California (Howard 2010). In contrast, more recent
4344 events leading to the regime shift in the benthic community of the upper estuary are well
4345 documented. These events began several decades ago with a series of invasive species
4346 introductions that would indelibly alter upper estuary benthic community structure (see Peterson
4347 and Vayssières 2010) and produce trophic consequences exhibited at all levels of the food web
4348 (e.g., bacteria – Werner and Hollibaugh 1993, Hollibaugh and Wong 1996; phytoplankton –
4349 Alpine and Cloern 1992, Jassby et al. 2002; zooplankton – Kimmerer et al. 1994, Kimmerer and
4350 Orsi 1996, Orsi and Mecum 1996; fishes – Feyrer et al. 2003, Stewart et al. 2004). Although
4351 hundreds of species have invaded the San Francisco Estuary (Cohen and Carlton 1998), perhaps
4352 the most consequential among those belonging to the benthic community are the clams
4353 *Corbicula fluminea* in freshwater and *C. amurensis* (previously *Potamocorbula amurensis*),
4354 which ranges throughout the estuary, but reaches its highest densities in intermediate salinities
4355 often found in Suisun or San Pablo bays (Peterson and Vayssières 2010). *C. fluminea* invaded
4356 the system in the mid-1940s so it has inhabited fresh and low salinity regions in the upper
4357 estuary since benthic surveys began (see Hazel and Kelly 1966, Painter 1966 and references
4358 therein). So, it is not possible to contrast the benthic community before and after its introduction
4359 or detect changes in the food web. *C. fluminea* may sequester more pelagic productivity into the

4360 benthic portion of the food web than the native freshwater mussels and other freshwater benthic
4361 inhabitants, but this is unknown. *C. fluminea* is among the dominant benthic organisms within
4362 the Delta and in Suisun Bay after consecutive high outflow years (Peterson and Vayssieres 2010)
4363 and is capable of exerting a strong negative effect on pelagic production (Lucas et al. 2002).
4364 However, the introduction of *C. amurensis* in 1986 and its rapid establishment (Nichols et al.
4365 1990, Carlton et al, 1990) led to a shift in community structure (Nichols and others 1990) that
4366 remains stable to recent years, including the POD years (Peterson and Vayssieres 2010). Since
4367 then, high rates of filter feeding by *C. amurensis* have been linked to greatly reduced
4368 phytoplankton biomass and the absence of seasonal phytoplankton blooms in the upper estuary
4369 (Alpine and Cloern 1992, Jassby et al. 2002, Jassby 2006). Decreases in zooplankton have been
4370 linked to these reductions in phytoplankton and also to direct feeding by *C. amurensis* on early
4371 life stages of zooplankton (Kimmerer et al. 1994, Kimmerer and Orsi 1996, Orsi and Mecum
4372 1996) and through zooplankton to a reduction in the abundances of fishes (Kimmerer 2002a) or a
4373 shift in distribution of others (Kimmerer 2006). By the early 2000s, both introduced clam
4374 species were well established and their effects on the upper estuary benthic community and food
4375 web had been promulgated for decades. Since Peterson and Vayssieres (2010) did not detect a
4376 change in the upper estuary benthic community in the 2000s, the contributions of these
4377 introduced clams to the regime shift are those of a major biological driver, acting at intermediate
4378 scales by diminishing the upper estuary pelagic food web. The clams themselves are affected by
4379 slower acting drivers, especially changes in salinity.

4380
4381 Overall, benthic community structure remains most strongly influenced by salinity, substrate
4382 (sediment texture) and current energy (e.g., Nichols 1979, Peterson and Vayssieres 2010).
4383 Species diversity peaks in stable marine and freshwater regions of the estuary and declines in the
4384 more variable, intermediate salinities. These intermediate salinity regions are often dominated
4385 by a few benthic organisms that support the highest benthic biomass (Peterson and Vayssieres
4386 2010). This highlights the productive nature of the estuarine mixing zone and large benefits
4387 derived by species tolerant of variable salinity conditions. *C. amurensis* appeared exceptional in
4388 regard to tolerance and adaptability. Neither salinity and nor substrate appeared to limit its
4389 colonization, but reduced competition with established species as a result of extreme
4390 environmental conditions before, during, and after 1986 may have been key to its establishment
4391 (Nichols et al. 1990). Within two years of its appearance in 1986, *Corbula* had spread
4392 throughout the estuary, inhabiting sub-tidal mud, sand, peat and clay substrates, salinities from
4393 <1 to 33 ppt, and reaching densities that exceeded 10,000 m⁻² at a few locations (Carlton et al.
4394 1990). In the Suisun Bay region, *Corbula* was able to immediately change the benthic
4395 community (Nichols et al. 1990, Peterson and Vayssieres 2010). In 1988, when low outflow
4396 conditions were expected to favor the clam *Mya arenaria* (indicator species of the intermediate
4397 salinity benthic community), *Corbula* remained dominant in Suisun Bay. Nichols et al. (1990)
4398 attributed this to alteration of benthic community dynamics by *Corbula*.

4399
4400 The benthic community shifts of most direct relevance to the POD took place in Suisun Bay and
4401 the western Delta, so this smaller region will be the focus of a more detailed examination of
4402 community changes. As mentioned previously, the only large and significant community
4403 changes occurred after the introduction of *Corbula* in 1986 and no additional changes were
4404 detected in the early 2000s (Peterson and Vayssieres 2010). Here we describe dominant benthic
4405 community members in Suisun Bay and the western Delta, contrasting those present in low and

4406 high outflow (“dry” and “wet” years for Peterson and Vayssieres 2010) as well as before and
4407 after the introduction of *Corbula*. We summarized the species abundance descriptions of
4408 Peterson and Vayssieres (2010). During low outflow years prior to *Corbula* the Suisun Bay
4409 community structure, based on Grizzly Bay sampling, strongly resembled that of San Pablo Bay
4410 and included the clams *Mya arenaria*, *Macoma petalum*, *Musculista senhousia*, and amphipods
4411 *Ampelisca abdita*, *Monocorophium acherusicum*, *Grandidierella japonica*, polychaete
4412 *Streblospio benedicti* and several other species. In high outflow years prior to *Corbula* the
4413 benthic assemblage was dominated by organisms normally found in the lower Sacramento River,
4414 including the amphipods *Americorophium stimpsoni* and *A. spinicorne*, the clam *Corbicula*
4415 *fluminea* and mermithid nematodes. During the years since its invasion *C. amurensis* has been
4416 numerically dominant under all outflow types. It is joined in low outflow years by *Nippoleucon*
4417 *hinumensis*, *A. abdita*, *G. japonica*, and a group of polychaete and oligochaete worms. In high
4418 outflow years *Corbula* is joined by *Corophium alienense*, *A. stimpsoni*, *Gammarus daiberi*, *C.*
4419 *heteroceratum* and a few polychaete and oligochaete worms. Shifting upstream into the western
4420 Delta, the benthic community in the lower Sacramento River was variable over time and suffered
4421 a distinct overall decline in organism numbers after *Corbula*. *Corbula* was present in all outflow
4422 types, but only ranked among the dominant species during low outflow years (Peterson and
4423 Vayssieres 2010). The community was dominated in most pre-*Corbula* years by the amphipods
4424 *A. stimpsoni* and *A. spinicorne*, the clam *C. fluminea* and oligochaete worms *Varichaetadrilus*
4425 *angustipenis* and *Limnodrilus hoffmeisteri*. In low outflow years, additional species joined the
4426 community: polychaete worms *Boccardiella ligerica* and *Laomome sp.* were abundant with the
4427 former numerically dominant and the amphipods *G. japonica*, *Monocorophium acherusicum* and
4428 *Melita nitida* and isopod *Synidotea laevidorsalis* were also abundant. In high outflow years after
4429 the *Corbula* invasion the amphipods *A. stimpsoni*, *A. spinicorne* and *Gammarus daiberi*,
4430 polychaetes *Laomome sp.* and *Marenzelleria viridis*, as well as oligochaete and mermithid
4431 nematode worms were present. During dry year post-invasion, *C. amurensis* and the polychaete
4432 *Boccardiella ligerica* were regularly abundant. The presence of *C. amurensis* after 1986 resulted
4433 in 2 snail species and a clam dropping out of Suisun Bay collections and sharp declines in other
4434 competing bivalves in both Suisun Bay and the western Delta (Peterson and Vayssieres 2010).
4435 Similarly, after 1986 abundance declined for most amphipod species except for *Gammarus*
4436 *daiberi*, which increased in both locations. Five additional recent invaders were consistently
4437 associated with *C. amurensis* in the upper estuary and formed a relatively stable assemblage:
4438 amphipods *Corophium alienense*, *Gammarus daiberi*, polychaetes *Marenzelleria viridis*, and
4439 *Laomome sp.* and the cumacean *Nippoleucon hinumensis*. The post-*Corbula* benthic community
4440 bears little resemblance to the previous community.

4441
4442 Although the presence of *Corbula* strongly affected the pelagic food web as described above and
4443 the change in community structure undoubtedly changed feeding opportunities, these changes
4444 also created new feeding opportunities for organisms capable of feeding at or near the bottom.
4445 *C. amurensis* was rapidly incorporated as a major portion of the diets of several species of diving
4446 birds and bottom-feeding fishes (Nichols et al. 1990). White sturgeon and adult splittail utilized
4447 *C. amurensis* as a food source, but there were consequences. The inability of *C. amurensis* to
4448 efficiently process and excrete selenium led to bioaccumulation in white sturgeon, Dungeness
4449 crab, and splittail (Stewart et al. 2004). Stewart et al. (2004) also noted that accumulation levels
4450 observed were sufficient to cause reproductive problems and developmental deformities in
4451 fishes. Other introduced organisms, particularly amphipods, provided additional links in the

4452 food web. With the severe decline of the mysids as a food source, young striped bass rapidly
4453 shifted to feeding on amphipods, particularly gammarid amphipods (probably *G. daiberi*)
4454 (Bryant and Arnold 2007). Similarly, gammarid and *Corophium* spp. amphipods now provide
4455 substantial summer and fall food for longfin smelt (S. Slater, CDFG, unpublished data). Since
4456 these organisms are closely associated with bottom during the day (Chapman 1988, Kimmerer et
4457 al. 1998), benthic-oriented feeding may partially explain the apparent shift of longfin smelt to the
4458 lower portion of the water column (see Longfin smelt section, Figure 28). Thus, to some degree
4459 even pelagic fishes have adapted somewhat to the changes in the benthic community.
4460

4461 **The new littoral regime in the Sacramento-San Joaquin Delta**

4462
4463 The regime shift of the Sacramento-San Joaquin ecosystem manifests in the littoral (shallow,
4464 nearshore) zone as a replacement of seasonally inundated tule marshes with perennial beds of
4465 SAV. While several species of aquatic macrophytes are native and still present today (coontail
4466 *Ceratophyllum demersum*, sago pondweed *Stuckenia pectinata* and *S. filiformis*, American
4467 pondweed *Potamogeton nodosus*, and common waterweed *Elodea canadensis*), the SAV
4468 community today is dominated by Brazilian waterweed *E. densa*. Other invasive aquatic
4469 macrophytes are also present (Eurasian watermilfoil *Myriophyllum spicatum*, Carolina fanwort
4470 *Cabomba caroliniana*, and curlyleaf pondweed *Potamogeton crispus*), but do not comprise
4471 nearly the proportion of the biomass as *E. densa* (roughly 85%, L. Conrad, DWR, unpublished
4472 data). Indeed, *E. densa* commonly exists in monospecific patches and is often the dominant
4473 species when it co-occurs with other macrophytes (Santos et al. 2010). The invasive floating
4474 macrophyte water hyacinth *E. crassipes* is also present and has become quite abundant; however,
4475 there is currently a very successful control program in place so water hyacinth is not currently
4476 having substantial effects in the Delta.

4477
4478 The exact date of the introduction of *E. densa* is unknown but it was established in the Delta by
4479 the early 1980s (Brown and Michniuk 2007). The plant was a nuisance to navigation throughout
4480 the Delta by the mid-1980s (CDBW 2001) and was well established in all but the northern
4481 portion of the Delta by the early 2000s. In June of each year from 2004 to 2008, the Center for
4482 Spatial Technology and Remote Sensing at UC Davis (CSTARS) used hyperspectral imagery of
4483 the entire Delta to create maps of the SAV distribution for each year. Total SAV coverage was
4484 highest in 2005 and 2006, covering approximately 2300 ha, or 7% of the water surface. More
4485 recently, this figure has declined, with SAV covering 927 ha (3% of the water surface) in 2008
4486 (E. Hestir, UCD, personal communication). This decrease in coverage is likely a result of an
4487 herbicide application program targeting *E. densa* initiated in 2001 by the California Department
4488 of Boating and Waterways.

4489
4490 The expansion of *E. densa* is directly important because it alters the characteristics of littoral
4491 habitat, but it also acts as an “ecosystem engineer” (Jones et al. 1994, Jones et al. 1997). The
4492 species is known for its ability to alter conditions of the physical habitat in other systems
4493 (Champion and Tanner 2000). SAV cover is negatively associated with turbidity in the Delta
4494 (Hestir 2010), as is *E. densa* specifically (Santos et al. 2010). In fact, the decreasing turbidity
4495 trend in the Delta between 1975 and 2008 is highly correlated with the increase in SAV cover.
4496 As discussed above, much of the increased water clarity in the Delta may be attributable to a

4497 decrease in sediment supply due to dam construction and erosion of sediment deposits from the
4498 Gold Rush era (Wright and Schoellhamer 2004). However, even after statistically controlling for
4499 reduced sediment inputs, the negative relationship between turbidity and SAV cover is still
4500 present and significant, with SAV accounting for an estimated 21–70% of the total trend (Hestir
4501 2010). These results are consistent with work from other systems showing that aquatic
4502 macrophytes attenuate water currents, thus increasing sedimentation (Yang 1998, Braskerud
4503 2001). These processes can form into a positive feedback loop in which SAV slows water
4504 velocities and increases sedimentation, which provides favorable habitat for more SAV. This
4505 positive feedback may be an important factor contributing to the current state of the Delta. The
4506 effect of *E. densa* on hydrodynamics is unknown but could be important. Hestir (2010)
4507 determined that water velocities in excess of 0.49 m/s limited SAV establishment (Hestir 2010).
4508 In some situations *E. densa* can likely completely block channels and in others an equilibrium
4509 might be reached with a deep channel with high water velocities between beds of SAV.

4510
4511 Beyond effects on water velocity, turbidity, and sedimentation, SAV can alter DO concentrations
4512 and temperature gradients. During summertime, when at its densest and tallest, *E. densa* can
4513 cause great swings in DO due to high rates of photosynthesis in beds during the day and high
4514 rates of respiration during the night (Wilcox et al. 1999). However, it is not clear to what extent
4515 *E. densa* may alter average levels of DO, and at what spatial scale. Water temperatures may be
4516 higher in dense beds of SAV because the plants can absorb heat during the day and block heat
4517 loss at night, causing vertical stratification of temperature in the water column (Grimaldo and
4518 Hymanson 1999, Wilcox et al. 1999). If water exchange between SAV beds and nearby open
4519 water areas is limited, increased densities of SAV may also cause horizontal temperature
4520 gradients (Stacey 2003).

4521
4522 In addition to altering the physical aspect of the environment, the expansion of *E. densa* also has
4523 a profound effect on the littoral zone biological community, at multiple trophic levels. It
4524 influences the general aquatic macrophyte community by preventing some species from
4525 becoming established and facilitating the persistence of other species (Santos et al. 2010). Its
4526 widespread dominance in the Delta is a result of its bimodal growth pattern, with peaks in the
4527 late spring and fall. Growth in the fall allows a greater proportion of its biomass to persist
4528 through the winter, giving it a competitive advantage over most other macrophytes at the start of
4529 the growth season the following year. An exception is the native coontail (*C. demersum*), which
4530 does not root in the substrate and instead relies on external structures (such as other plants) for
4531 anchoring. Thus, coontail benefits from ample anchoring structure provided by *E. densa* and the
4532 two species commonly co-occur. Co-existence of *E. densa* with other aquatic macrophytes
4533 species is relatively uncommon (Santos et al. 2010).

4534
4535 The widespread establishment of *E. densa* led to a dramatic shift in the resident fish community.
4536 In particular, largemouth bass and other centrarchid fishes are now prevalent. Historically, the
4537 littoral fish community in the Delta was probably a blend of fresh and brackish water fishes
4538 whose distributions varied with a dynamic salinity gradient. Freshwater fishes included a suite of
4539 cyprinids such as hitch *Lavinia exilicauda*, thicktail chub, Sacramento blackfish *Orthodon*
4540 *microlepidotus*, along with Sacramento sucker *Catostomus occidentalis* and Sacramento
4541 pikeminnow. Species with broad salinity tolerances were also present, including splittail, tule
4542 perch *Hysterocarpus traski*, prickly sculpin *Cottus asper*, threespine stickleback *Gasterosteus*

4543 *aculeatus*, and Sacramento perch, the only native centrarchid (Moyle 2002). Predators included
4544 Sacramento pikeminnow, thickettail chub, and Sacramento perch. These species were all strongly
4545 associated with littoral vegetated habitats and were likely also found in more open waters away
4546 from the littoral zone. By the early 1980s, centrarchids comprised 33% and catfish 17% of the
4547 total catch in Delta-wide electrofishing surveys conducted by CDFG, compared to approximately
4548 12% native species (Brown and Michniuk 2007). In 2001–2003, centrarchids dominated CDFG
4549 electrofishing catches and contributed 61% of the catch with only 3% native fish (Brown and
4550 Michniuk 2007). Recent data from a two-year electrofishing effort conducted at UC Davis
4551 between 2008 and 2010 confirmed the centrarchids as unquestionably dominant, comprising
4552 approximately 65% of all fish sampled, with native residents accounting for only 2% of the
4553 catch. Interestingly, catfish populations have apparently declined, also only comprising 2% of
4554 sampled fishes (L. Conrad, DWR, unpublished data).

4555
4556 In 2008–2010, largemouth bass make up approximately 35% of the centrarchids residing in the
4557 littoral zone (L. Conrad, DWR, unpublished data). From the perspective of the POD, the rise of
4558 largemouth bass is of concern because the species is a voracious predator that has been known to
4559 cause dramatic shifts in community composition in other systems (Schindler et al. 1997). Along
4560 with other centrarchids, they are positively associated with SAV and likely benefit from clearer
4561 waters and warmer temperatures that characterize the present-day littoral ecosystem. However,
4562 largemouth bass and other centrarchids may also profit from SAV in an energetic sense.
4563 Decreased water velocities, warmer temperatures, and enhanced nutrient cycling allow
4564 periphyton to grow on plant stems and leaves, in turn supporting an enriched community of
4565 epibenthic prey (e.g., aquatic insects, amphipods, snails) that are consumed by juvenile fishes.
4566 Adult largemouth bass can then prey on juvenile fishes residing in the nearshore, though SAV
4567 probably offers juveniles a predator refuge.

4568
4569 A recent comparison of stable isotope signatures between pelagic and nearshore fishes support
4570 this picture of a self-contained food-web in the littoral zone. While pelagic fish were dependent
4571 on a zooplankton-phytoplankton trophic pathway, littoral fishes had carbon isotope ratios
4572 consistent with those of SAV and epiphytic macroalgae (Grimaldo et al. 2009b). Analysis of diet
4573 samples collected from largemouth bass throughout the Delta are consistent with this result: the
4574 most common fish species identified in largemouth bass stomach contents are other centrarchids,
4575 including juvenile largemouth bass, followed by prickly sculpin, and shimofuri and yellowfin
4576 gobies. In contrast, pelagic fishes are rare in the diet of largemouth bass, with only 12 threadfin
4577 shad and 1 striped bass found over 1500 samples analyzed (L. Conrad, DWR, unpublished data).
4578 After fish prey, adult largemouth bass appear to rely heavily on the red swamp crayfish,
4579 *Procambrus clarkii* (Nobriga and Feyrer 2007). These crayfish are also non-native and appear to
4580 be prevalent in beds of *E. densa*, but to date their relative abundance across nearshore habitat
4581 types has not been quantified. The nature of feedbacks between *E. densa*, largemouth bass,
4582 crayfish, and other residents of the nearshore are not yet fully described for the current littoral
4583 ecosystem. Studies of the role of *P. clarkii* in the littoral food web and nutrient cycling are
4584 warranted, given that this species has previously been identified as an ecosystem engineer and is
4585 associated with catastrophic regime shifts, like *E. densa* (Matsuzaki et al. 2009).

4586
4587 The segregation of littoral and pelagic food webs suggests the separation of whole ecosystem
4588 processes between the two habitats. Formation of dense beds of *E. densa* may result in a

4589 “biological barrier” for fish and other biota between nearshore open waters and tidal wetlands
4590 (Brown 2003) or along migration routes. These barriers may greatly diminish the benefits of
4591 natural and restored tidal wetlands (Brown 2003) and upstream spawning areas for native fishes.
4592

4593 **2010 POD Program**

4594
4595 The POD work for 2010 includes 39 continuing study elements and 32 new elements. Brief
4596 descriptions of each element are included in Appendix 1. The coordinated POD study elements
4597 address a number of the questions and issues described in the body of this report. The linkages
4598 of each study element with the POD conceptual models are listed in Table 1, including
4599 information on targeted seasons for each element.

4600
4601 The funding support for the 2010 POD program from DWR, USBR, and SWRCB is
4602 approximately \$7,547,000 (Table 2). In spite of an increase in allocated funding from these
4603 sources of only 1% relative to 2009, the 2010 POD program has 24 more study elements than
4604 conducted in 2009, bringing the total number of POD studies to 76 during 2010. The increased
4605 number of studies is largely a result of a proposal solicitation (see
4606 <http://www.water.ca.gov/iep/activities/research.cfm>) which resulted in 14 new solicited elements
4607 and 15 new directed elements that are broadly distributed across the POD conceptual model topic
4608 areas. Several projects funded by research grants from the Delta (previously CALFED) Science
4609 Program that are thematically related to the POD efforts are also part of the coordinated POD
4610 program and listed and identified in Appendix 1 and Table 2. 2010 funding for these projects is
4611 estimated at \$3,019,000. In some instances, funding for 2010 work was obligated with 2009
4612 funds, therefore not reflected in the table below.

4613

4614 **References Cited**

4615

4616 **Peer-reviewed literature, final reports and IEP Newsletter articles:**

4617

- 4618 Aasen, G. 2010. Fish salvage at the State Water Project's and Central Valley Project's fish
4619 facilities during 2009. IEP Newsletter 23(2):72–79.
- 4620 Alpine, A.E. and J.E. Cloern. 1992. Trophic interactions and direct physical effects control
4621 phytoplankton biomass and production in an estuary. *Limnology and Oceanography* 37:946–
4622 955.
- 4623 Allee, W.C. 1931. *Animal Aggregations. A study in General Sociology.* University of Chicago
4624 Press, Chicago.
- 4625 Amweg, E.L., D.P. Weston, and N.M. Ureda. 2005. Use and toxicity of pyrethroid pesticides in
4626 the Central Valley, California, USA. *Environmental Toxicology and Chemistry* 24:966–972.
- 4627 Anderson, L. and M. C. Hoshovsky. 2000. *Egeria densa.* in C.C. Bossard, , J.M. Randall, and M.
4628 C. Hoshovsky, editors. *Invasive Plants of California's Wildlands.* University of California
4629 Press, Berkeley, CA.
- 4630 Anderson, S.L., G. Cherr, D. Nacci, D. Schlenk, and E. Gallagher. 2007. Biomarkers and the
4631 pelagic organism decline: conclusions of the POD Biomarker Task Force, San Francisco,
4632 California. Report to POD Management Team. 84 pp. plus appendices

- 4633 Arnold, J.D. and H.S. Yue. 1997. Prevalence, relative abundance, and mean intensity of
4634 Pleurocercoids of *Proteocephalus* sp. in young striped bass in the Sacramento-San Joaquin
4635 Estuary. *California Fish and Game* 83(3):105–117.
- 4636 Arthur, J.F., M.D. Ball, and S.Y. Baughman. 1996. Summary of federal and state water project
4637 environmental impacts in the San Francisco Bay-Delta Estuary, California. Pages 445–495
4638 in J.T. Hollibaugh, editor. *San Francisco Bay: the ecosystem*. Pacific Division, American
4639 Association for the Advancement of Science, San Francisco, California.
- 4640 Atwater, B.F. 1979. Ancient processes at the site of southern San Francisco Bay, movement of
4641 the crust and changes in sea level. Pages 31–45 in T.J. Conomos, editor. *San Francisco Bay-*
4642 *The Urbanized Estuary: American Association for the Advancement of Science, Pacific*
4643 *Division, San Francisco*.
- 4644 Bailey, H.C., C. Alexander, C. DiGiorgio, M. Miller, S. Doroshov and D. Hinton. 1994. The
4645 effect of agricultural discharge on striped bass in California's Sacramento-San Joaquin
4646 drainage. *Ecotoxicology* 3:123–142.
- 4647 Ball M.D. and J.F. Arthur. 1979. Planktonic chlorophyll dynamics in the northern San Francisco
4648 Bay and Delta. Pages 265–286 in T.J. Conomos, editor. *San Francisco Bay: the urbanized*
4649 *estuary, Investigations into the Natural History of San Francisco Bay and Delta with*
4650 *reference to the influence of Man. Pacific Division, American Association for the*
4651 *Advancement of Science, San Francisco, California*.
- 4652 Baskerville-Bridges B., J.C. Lindberg, and S.I. Doroshov. 2002. The effect of light intensity, alga
4653 concentration, and prey density on the feeding behavior of delta smelt larvae. Pages 219–228
4654 in F. Freyer, L.R. Brown, R.L. Brown, and J.J. Orsi, editors. *Early life history of fishes in the*
4655 *San Francisco Estuary and watershed. American Fisheries Society Symposium 39. American*
4656 *Fisheries Society*.
- 4657 Braskerud, B.C. 2001. The influence of vegetation on sedimentation and resuspension of soil
4658 particles in small constructed wetlands. *Journal of Environmental Quality* 30:1447–1457.
- 4659 Baxa, D.V., T. Kurobe, K.A. Ger, P.W. Lehman and S.J. Teh. 2010. Estimating the abundance of
4660 toxic *Microcystis* in the San Francisco Estuary using quantitative real-time PCR. *Harmful*
4661 *Algae* 9:342–349.
- 4662 Baxter, R.D. 1999. Osmeridae. Pages 179–216 in J. Orsi, editor. *Report on the 1980–1995 fish,*
4663 *shrimp and crab sampling in the San Francisco Estuary. Interagency Ecological Program for*
4664 *the Sacramento-San Joaquin Estuary*.
- 4665 Baxter, R., R. Breuer, L. Brown, M. Chotkowski, F. Feyrer, M. Gingras, B. Herbold, A. Mueller-
4666 Solger, M. Nobriga, T. Sommer, and K. Souza. 2008. Pelagic organism decline progress
4667 report: 2007 synthesis of results. *Interagency Ecological Program for the San Francisco*
4668 *Estuary, Technical Report 227*.
- 4669 Beggel S., I. Werner, R.E. Connon, and J. Geist. 2010. Sublethal toxicity of commercial
4670 insecticide formulations and their active ingredients to larval fathead minnow (*Pimephales*
4671 *promelas*). *Science of the Total Environment* 408:3169–3175.
- 4672 Beisner, B.E., D.T. Haydon, and K. Cuddington. 2003. Alternative stable states in ecology.
4673 *Frontiers in Ecology and the Environment* 1:376–82.
- 4674 Bennett, W.A. 2005. Critical assessment of the delta smelt population in the San Francisco
4675 estuary, California. *San Francisco Estuary and Watershed Science*. Vol. 3, Issue 2, Article 1.
4676 <http://repositories.cdlib.org/jmie/sfews/vol3/iss2/art1>
- 4677 Bennett, W.A., J.A. Hobbs, and S.J. Teh. 2008. Interplay of environmental forcing and growth-
4678 selective mortality in the poor year-class success of delta smelt in 2005. Final report: “fish

4679 otolith and condition study 2005". Prepared for the POD Management Team of the
4680 Interagency Ecological Program for the San Francisco Estuary.

4681 Bennett, W.A. and E. Howard. 1997. El Niños and the decline of striped bass. IEP Newsletter
4682 10(4):17–21.

4683 Bennett, W.A. and E. Howard. 1999. Climate change and the decline of striped bass.
4684 IEP Newsletter 12(2):53–56.

4685 Bennett, W.A., W.J. Kimmerer, and J.R. Burau. 2002. Plasticity in vertical migration by native
4686 and exotic estuarine fishes in a dynamic low-salinity zone. *Limnology and Oceanography*
4687 47:1496–1507.

4688 Bennett, W.A. and P.B. Moyle. 1996. Where have all the fishes gone? Interactive factors
4689 producing fish declines in the Sacramento San Joaquin Estuary. Pages 519–542 in J.T.
4690 Hollibaugh, editor, *San Francisco Bay: The Ecosystem*. American Association for the
4691 Advancement of Science, San Francisco.

4692 Bennett, W.A., D.J. Ostrach, and D.E. Hinton. 1995. Larval striped bass condition in a drought-
4693 stricken estuary: evaluating pelagic food web limitation. *Ecological Applications* 5:680–692.

4694 Berec, L., E. Angulo, and F. Courchamp. 2006. Multiple Allee effects and population
4695 management. *Trends in Ecology and Evolution* 22:185–191.

4696 Bergamaschi, B.A., K.M. Kuivila, and M.S. Fram. 2001. Pesticides associated with suspended
4697 sediments entering San Francisco Bay following the first major storm of water year 1996.
4698 *Estuaries* 24:368–380.

4699 Biggs, R., S.R. Carpenter, and W.A. Brock. 2009. Turning back from the brink: Detecting an
4700 impending regime shift in time to avert it. *Proceedings of the National Academy of Sciences*
4701 U.S.A. 106:826–831.

4702 Blommer, G.L. and A.W. Gustaveson. 2002. Life history and population dynamics of threadfin
4703 shad in Lake Powell 1968–2001. Utah Division of Wildlife Resources, Project F-46-R,
4704 Publication Number 02-23, Salt Lake City, Utah.

4705 Bouley, P. and W.J. Kimmerer. 2006. Ecology of a highly abundant, introduced cyclopoid
4706 copepod in a temperate estuary. *Marine Ecology Progress Series* 324:219–228.

4707 Brander, S.M., I. Werner, J.W. White, and L.A. Deanovic. 2009. Toxicity of a dissolved
4708 pyrethroid mixture to *Hyalella azteca* at environmentally relevant concentrations.
4709 *Environmental Toxicology and Chemistry* 28:1493–1499.

4710 Braskerud, B.C. 2001. The influence of vegetation on sedimentation and resuspension of soil
4711 particles in small constructed wetlands. *Journal of Environmental Quality* 30:1447–1457.

4712 Breitburg, D.L., B.C. Crump, J.O. Dabiri, and C.L. Gallegos. 2010. Ecosystem engineers in the
4713 pelagic realm: alteration of habitat by species ranging from microbes to jellyfish. *Integrative*
4714 *and Comparative Biology*. Advance Access published June 16, 2010. doi:10.1093/icb/icq051

4715 Brett M.T. and D.C. Müller-Navarra. 1997. The role of highly unsaturated fatty acids in aquatic
4716 food web processes. *Freshwater Biology* 38:483–499.

4717 Brown, L.R. 2003. Will tidal wetland restoration enhance populations of native fishes? in L.R.
4718 Brown, editor. *Issues in San Francisco Estuary Tidal Wetlands Restoration*. San Francisco
4719 Estuary and Watershed Science Vol. 1, Issue 1, Article 2.
4720 <http://repositories.cdlib.org/jmie/sfews/vol1/iss1/art2>.

4721 Brown, L.R. and M.L. Bauer. 2009. Effects of hydrologic infrastructure on flow regimes of
4722 California's Central Valley rivers: implications for fish populations. *River Research and*
4723 *Applications* 26:751–765.

- 4724 Brown, L.R., W. Kimmerer, and R. Brown. 2009. Managing water to protect fish: a review of
4725 California's environmental water account. *Environmental Management* 43:357–368.
- 4726 Brown, L. R. and D. Michniuk 2007. Littoral fish assemblages of the alien-dominated
4727 Sacramento–San Joaquin Delta, California 1980–1983 and 2001–2003. *Estuaries and Coasts*
4728 30:186–200.
- 4729 Brown, L.R. and P.B. Moyle. 1991. Changes in habitat and microhabitat use of an assemblage of
4730 stream fishes in response to predation by Sacramento squawfish (*Ptychocheilus grandis*).
4731 *Canadian Journal of Fisheries and Aquatic Sciences* 48:849–856.
- 4732 Brown, L.R. and P.B. Moyle. 2005. Native fish communities of the Sacramento-San Joaquin
4733 watershed, California: a history of decline. Pages 75–98 in F. Rinne, R. Hughes, and R.
4734 Calamusso, editors. *Fish Communities of Large Rivers of the United States*. American
4735 Fisheries Society, Bethesda, Maryland.
- 4736 Brown, R., S. Greene, P. Coulston and S. Barrow. 1996. An evaluation of the effectiveness of
4737 fish salvage operations at the intake to the California aqueduct, 1979–1993. Pages 497–518
4738 in J. T. Hollibaugh, editor. *San Francisco Bay: the ecosystem*. Pacific Division of the
4739 American Association for the Advancement of Science, San Francisco, CA.
- 4740 Brown, T. 2009. Phytoplankton community composition: the rise of the flagellates. *IEP*
4741 *Newsletter* 22(3):20–28.
- 4742 Bryant, M.E. and J.D. Arnold. 2007. Diets of age-0 striped bass in the San Francisco Estuary,
4743 1973–2002. *California Fish and Game* 93(1):1–22.
- 4744 Burkhardt-Holm, P., and K. Scheurer. 2007. Application of the weight-of-evidence approach to
4745 assess the decline of brown trout (*Salmo trutta*) in Swiss rivers. *Aquatic Sciences* 69:51–70.
- 4746 Byrne, R, B.L. Ingram, S. Starratt, F. Malamud-Roam, J.N. Collins, and M.E. Conrad. 2001.
4747 Carbon-isotope, diatom, and pollen evidence for Late Holocene salinity change in a brackish
4748 marsh in the San Francisco Estuary. *Quaternary Research* 55:66–76.
- 4749 California Department of Boating and Waterways (CDBW). 2001. Environmental impact report
4750 for the *Egeria densa* control program, Chapter 1.
- 4751 California Department of Fish and Game (CDFG). 2009a. Report to the Fish and Game
4752 Commission: A status review of the longfin smelt (*Spirinchus thaleichthys*) in California. 46
4753 pp.
- 4754 California Department of Fish and Game (CDFG). 2009b. Effects analysis – State Water Project
4755 effects on longfin smelt. 58 pp. plus appendices
- 4756 Carlton, J.T. 1979. Introduced invertebrates of San Francisco Bay. Pages 427–444 in T. J.
4757 Conomos, editor. *San Francisco Bay: The Urbanized Estuary, Investigations into the Natural*
4758 *History of San Francisco Bay and Delta with reference to the influence of Man*. Pacific
4759 Division, American Association for the Advancement of Science, San Francisco.
- 4760 Carpenter, S.R. and M.G. Turner. 2000. Hares and tortoises: interactions of fast and slow
4761 variables in ecosystems. *Ecosystems* 3:495–497.
- 4762 Carpenter, S.R., J.J. Cole, J.R. Hodgson, J.F. Kitchell, M.L. Pace, D. Bade, K.L. Cottingham,
4763 T.E. Essington, J.N. Houser, and D.E. Schindler. 2001. Trophic cascades, nutrients, and lake
4764 productivity: whole-lake experiments. *Ecological Monographs* 71:163–186.
- 4765 Carlton, J.T., J.K. Thompson, L.E. Schemel, and F.H. Nichols. 1990. Remarkable invasion of
4766 San Francisco Bay (California, USA) by the Asian clam *Potamocorbula amurensis* I.
4767 Introduction and dispersal. *Marine Ecology Progress Series* 66:81–94.

- 4768 Cavallo, B., J. Merz, C. Turner, and P. Bergman. 2009. Review of delta smelt and longfin smelt
4769 monitoring program at Contra Costa and Pittsburg Power Plants. Technical Report to Mirant
4770 Corporation, Cramer Fish Sciences, Auburn, CA.
- 4771 Central Valley Regional Water Quality Control Board (CVRWQCB). 2010. NPDES Permit
4772 Renewal Issues Aquatic Life and Wildlife Preservation Sacramento Regional County
4773 Sanitation District Sacramento Regional Wastewater Treatment Plant. Staff report available
4774 at
4775 http://www.waterboards.ca.gov/centralvalley/board_decisions/tentative_orders/aquatictox/aquatictox_iss_pap.pdf
4776
- 4777 Champion, P.D. and Tanner, C.C. 2000. Seasonality of macrophytes and interaction with flow in
4778 a New Zealand lowland stream. *Hydrobiologia* 441:1–12.
- 4779 Chapin, F.S., III, G.P. Kofinas, and C. Folke, editors. 2009. Principles of ecosystem stewardship:
4780 Resilience-based natural resource management in a changing world. Springer, New York.
- 4781 Chapin, F.S., III, A.L. Lovcraft, E.S. Zavaleta, J. Nelson, M.D. Robards, G.P. Kofinas, S.F.
4782 Trainor, G. Peterson, H.P. Huntington, and R.L. Naylor. 2006. Policy strategies to address
4783 sustainability of Alaskan boreal forests in response to a directionally changing climate.
4784 *Proceedings of the National Academy of Sciences* doi:10.1073/pnas.0606955103.
- 4785 Chigbu, P. 2000. Population biology of longfin smelt and aspects of the ecology of other major
4786 planktivorous fishes in Lake Washington. *Journal of Freshwater Ecology* 15:543–558.
- 4787 Clark, K.W., M.D. Bowen, R.B. Mayfield, K.P. Zehfuss, J.D. Taplin, and C.H. Hanson. 2009.
4788 Quantification of Pre-Screen Loss of Juvenile Steelhead in Clifton Court Forebay. State of
4789 California. The California Natural Resources Agency. Department of Water Resources.
4790 Fishery Improvements Section Bay-Delta Office. 119 pp.
- 4791 Cloern, J.E. 2001. Our evolving conceptual model of the coastal eutrophication problem. *Marine*
4792 *Ecology Progress Series* 210:223–253.
- 4793 Cloern, J.E. 2007. Habitat connectivity and ecosystem productivity: Implications from a simple
4794 model. *American Naturalist* 169:E21–E33.
- 4795 Cloern, J.E. and A.D. Jassby. 2008. Complex seasonal patterns of primary producers at the land–
4796 sea interface. *Ecology Letters* 11:1294–1303.
- 4797 Cloern, J.E. K.A. Hieb, T. Jacobson, B. Sansó, E. Di Lorenzo, M.T. Stacey, J.L. Largier, W.
4798 Meiring, W.T. Peterson, T.M. Powell, M. Winder, and A.D. Jassby. 2010. Biological
4799 communities in San Francisco Bay track large-scale climate forcing over the North Pacific.
4800 *Geophysical Research Letters* 37:L21602.
- 4801 Cohen, A.N., and J.T. Carlton. 1998. Accelerating invasion rate in a highly invaded estuary.
4802 *Science* 279:555–558
- 4803 Connon, R.E., J. Geist, J. Pfeiff, A.S. Loguinov, L.S. D’Abronzio, H. Wintz, C.D. Vulpe, and I.
4804 Werner. 2009. Linking mechanistic and behavioral responses to sublethal esfenvalerate
4805 exposure in the endangered delta smelt; *Hypomesus transpacificus* (Fam. Osmeridae). *BMC*
4806 *Genomics* 10:608.
- 4807 Contra Costa Water District. 2010. Historical fresh water and salinity conditions in the western
4808 Sacramento-San Joaquin Delta and Suisun Bay, a summary of historical reviews, reports,
4809 analyses and measurements. Water Resources Department, Contra Costa Water District,
4810 Concord, California. Technical Memorandum WR10-001.
- 4811 Courchamp, F., T. Clutton-Brock, and B. Grenfell. 1999. Inverse density dependence and the
4812 Allee effect. *Trends in Ecology and Evolution* 14:405–410.

- 4813 Crain, P. and P.B. Moyle. In press. Biology, history, status and conservation of Sacramento
4814 perch, *Archoplites interruptus*. San Francisco Estuary and Watershed Science.
4815 Crutzen, P.J. 2002. Geology of mankind. *Nature* 415:23.
4816 Daskalov, G.M., A.N. Grishin, S. Rodionov, and V. Mihneva. 2007. Trophic cascades triggered
4817 by overfishing reveal possible mechanisms of ecosystem regime shifts. *Proceedings of the*
4818 *National Academy of Sciences* 104:10518–10523.
4819 Davis, J., L. Sim, and J. Chambers. 2010. Multiple stressors and regime shifts in shallow aquatic
4820 ecosystems in antipodean landscapes. *Freshwater Biology* 55 (Suppl. 1):5–18.
4821 Davis, J.A., D. Yee, J.N. Collins, S.E. Schwarzbach, and S.N. Luoma. 2003. Potential for
4822 increased mercury accumulation in the estuary food web. *San Francisco Estuary and*
4823 *Watershed Science*. Vol. 1 Issue 1 Article 4.
4824 Dege, M. and L.R. Brown. 2004. Effect of outflow on spring and summertime distribution and
4825 abundance of larval and juvenile fishes in the upper San Francisco Estuary. Pages 49–66 in
4826 F. Feyrer, L.R. Brown, R.L. Brown, and J.J. Orsi, editors. *Early life history of fishes in the*
4827 *San Francisco Estuary and watershed*. American Fisheries Society Symposium 39.
4828 Deng, D., K. Zheng, F. Teh, P.W. Lehman, and S.J. Teh. 2010. Toxic threshold of dietary
4829 microcystin (-LR) for Quart Medaka. *Toxicol* 55:787–794.
4830 Dennis, B. 1989. Allee effects: population growth, critical density, and the chance of extinction.
4831 *Natural Resource Modeling* 3:481–538.
4832 Dettinger, M.D. 2005. From climate-change spaghetti to climate-change distributions for 21st
4833 Century California. *San Francisco Estuary and Watershed Science* Vol. 3 Issue 1 Article 4.
4834 <http://repositories.cdlib.org/jmie/sfews/vol3/iss1/art4>.
4835 Dill, W.A. and A.J. Cordone. 1997. History And status of introduced fishes in California, 1871 –
4836 1996. *Fish Bulletin* 178. California Department of Fish and Game.
4837 Dryfoos, R.L. 1965. The life history and ecology of the longfin smelt in Lake Washington. Ph.D.
4838 Dissertation. University of Washington, Tacoma WA.
4839 Dugdale, R.C., F.P. Wilkerson, V.E. Hogue, and A. Marchi. 2007. The role of ammonium and
4840 nitrate in spring bloom development in San Francisco Bay. *Estuarine, Coastal and Shelf*
4841 *Science* 73:17–29.
4842 Edmunds, J.L., K.M. Kuivila, B.E. Cole, and J.E. Cloern. 1999. Do herbicides impair
4843 phytoplankton primary production in the Sacramento - San Joaquin River Delta? Pages 81–
4844 88 in D.W. Morganwalp and H.T. Buxton, editors. *U.S. Geological Survey Toxic Substances*
4845 *Hydrology Program - Proceedings of the Technical Meeting, Charleston, South Carolina,*
4846 *March 8–12, 1999, v. 2 - Contamination of Hydrologic Systems and Related Ecosystems:*
4847 *U.S. Geological Survey Water-Resources Investigations Report 99-4018B*.
4848 Enright, C. and S.D. Culbertson. 2010. Salinity trends, variability, and control in the northern
4849 reach of the San Francisco Estuary. *San Francisco Estuary and Watershed Science*, 7(2).
4850 EPA (U.S. Environmental Protection Agency). 2009. Draft 2009 update aquatic life ambient
4851 water quality criteria for ammonia – freshwater. U.S. Environmental Protection Agency,
4852 Office of Water, Office of Science and Technology, Washington, DC.
4853 Feijoó, C., M.E. Garcia, F. Momo, and J. Toja. 2002. Nutrient absorption by the submerged
4854 macrophyte *Egeria densa* Planch.: Effect of ammonium and phosphorous availability in the
4855 water column on growth and nutrient uptake. *Limnetica* 21(1–2):93–104.
4856 Feyrer, F. 2004. Ecological segregation of native and alien larval fish assemblages in the
4857 southern Sacramento-San Joaquin Delta. Pages 67–80 in F. Feyrer, L.R. Brown, R.L. Brown,

4858 and J.J. Orsi, editors. Early Life History of Fishes in the San Francisco Estuary and
4859 Watershed. American Fisheries Society, Symposium 39, Bethesda, Maryland.

4860 Feyrer, F. and M.P. Healey. 2003. Fish community structure and environmental correlates in the
4861 highly altered southern Sacramento-San Joaquin Delta Environmental Biology of Fishes
4862 66:123–132.

4863 Feyrer, F., B. Herbold, S.A. Matern, and P.B. Moyle. 2003. Dietary shifts in a stressed fish
4864 assemblage: Consequences of a bivalve invasion in the San Francisco Estuary.
4865 Environmental Biology of Fishes 67:277–288.

4866 Feyrer, F., K. Newman, M. Nobriga and T. Sommer. 2010. Modeling the effects of future
4867 outflow on the abiotic habitat of an imperiled estuarine fish. Estuaries and Coasts. DOI:
4868 10.1007/s12237-010-9343-9.

4869 Feyrer, F., M. Nobriga, and T. Sommer. 2007. Multi-decadal trends for three declining fish
4870 species: habitat patterns and mechanisms in the San Francisco Estuary, California, U.S.A.
4871 Canadian Journal of Fisheries and Aquatic Sciences 64:723–734.

4872 Feyrer, F., T. Sommer, and S.B. Slater. 2009. Old school vs. new school: status of threadfin shad
4873 (*Dorosoma petenense*) five decades after its introduction to the Sacramento-San Joaquin
4874 Delta. San Francisco Estuary and Watershed Science Vol. 7 Issue 1. Retrieved from:
4875 <http://escholarship.org/uc/item/4dt6p4bv>

4876 Fish, M., D. Contreras, V. Afentoulis, J. Messineo, and K. Hieb. 2009. 2008 Fishes annual status
4877 and trends report for the San Francisco Estuary. IEP Newsletter 22(2):17–36.

4878 Folke, C., S. Carpenter, T. Elmqvist, L. Gunderson, C. S. Holling, and B. Walker. 2002.
4879 Resilience and sustainable development: building adaptive capacity in a world of
4880 transformations. *Ambio* 31:437–440.

4881 Folke, C., S. Carpenter, B. Walker, M. Scheffer, T. Elmqvist, L. Gunderson, and C.S. Holling.
4882 2004. Regime shifts, resilience, and biodiversity in ecosystem management. *Annual Reviews*
4883 *in Ecology Evolution and Systematics* 35:557–581.

4884 Foott, J. S., K. True, and R. Stone. 2006. Histological evaluation and viral survey of juvenile
4885 longfin smelt (*Spirinchus thaleichthys*) and threadfin shad (*Dorosoma petenense*) collected in
4886 the Sacramento-San Joaquin River Delta, April–October 2006. California Nevada Fish
4887 Health Center, U.S. Fish and Wildlife Service, Anderson, California. 11 pp.

4888 Foott, J. S. and R. Stone. 2008. Histological evaluation and viral survey of juvenile longfin smelt
4889 (*Spirinchus thaleichthys*) and threadfin shad (*Dorosoma petenense*) collected from the
4890 Sacramento-San Joaquin River Delta, April–November 2007. California Nevada Fish Health
4891 Center U.S. Fish and Wildlife Service, Anderson, California. 16 pp.

4892 Foss, S.F. and L.W. Miller. 2004. Growth and growth rate variability of larval striped bass in the
4893 San Francisco Estuary. Pages 203–217 in F. Feyrer, L.R. Brown, R.L. Brown, and J.J. Orsi,
4894 editors. Early Life History of Fishes in the San Francisco Estuary and Watershed. American
4895 Fisheries Society, Symposium 39, Bethesda, Maryland.

4896 Frank, K.T., B. Petrie, J. S. Choi, W.C. Leggett. 2005. Trophic cascades in a formerly cod-
4897 dominated ecosystem. *Science* 308:1621–1623.

4898 Fulton, R.S. and H.W. Paerl. 1987. Toxic and inhibitory effects of the blue-green alga
4899 *Microcystis aeruginosa* on herbivorous zooplankton. *Journal of Plankton Research* 9:837–
4900 855.

4901 Gartz, R. and S. Vu. 2006. POD condition and diet: 2005 3.c. Report to POD management team.

4902 Gascoigne, J., L. Berec, S. Gregory, and F. Courchamp. 2009. Dangerously few liaisons: a
4903 review of mate-finding Allee effects. *Population Ecology* 51:355–372.

- 4904 Gascoigne, J.C. and R.N. Lipcius. 2004. Allee effects driven by predation. *Journal of Applied*
4905 *Ecology* 41:801–810.
- 4906 Geist, J., I. Werner, K.J. Eder, and C.M. Leutenegger. 2007. Comparisons of tissue-specific
4907 transcription of stress response genes with whole animal endpoints of adverse effect in
4908 striped bass (*Morone saxatilis*) following treatment with copper and esfenvalerate. *Aquatic*
4909 *Toxicology* 85:28–39.
- 4910 Ger, K.A., P. Arneson, C.R. Goldman, and S.J. Teh. 2010b. Species specific differences in the
4911 ingestion of *Microcystis* cells by the calanoid copepods *Eurytemora affinis* and
4912 *Pseudodiaptomus forbesi*. *Journal of Plankton Research* 32:1479–1484.
- 4913 Ger, K.A. S.J. Teh, D.V. Baxa, S. Lesmeister, and C.R. Goldman. 2010a. The effects of dietary
4914 *Microcystis aeruginosa* and microcystin on the copepods of the upper San Francisco Estuary.
4915 *Freshwater Biology* 55:1548–1559.
- 4916 Ger, Kemal A., Swee J. Teh and Charles R. Goldman. 2009. Microcystin-LR toxicity on
4917 dominant copepods *Eurytemora affinis* and *Pseudodiaptomus forbesi* of the upper San
4918 Francisco Estuary. *Science of the Total Environment* 407:4852–4857.
- 4919 Giddings, J.M., L.W. Hall Jr., K.R. Solomon. 2000. Ecological risks of diazinon from
4920 agricultural use in the Sacramento - San Joaquin River Basins, California. *Risk Analysis*
4921 20:545–572.
- 4922 Gingras, M. 1997. Mark/recapture experiments at Clifton Court Forebay to estimate pre-
4923 screening loss to juvenile fishes: 1976–1993. Interagency Ecological Program. Technical
4924 Report 55.
- 4925 Gifford, S.M., G. Rollwagen-Bollens, S.M. Bollens. 2007. Mesozooplankton omnivory in the
4926 upper San Francisco Estuary. *Marine Ecological Progress Series* 348:33–46.
- 4927 Glibert, P.M. 2010. Long-term changes in nutrient loading and stoichiometry and their
4928 relationships with changes in the food web and dominant pelagic fish species in the San
4929 Francisco Estuary, California. *Reviews in Fisheries Science* 18:211–232.
- 4930 Goldschmidt, T., F. Witte and J. Wanink. 1993. Cascading effects of the introduced Nile Perch
4931 on the detritivorous/ phytoplanktivorous species in the sublittoral areas of Lake Victoria.
4932 *Conservation Biology* 7:686–700.
- 4933 Gould, A.L. and W.J. Kimmerer. 2010. Development, growth, and reproduction of the cyclopoid
4934 copepod *Limnoithona tetraspina* in the upper San Francisco Estuary. *Marine Ecology*
4935 *Progress Series* 412:163–177.
- 4936 Gregory, R.S. and C.D. Levings. 1998. Turbidity reduces predation on migrating juvenile Pacific
4937 salmon. *Transactions of the American Fisheries Society* 127:275–285.
- 4938 Gregory, S.D., C.J.A. Bradshaw, B.W. Brook, and F. Courchamp. 2010. Limited evidence for the
4939 demographic Allee effect from numerous species across taxa. *Ecology* 91:2151–2161.
- 4940 Griffith, J.S. 1978. Effects of low temperature on the survival and behavior of threadfin shad,
4941 *Dorosoma petenense*. *Transactions of the American Fisheries Society* 107:63–70.
- 4942 Grimaldo, L. and Z. Hymanson. 1999. What is the impact of the introduced Brazilian waterweed
4943 *Egeria densa* to the Delta ecosystem? *IEP Newsletter* 12 (1):43–45.
- 4944 Grimaldo, L.F, R.E. Miller, C.M. Peregrin, and Z.P Hymanson. 2004. Spatial and temporal
4945 distribution of ichthyoplankton in three habitat types of the Sacramento-San Joaquin Delta.
4946 Pages 81–96 in F. Feyrer, L.R. Brown, R.L. Brown, and J.J. Orsi, editors. *Early Life History*
4947 *of Fishes in the San Francisco Estuary and Watershed*. American Fisheries Society,
4948 Symposium 39, Bethesda, Maryland.

- 4949 Grimaldo, L.F., T. Sommer, N. Van Ark, G. Jones, E. Holland, P. Moyle, B. Herbold, and P.
4950 Smith. 2009a. Factors affecting fish entrainment into massive water diversions in a tidal
4951 freshwater estuary: Can fish losses be managed? *North American Journal of Fisheries*
4952 *Management* 29:1253–1270.
- 4953 Grimaldo, L. F., A.R. Stewart, and W. Kimmerer. 2009b. Dietary segregation of pelagic and
4954 littoral fish assemblages in a highly modified tidal freshwater estuary. *Marine and Coastal*
4955 *Fisheries: Dynamics Management, and Ecosystem Science* 1:200–217.
- 4956 Groffman, P.M., J.S. Baron, T. Blett, A.J. Gold, I. Goodman, L.H., Gunderson, et al. 2006.
4957 Ecological thresholds: the key to successful environmental management or an important
4958 concept with no practical application? *Ecosystems* 9:1–13.
- 4959 Gross, E.S., M.L. MacWilliams, C.D. Holleman, and T.A. Hervier. 2010. POD 3-D Particle
4960 tracking modeling study: particle tracking model testing and applications report. Final report
4961 to POD management team.
- 4962 Hart, J. 2010. The once and future Delta – mending the broken heart of California. *Bay Nature*
4963 *Magazine* April–June 2010.
- 4964 Hartman, K.J. 2003. Population-level consumption by Atlantic coastal striped bass and the
4965 influence of population recovery upon prey communities. *Fisheries Management and*
4966 *Ecology* 10:281–288.
- 4967 Hastings, A. and D.B. Wysham. 2010. Regime shifts in ecological systems can occur with no
4968 warning. *Ecology Letters* 13:464–472.
- 4969 Haunstein, E. and C. Ramirez. 1986. The influence of salinity on the distribution of *Egeria densa*
4970 in the Valdivia River Basin Chile. *Archiv Fur Hydrobiologie* 107:511–520.
- 4971 Hayes, D.B., C. P. Ferreri, and W.M. Taylor. 1996. Linking fish habitat to their population
4972 dynamics. *Canadian Journal of Fisheries and Aquatic Sciences* 53(Supplement 1):383–390.
- 4973 Hazel, C.R. and D. W. Kelley. 1966. Zoobenthos of the Sacramento-San Joaquin Delta. Pages
4974 113–133 in D.W. Kelley, editor. *Ecological studies of the Sacramento-San Joaquin Estuary.*
4975 *Part 1: Zooplankton, zoobenthos, and fishes of San Pablo and Suisun Bays, zooplankton and*
4976 *zoobenthos of the Delta.* California Department of Fish and Game, Fish Bulletin 133.
- 4977 Healey, M.C., M.D. Dettinger, and R.B. Norgaard, editors. 2008. *The State of Bay-Delta*
4978 *Science, 2008.* CALFED Science Program, Sacramento, CA. 174 pp.
- 4979 Hennessy, A. 2008. Zooplankton monitoring 2007. *IEP Newsletter* 21(2):16–21.
- 4980 Hennessy, A. 2010. Zooplankton monitoring 2009. *IEP Newsletter* 23(2):15–22.
- 4981 Hennessy, A. and K.A. Hieb. 2007. Zooplankton monitoring 2006. *IEP Newsletter* 20(1):9–13.
- 4982 Herbold, B., A.D. Jassby, and P.B. Moyle. 1992. Status and trends of aquatic resources of the
4983 San Francisco Bay and Delta. San Francisco Estuary Project. 210 pp plus 2 appendices.
- 4984 Hestir, E.L. 2010. Trends in estuarine water quality and submerged aquatic vegetation invasion.
4985 Ph.D dissertation, University of California, Davis.
- 4986 Heubach, W., R.J. Toth, and A.M. McCready. 1963. Food of young-of-the-year striped bass
4987 (*Roccus saxatilis*) in the Sacramento-San Joaquin River system. *California Fish and Game*
4988 49:224–239.
- 4989 Hieb, K. 2007. Common shrimp of the San Francisco Estuary. *IEP Newsletter* 20(2):14–18.
- 4990 Hieb, K. and R. Baxter. 1993. Delta outflow/San Francisco Bay. Pages 101–116 in P.L.
4991 Herrgesell, editor. 1991 Annual Report - Interagency Ecological Studies Program for the
4992 Sacramento-San Joaquin Estuary.
- 4993 Hinton, D.E. and D.J. Lauren. 1990. Integrative histopathological approaches to detecting effects
4994 of environmental stressors on fishes. *American Fisheries Society Symposium* 8:51–66.

- 4995 Hobbs, J. A., W. A. Bennett, and J. E. Burton. 2006. Assessing nursery habitat quality for native
4996 smelts (Osmeridae) in the low-salinity zone of the San Francisco estuary. *Journal of Fish*
4997 *Biology* 69:907–922.
- 4998 Hobbs, J.A., L.S. Lewis, N. Ikemiyagi, T. Sommer and R.D. Baxter. 2010. The use of otolith
4999 strontium isotopes ($^{87}\text{Sr}/^{86}\text{Sr}$) to identify nursery habitat for a threatened estuarine fish.
5000 *Environmental Biology of Fishes* 89:557–569.
- 5001 Hoff, T.B., J.B. McLaren, and J.C. Cooper. 1988. Stock characteristics of Hudson River striped
5002 bass. Pages 59–68 in L.W. Barnthouse and R.J. Klauda, editors. *Science, law and Hudson*
5003 *River power plants: a case study in environmental impact assessment*. American Fisheries
5004 Society, Bethesda, Maryland.
- 5005 Hollibaugh, J.T., and P.S. Wong. 1996. Distribution and activity of bacterioplankton in San
5006 Francisco Bay. Pages 263–288 in J. T. Hollibaugh, editor. *San Francisco Bay: the ecosystem*.
5007 Pacific Division of the American Association for the Advancement of Science, San
5008 Francisco, CA.
- 5009 Hornberger, M.I., S.N. Luoma, D. Cain, F. Parchaso, C. Brown, R. Bouse, C. Wellise, and J.
5010 Thompson. 2000. Linkage of bioaccumulation and biological effects to changes in pollutant
5011 loads in South San Francisco Bay. *Environmental Science and Technology* 34:2401–2409.
- 5012 Houde, E.D. 1987. Fish early life dynamics and recruitment variability. *American Fisheries*
5013 *Society Symposium* 2:17–29.
- 5014 Howard, J. 2010. Sensitive freshwater mussel surveys in the Pacific Southwest Region:
5015 assessment of conservation status. Final report to the USDA Forest Service, May 2010.
5016 Available at <http://nrm.dfg.ca.gov/documents/ContextDocs.aspx?cat=BIOS>
- 5017 Huang, C.H. and D.L. Sedlak. 2001. Analysis of estrogenic hormones in municipal wastewater
5018 effluent and surface water using enzyme-linked immunosorbent assay and gas
5019 chromatography/tandem mass spectrometry. *Environmental Toxicology and Chemistry*
5020 20:133–139.
- 5021 Hudson, P.L., R.W. Griffiths, and T.J. Wheaton. 1992. Review of habitat classification schemes
5022 appropriate to streams, rivers, and connecting channels in the Great Lakes drainage basin.
5023 Pages 73–107, in W.D.N. Busch and P.G. Sly, editors. *The development of an aquatic habitat*
5024 *classification system for lakes*. CRC Press, Ann Arbor, MI.
- 5025 IEP (Interagency Ecological Program for the San Francisco Estuary). 2005. Interagency
5026 Ecological Program 2005 Work plan to evaluate the decline of pelagic species in the upper
5027 San Francisco Estuary. Available at:
5028 [http://www.science.calwater.ca.gov/pdf/workshops/POD/2005_IEP-](http://www.science.calwater.ca.gov/pdf/workshops/POD/2005_IEP-POD_Workplan_070105.pdf)
5029 [POD_Workplan_070105.pdf](http://www.science.calwater.ca.gov/pdf/workshops/POD/2005_IEP-POD_Workplan_070105.pdf).
- 5030 IEP (Interagency Ecological Program for the San Francisco Estuary). 2006. Interagency
5031 Ecological Program Synthesis of 2005 Work to Evaluate the Pelagic Organism Decline
5032 (POD) in the Upper San Francisco Estuary. Available at:
5033 [http://science.calwater.ca.gov/pdf/workshops/IEP_POD_2005WorkSynthesis-](http://science.calwater.ca.gov/pdf/workshops/IEP_POD_2005WorkSynthesis-draft_111405.pdf)
5034 [draft_111405.pdf](http://science.calwater.ca.gov/pdf/workshops/IEP_POD_2005WorkSynthesis-draft_111405.pdf)
- 5035 IPCC (Intergovernmental Panel on Climate Change). 2001. *Climate change 2001: Impacts,*
5036 *adaptation and vulnerability*. World Meteorological Organization, Geneva, Switzerland.
- 5037 Jang, M., K. Ha, M.C. Lucas, G. Joo, and N. Takamura. 2004. Changes in microcystin
5038 production by *Microcystis aeruginosa* exposed to phytoplanktivorous and omnivorous fish.
5039 *Aquatic Toxicology* 68:5159.

- 5040 Jassby, A.D. 2008. Phytoplankton in the upper San Francisco Estuary: recent biomass trends,
5041 their causes and their trophic significance. San Francisco Estuary and Watershed Science.
5042 Vol. 6, Issue 1, Article 2.
- 5043 Jassby, A.D. and J. E. Cloern. 2000. Organic matter sources and rehabilitation of the
5044 Sacramento-San Joaquin Delta (California, USA). Aquatic Conservation: Marine and
5045 Freshwater Ecosystems 10:323–352.
- 5046 Jassby, A. D., J. E. Cloern, and B. E. Cole. 2002. Annual primary production: patterns and
5047 mechanisms of change in a nutrient-rich tidal ecosystem. Limnology and Oceanography 47:
5048 698–712.
- 5049 Jassby, A. D., W. J. Kimmerer, S. G. Monismith, C. Armor, J. E. Cloern, T. M. Powell, J. R.
5050 Schubel, and T. J. Vendlinski. 1995. Isohaline position as a habitat indicator for estuarine
5051 populations. Ecological Applications 5: 272–289.
- 5052 Jassby, A. D., A. B. Mueller-Solger, and M. Vayssieres. 2005. Subregions of the Sacramento-
5053 San Joaquin Delta: identification and use. IEP Newsletter 18(2):68–75.
- 5054 Jassby, A.D., and E.E. Van Nieuwenhuyse. 2005. Low dissolved oxygen in an estuarine channel
5055 (San Joaquin River, California): mechanisms and models based on long-term time series. San
5056 Francisco Estuary and Watershed Science Vol. 3, Issue 2, Article 2.
- 5057 Jin, M., L. Li, C. Xu, Y. Wen, and M. Zhao. 2010. Estrogenic activities of two synthetic
5058 pyrethroids and their metabolites. Journal of Environmental Sciences 22:290–296.
- 5059 Johnson, J.E. 1970. Age, growth, and population dynamics of threadfin shad *Dorosoma*
5060 *petenense* (Gunther), in central Arizona reservoirs. Transactions of the American Fisheries
5061 Society 99:739–753.
- 5062 Johnson, M.L., I. Werner, S. Teh, and F. Loge. 2010. Evaluation of chemical, toxicological, and
5063 histopathological data to determine their role in the pelagic organism decline. Final report to
5064 the California State Water Resources Control Board and Central Valley Regional Water
5065 Quality Control Board. University of California, Davis.
- 5066 Jones, C.G., J.H. Lawton, and M. Shachak. 1994. Organisms as ecosystem engineers. Oikos
5067 69:373–86.
- 5068 Jones, C.G., J.H. Lawton, and M. Shachak. 1997. Positive and negative effects of organisms as
5069 physical ecosystem engineers. Ecology 78:1946–1957.
- 5070 Kano, R.M. 1990 Occurrence and abundance of predator fish in Clifton Court Forebay,
5071 California. Department of Fish and Game. Interagency Ecological Study Program for the
5072 Sacramento-San Joaquin Estuary. Technical Report 24.
- 5073 Kashuba, S.A., and W.J. Matthews. 1984. Physical condition of larval shad during spring–
5074 summer in a southeastern reservoir. Transactions of the American Fisheries Society 113:199–
5075 2004.
- 5076 Kimmerer, W. J. 2002a. Physical, biological, and management responses to variable freshwater
5077 flow into the San Francisco Estuary. Estuaries 25: 1275–1290.
- 5078 Kimmerer, W.J. 2002b. Effects of freshwater flow on abundance of estuarine organisms:
5079 physical effects or trophic linkages. Marine Ecology Progress Series 243:39–55.
- 5080 Kimmerer, W.J. 2004. Open water processes of the San Francisco Estuary: from physical
5081 forcing to biological responses. San Francisco Estuary and Watershed Science. Vol. 2, Issue
5082 1, Article 2.
- 5083 Kimmerer, W.J. 2005. Long-term changes in apparent uptake of silica in the San Francisco
5084 Estuary. Limnology and Oceanography 50:793–798.

- 5085 Kimmerer, W.J. 2006. Response of anchovies dampens effects of the invasive bivalve *Corbula*
5086 *amurensis* on the San Francisco Estuary foodweb. *Marine Ecology Progress Series* 324:207–
5087 218.
- 5088 Kimmerer, W.J. 2008a. Losses of Sacramento River Chinook salmon and delta smelt to
5089 entrainment in water diversions in the Sacramento-San Joaquin Delta. *San Francisco Estuary*
5090 *and Watershed Science*. Vol. 6, Issue 2, Article 2.
- 5091 Kimmerer, W.J. 2008b. Long-term changes in an estuarine zooplankton community: species
5092 introductions and clam grazing. Report to the Department of Water Resources, Contract B-
5093 81623. Romberg Tiburon Center, San Francisco State University.
- 5094 Kimmerer, W., S. Avent, S. Bollens, F. Feyrer, L. Grimaldo, P. Moyle, M. Nobriga, and T.
5095 Visintainer. 2005. Variability in length-weight relationships used to estimate biomass of
5096 estuarine fishes from survey data. *Transactions of the American Fisheries Society* 134:481–
5097 495
- 5098 Kimmerer, W.J., J.H. Cowan, Jr., L.W. Miller, and K.A. Rose. 2000. Analysis of an estuarine
5099 striped bass (*Morone saxatilis*) population: influence of density-dependent mortality between
5100 metamorphosis and recruitment. *Canadian Journal of Fisheries and Aquatic Sciences* 57:
5101 478–486.
- 5102 Kimmerer, W. J., J. H. Cowan, Jr., L. W. Miller, and K. A. Rose. 2001. Analysis of an estuarine
5103 striped bass population: effects of environmental conditions during early life. *Estuaries* 24:
5104 557–575.
- 5105 Kimmerer, K.J., N. Ferm, M.H. Nicolini, and C. Penalva. 2005. Chronic food limitation of egg
5106 production in populations of copepods of the genus *Acartia* in the San Francisco Estuary.
5107 *Estuaries* 28:541–550.
- 5108 Kimmerer, W.J., E. Gartside, J.J. Orsi. 1994. Predation by an introduced clam as the likely cause
5109 of substantial declines in zooplankton of San Francisco Bay. *Marine Ecology Progress Series*
5110 113: 81–93.
- 5111 Kimmerer, W. J., E. S. Gross, and M. L. MacWilliams. 2009. Is the response of estuarine nekton
5112 to freshwater flow in the San Francisco Estuary explained by variation in habitat volume?
5113 *Estuaries and Coasts* 32:375–389.
- 5114 Kimmerer, W. and M. Nobriga. 2008. Investigating dispersal in the Sacramento-San Joaquin
5115 Delta using a particle tracking model. *San Francisco Estuary and Watershed Science*. Vol. 6,
5116 Issue 1, Article 4.
- 5117 Kimmerer, W. J. and J. J. Orsi. 1996. Changes in the zooplankton of the San Francisco Bay
5118 Estuary since the introduction of the clam *Potamocorbula amurensis*. Pages 403–424 in J.T.
5119 Hollibaugh, editor. *San Francisco Bay: the ecosystem*. Pacific Division of the American
5120 Association for the Advancement of Science. San Francisco, California, USA.
- 5121 Kimsey, J.B., R.H. Hagy, and W. McGammon. 1957. Progress report of the Mississippi threadfin
5122 shad, *Dorosoma petenense atchafaylae*, in the Colorado River for 1966. *California Fish and*
5123 *Game Inland Fisheries Administrative Report* 57–23.
- 5124 Knowles, N. 2010. Potential inundation due to rising sea levels in the San Francisco Bay Region.
5125 *San Francisco Estuary and Watershed Science*. Vol. 8, Issue 1. Retrieved from:
5126 <http://escholarship.org/uc/item/8ck5h3qn>
- 5127 Knowles, N. and D. Cayan. 2002. Potential effects of global warming on the Sacramento/San
5128 Joaquin watershed and the San Francisco estuary. *Geophysical Research Letters* 29:38-1–38-
5129 4.

- 5130 Knowles, N. and D. Cayan. 2004. Elevational dependence of projected hydrologic changes in
5131 the San Francisco Estuary and watershed. *Journal Climatic Change* 62:319–336.
- 5132 Kohlhorst, D.W. 1999. Status of striped bass in the Sacramento-San Joaquin estuary. *California*
5133 *Fish and Game* 85:31–36.
- 5134 Kohlhorst, D.W. 1976. Eutrophication and fishery resources – a literature review applicable to
5135 future conditions in the Sacramento-San Joaquin Estuary. CDFG, Anadromous Fish. Branch
5136 Admin. Report No 76-8, 20 pp.
- 5137 Kuivila, K. M., and C.G. Foe. 1995. Concentrations, transport and biological effects of dormant
5138 spray pesticides in the San Francisco Estuary, California. *Environmental Toxicology and*
5139 *Chemistry* 14:1141–1150.
- 5140 Kuivila, K.M., and M. L. Hladik. 2008. Understanding the occurrence and transport of current-
5141 use pesticides in the San Francisco Estuary Watershed. *San Francisco Estuary and Watershed*
5142 *Science*. Vol. 6, Issue 3, Article 2.
- 5143 Kuivila, K. and G.E. Moon. 2004. Potential exposure of larval and juvenile delta smelt to
5144 dissolved pesticides in the Sacramento–San Joaquin Delta, California. Pages 229–241 in F.
5145 Feyrer, L. R. Brown, R. L. Brown, and J. J. Orsi, editors. *Early life history of fishes in the*
5146 *San Francisco Estuary and watershed*. American Fisheries Society, Symposium 39, Bethesda,
5147 Maryland.
- 5148 Lavado, R. J.E. Loyo-Rosales, E. Floyd, E.P. Kolodziej, S.A. Snyder, D.L. Sedlak, and D.
5149 Schlenk. 2009. Site-Specific Profiles of Estrogenic Activity in Agricultural Areas of
5150 California’s Inland Waters. *Environmental Science and Technology* 43:9110–9116.
- 5151 Lehman, P. W. 1996. Changes in chlorophyll a concentration and phytoplankton community
5152 composition with water-year type in the upper San Francisco Bay Estuary. Pages 351–374 in
5153 J. T. Hollibaugh, editor. *San Francisco Bay: The Ecosystem*. Pacific Division of the
5154 American Association for the Advancement of Science.
- 5155 Lehman, P. W. 2000. Phytoplankton biomass, cell diameter and species composition in the low
5156 salinity zone of northern San Francisco Bay Estuary. *Estuaries* 23:216–230.
- 5157 Lehman, P. W. 2007. The influence of phytoplankton community composition on primary
5158 productivity along the riverine to freshwater tidal continuum in the San Joaquin River,
5159 California. *Estuaries and Coasts* 30:82–93.
- 5160 Lehman, P. W., G. Boyer, C. Hall, S. Waller, and K. Gehrts. 2005. Distribution and toxicity of a
5161 new colonial *Microcystis aeruginosa* bloom in the San Francisco Bay Estuary, California.
5162 *Hydrobiologia* 541:87–99.
- 5163 Lehman, P.W., G. Boyer, M. Satchwell, and S. Waller. 2008a. The influence of environmental
5164 conditions on the seasonal variation of *Microcystis* cell density and microcystins
5165 concentration in San Francisco Estuary. *Hydrobiologia* 600:187–204.
- 5166 Lehman, P.W., S. Mayr, L. Mecum, and C. Enright. 2010. The freshwater tidal wetland Liberty
5167 Island, CA was both a source and sink of inorganic and organic material to the San Francisco
5168 Estuary. *Aquatic Ecology* 44:359–372.
- 5169 Lehman, P.W., J. Sevier, J. Giulianotti, and M. Johnson. 2004. Sources of oxygen demand in the
5170 lower San Joaquin River, California. *Estuaries* 27:405–418.
- 5171 Lehman, P.W., T. Sommer, and L. Rivard. 2008b. The influence of floodplain habitat on the
5172 quantity and quality of riverine phytoplankton carbon produced during the flood season in
5173 San Francisco Estuary. *Aquatic Ecology* 42:363–378.

- 5174 Lehman, P.W., S. Teh, G.L. Boyer, M. Nobriga, E. Bass and C. Hogle. 2010. Initial impacts of
5175 *Microcystis* on the aquatic food web in the San Francisco Estuary. *Hydrobiologia* 637:229–
5176 248.
- 5177 Leonard, L.A., P.A. Wren, and R.L. Beavers. 2002. Flow dynamics and sedimentation in
5178 *Spartina alterniflora* and *Phragmites australis* marshes of the Chesapeake Bay. *Wetlands*
5179 22:415–424.
- 5180 Lindley, S. T., C.B. Grimes, M.S. Mohr, W. Peterson, J. Stein, J.T. Anderson, L.W. Botsford, ,
5181 D.L. Bottom, C.A. Busack, T.K. Collier, J. Ferguson, J.C. Garza, A.M. Grover, D.G. Hankin,
5182 R.G. Kope, P.W. Lawson, A. Low, R.B. MacFarlane, K. Moore, M. Palmer-Zwahlen, F.B.
5183 Schwing, J. Smith, C. Tracy, R. Webb, B.K. Wells, T.H. Williams. 2009. What caused the
5184 Sacramento River fall Chinook stock collapse? Pre-publication report to the Pacific Fishery
5185 Management Council, Portland, Oregon. [http://www.pcouncil.org/wp-](http://www.pcouncil.org/wp-content/uploads/H2b_WGR_0409.pdf)
5186 [content/uploads/H2b_WGR_0409.pdf](http://www.pcouncil.org/wp-content/uploads/H2b_WGR_0409.pdf) .
- 5187 Linkov, I., D. Loney, S. Cormier, F.K. Satterstrom, T. Bridges. 2009. Weight-of-evidence
5188 evaluation in environmental assessment: Review of qualitative and quantitative approaches.
5189 *Science of the Total Environment* 407:5199–5205.
- 5190 Linville, R. G., S. N. Luoma, et al. (2002). Increased selenium threat as a result of invasion of
5191 the exotic bivalve *Potamocorbula amurensis* into the San Francisco Bay-Delta. *Aquatic*
5192 *Toxicology*. 57: 1–2.
- 5193 Livingston, R.J., X. Niu, F.G. Lewis, III, and G.C. Woodsum. 1997. Freshwater input to a gulf
5194 estuary: long-term control of trophic organization. *Ecological Applications*:277–299.
- 5195 Lluch-Belda, D., R.J.M.Crawford, T. Kawasaki, A.D. McCall, R.H. Parrish, R.A. Schwartzlose,
5196 and P.E. Smith. 1989. Worldwide fluctuations of sardine and anchovy stocks: The regime
5197 problem. *South African Journal of Marine Science* 8:195–205.
- 5198 Lopez, C.B., J.E. Cloern, T.S. Schraga, A.J. Little, L.V. Lucas, J.K. Thompson, and J.R. Burau.
5199 2006. Ecological values of shallow-water habitats: Implications for the restoration of
5200 disturbed ecosystems. *Ecosystems* 9:422–440.
- 5201 Lott, J. 1998. Feeding habits of juvenile and adult delta smelt from the Sacramento-San Joaquin
5202 River Estuary. *IEP Newsletter* 11(1):14–19.
- 5203 Lotze, H.K., H.S. Lenihan, B.J. Bourque, R.H. Bradbury, R.G. Cooke, M.C. Kay, S.M. Kidwell,
5204 M.X. Kirby, C.H. Peterson, J.B.C. Jackson. 2006. Depletion, degradation, and recovery
5205 potential of estuaries and coastal seas. *Science* 312:1806–1809.
- 5206 Lucas, L.V., J.E. Cloern, J.K. Thompson, and N.E. Mosen. 2002. Functional variability of
5207 habitats within the Sacramento-San Joaquin Delta: restoration implications. *Ecological*
5208 *Applications* 12:1528–1547.
- 5209 Lucas, L.V., J.R. Koseff, S.G. Monismith, and J.K. Thompson. 2009a. Shallow water processes
5210 govern system-wide phytoplankton bloom dynamics - A modeling study. *Journal of Marine*
5211 *Systems* 75:70–86.
- 5212 Lucas, L.V., D.M. Sereno, J.R. Burau, T.S. Schraga, C.B. Lopez, M.T. Stacey, K.V. Parchevsky,
5213 and V.P. Parchevsky. 2006. Intratidal variability of water quality in a shallow tidal lagoon:
5214 Mechanisms and implications. *Estuaries and Coasts* 29:711–730.
- 5215 Lucas, L.V., J.K. Thompson, and L.R. Brown. 2009b. Why are diverse relationships observed
5216 between phytoplankton biomass and transport time? *Limnology and Oceanography* 54:381–
5217 390.
- 5218 Lund, J., E. Hanak, W. Fleenor, R. Howitt, J. Mount, and P. Moyle. 2007. Envisioning futures
5219 for the Sacramento–San Joaquin Delta. Public Policy Institute of California.

- 5220 Lund, J., E. Hanak, W. Fleenor, W. Bennett, R. Howitt, J. Mount, and P. Moyle. 2008.
5221 Comparing Futures for the Sacramento-San Joaquin Delta. Public Policy Institute of
5222 California.
- 5223 Mac Nally, R., J.R. Thomson, W.J. Kimmerer, F. Feyrer, K.B. Newman, A. Sih, W.A. Bennett,
5224 L. Brown, E. Fleishman, S.D. Culberson, and G. Castillo. 2010. An analysis of pelagic
5225 species decline in the upper San Francisco Estuary using Multivariate Autoregressive
5226 modeling (MAR). *Ecological Applications* 20:167–180.
- 5227 Mager R, S.I. Doroshov, J.P. Van Eenennaam, and R.L. Brown. 2004. Early life stages of delta
5228 smelt. Pages 169–180 in F. Freyer, L.R. Brown, R.L. Brown, and J.J. Orsi, editors. Early life
5229 history of fishes in the San Francisco Estuary and watershed. American Fisheries Society
5230 Symposium 39. Bethesda, MD: American Fisheries Society.
- 5231 Malamud-Roam, F., M. Dettinger, B.L. Ingram, M.K. Hughes, and J.L. Florsheim. 2007.
5232 Holocene climates and connections between the San Francisco Bay Estuary and its
5233 watershed: a review. *San Francisco Estuary and Watershed Science*. Vol. 5, Issue 1, Article
5234 3. <http://repositories.cdlib.org/jmie/sfews/vol5/iss1/art3>
- 5235 Manly, B.J.F. and M.A. Chotkowski. 2006. Two new methods for regime change analysis.
5236 *Archiv für Hydrobiologie* 167:593–607.
- 5237 Marteinsdottir, G. and A. Steinarsson 1998. Maternal influence on the size and viability of
5238 Iceland cod, *Gadus morhua*, eggs and larvae. *Journal of Fish Biology* 52:1241–1258.
- 5239 Masuzaki, S. S., N. Usio, N. Takamura, and I. Washitani. 2009. Contrasting impacts of invasive
5240 engineers on freshwater ecosystems: an experiment and meta – analysis. *Oecologia* 158:673–
5241 686.
- 5242 McLean, R.B., J.S. Griffith, and M.V. McGee. 2006. Threadfin shad, *Dorosoma petenense*
5243 Gunther, mortality: causes and ecological implications in a South-eastern United States
5244 reservoir. *Journal of Fish Biology* 27:1–12.
- 5245 McLean, R.B., P.T. Single, D.M. Lodge, and R.A. Wallace. 1982. Synchronous spawning of
5246 threadfin shad. *Copeia* 4:952–955.
- 5247 McLeod, K. and H. Leslie, editors. 2009. *Ecosystem-based management for the oceans*. Island
5248 Press. 392 p.
- 5249 McLeod, K.L., J. Lubchenco, S.R. Palumbi, and A.A. Rosenberg. 2005. Scientific Consensus
5250 Statement on Marine Ecosystem-Based Management. Signed by 221 academic scientists and
5251 policy experts with relevant expertise and published by the Communication Partnership for
5252 Science and the Sea at <http://compassonline.org/?q=EBM>.
- 5253 McManus, G.B., J.K. York and W.J. Kimmerer. 2008. Miczooplankton dynamics in the low
5254 salinity zone of the San Francisco Estuary. *Verhandlungen Internationale Vereinigung für*
5255 *Theoretische und Angewandte Limnologie* 30:198–202.
- 5256 Meade, R.H. 1982. Sources, sinks, and storage of river sediment in the Atlantic drainage of the
5257 United States. *The Journal of Geology* 90: 235–252.
- 5258 Meffe, G.K. 1984. The effects of abiotic disturbance on coexistence of predator-prey fish
5259 species. *Ecology* 65:1525–1534.
- 5260 Melbourne, B.A. and A. Hastings. 2008. Extinction risk depends strongly on factors contributing
5261 to stochasticity. *Nature* 454:100–103.
- 5262 Meng, L. and J.J. Orsi. 1991. Selective predation by larval striped bass on native and introduced
5263 copepods. *Transactions of the American Fisheries Society* 120:187–192.
- 5264 Miller, D.M., S.P. Finn, A. Woodward, A. Torregrosa, M.E. Miller, D.R. Bedford, and
5265 A.M.Brasher. 2010. Conceptual ecological models to guide integrated landscape monitoring

5266 of the Great Basin. U.S. Geological Survey Scientific Investigations Report 2010-5133, 134
5267 pp.

5268 Miller M.A., R.M. Kudela, A. Mekebri, D. Crane, S.C. Oates, et al. 2010. Evidence for a novel
5269 marine harmful algal bloom: cyanotoxin (microcystin) transfer from land to sea otters. PLoS
5270 ONE 5(9): e12576. doi:10.1371/journal.pone.0012576

5271 Moisander, P., P.W. Lehman, M. Ochiai, and S. Corum. 2009. Diversity of the toxic
5272 cyanobacterium *Microcystis aeruginosa* in the Klamath River and San Francisco Bay delta,
5273 California. *Aquatic Toxicology* 57:19–31.

5274 Monsen, N.E., J.E. Cloern, and J.R. Burau. 2007. Effects of flow diversions on water and habitat
5275 quality; examples from California's highly manipulated Sacramento-San Joaquin Delta. *San*
5276 *Francisco Estuary and Watershed Science*. Vol. 5, Issue 3, Article 2.

5277 Mount, J.F. 1995. *California rivers and streams: the conflict between fluvial process and land*
5278 *use*. University of California Press. Berkeley.

5279 Mount J.F. and R. Twiss. 2005. Subsidence, sea level rise, seismicity in the Sacramento-San
5280 Joaquin Delta. *San Francisco Estuary and Watershed Science*. Vol. 3, Issue 1. Available
5281 from: <http://www.escholarship.org/uc/item/4k44725p>.

5282 Moyle, P.B. 2002. *Inland fishes of California*. Revised and expanded. University of California
5283 Press, Berkeley, California.

5284 Moyle, P.B. and W.A. Bennett. 2008. The future of the Delta ecosystem and its fish. Technical
5285 Appendix D in J. Lund, E. Hanak, W. Fleenor, W. Bennett, R. Howitt, J. Mount, and P.
5286 Moyle. *Comparing Futures for the Sacramento-San Joaquin Delta*. Public Policy Institute of
5287 California, San Francisco, CA.

5288 Moyle, P.B., W.A. Bennett, W.E. Fleenor, and J.R. Lund. 2010. Habitat variability and
5289 complexity in the upper San Francisco Estuary. *San Francisco Estuary and Watershed*
5290 *Science*. Vol. 8, Issue 3.

5291 Mueller-Solger, A.B., C.J. Hall, A.D. Jassby, and C.R. Goldman. 2006. Food resources for
5292 zooplankton in the Sacramento-San Joaquin Delta. Final Report to the Calfed Ecosystem
5293 Restoration Program, May 2006.

5294 Mueller-Solger, A.B., A.D.Jassby, and D.C. Mueller-Navarra. 2002. Nutritional quality of food
5295 resources for zooplankton (*Daphnia*) in a tidal freshwater system (Sacramento-San Joaquin
5296 River Delta), *Limnology and Oceanography* 47:1468–1476.

5297 Murrell, M.C. and J.T. Hollibaugh. 1998. Microzooplankton grazing in northern San Francisco
5298 Bay measured by the dilution method. *Aquatic Microbial Ecology* 15:53–63.

5299 Myers, R.A., J. Bridson, and N.J. Barrowman. 1995. Summary of worldwide spawner and
5300 recruitment data. Canadian Technical Report on Fisheries and Aquatic Sciences No. 2024.

5301 National Marine Fisheries Service (NMFS). 1997. Investigation of Scientific Information on the
5302 Impacts of California Sea Lions and Pacific Harbor Seals on Salmonids and on the Coastal
5303 Ecosystems of Washington, Oregon, and California. U.S. Dep. Commer., NOAA Tech.
5304 Memo. NMFS-NWFSC-28, 172 p.

5305 Newman, K.B. 2008. Sample design-based methodology for estimating delta smelt abundance.
5306 *San Francisco Estuary and Watershed Science*. Vol. 6, Issue 3, Article 3.
5307 <http://repositories.cdlib.org/jmie/sfews/vol6/iss3/art3>

5308 Nichols, F. H. 1979. Natural and anthropogenic influences on benthic community structure in the
5309 San Francisco Bay. Pages 409–426 in T. J. Conomos, editor. *San Francisco Bay: The*
5310 *Urbanized Estuary*. Pacific Division of the American Association for the Advancement of
5311 Science, San Francisco, CA.

- 5312 Nichols, F.H., J.E. Cloern, S.N. Luoma, and D.H. Peterson. 1986. The modification of an
5313 estuary. *Science* 231:567–573.
- 5314 Nichols, F.H., J.K. Thompson, and L.E. Schemel. 1990. Remarkable invasion of San Francisco
5315 Bay (California, USA) by the Asian clam *Potamocorbula amurensis*. II. Displacement of a
5316 former community. *Marine Ecology Progress Series* 66:95–101.
- 5317 Nixon, S.W. 1988. Physical energy inputs and the comparative ecology of lake and marine
5318 ecosystems. *Limnology and Oceanography, Part II* 33:1005–1025.
- 5319 Nobriga, M. 2002. Larval delta smelt composition and feeding incidence: environmental and
5320 ontogenetic influences. *California Fish and Game* 88:149–164.
- 5321 Nobriga, M.L. 2009. Bioenergetic modeling evidence for a context-dependent role of food
5322 limitation in California’s Sacramento-San Joaquin Delta. *California Fish and Game* 95:111–
5323 121.
- 5324 Nobriga, M. and F. Feyrer. 2007. Shallow-water piscivore-prey dynamics in the Sacramento-San
5325 Joaquin Delta. *San Francisco Estuary and Watershed Science* Vol. 5, Issue 2, Article 4.
- 5326 Nobriga, M.L., F. Feyrer, R.D. Baxter, and M. Chotkowski. 2005. Fish community ecology in an
5327 altered river Delta: spatial patterns in species composition, life history strategies and
5328 biomass. *Estuaries*:776–785.
- 5329 Nobriga, M.L., Z. Matica, and Z.P. Hymanson. 2004. Evaluating entrainment vulnerability to
5330 agricultural irrigation diversions: A comparison among open-water fishes. Pages 281–295 in
5331 F. Feyrer, L.R. Brown, R.L. Brown, and J.J. Orsi, editors. *Early Life History of Fishes in the*
5332 *San Francisco Estuary and Watershed*. American Fisheries Society, Symposium 39,
5333 Bethesda, Maryland.
- 5334 Nobriga, M.L., T. Sommer, F. Feyrer, and K. Fleming. 2008. Long-term trends in summertime
5335 habitat suitability for delta smelt, *Hypomesus transpacificus*. *San Francisco Estuary and*
5336 *Watershed Science*. Vol. 6, Issue 1, Article 1.
5337 <http://repositories.cdlib.org/jmie/sfews/vol6/iss1/art1>.
- 5338 Ogden, J.C., S.M. Davis, K.J. Jacobs, T. Barnes, and H.E. Fling. 2005. The use of conceptual
5339 ecological models to guide ecosystem restoration in south Florida. *Wetlands* 25:795–809.
- 5340 Ohlendorf, H.M. 2002. The birds of Kesterson Reservoir: a historical perspective. *Aquatic*
5341 *Toxicology* 57:1–10.
- 5342 Oros, D.R. and I. Werner. 2005. Pyrethroid Insecticides: an analysis of use patterns,
5343 distributions, potential toxicity and fate in the Sacramento-San Joaquin Delta and Central
5344 Valley. *San Francisco Estuary Institute*, Oakland, California.
- 5345 Orsi, J.J. and W.L. Mecum. 1986. Zooplankton distribution and abundance in the Sacramento-
5346 San Joaquin Delta in relation to certain environmental factors. *Estuaries* 9(4B):326–339.
- 5347 Orsi, J.J. and W.L. Mecum. 1996. Food limitation as the probable cause of a long-term decline in
5348 the abundance of *Neomysis mercedis* the opossum shrimp in the Sacramento-San Joaquin
5349 estuary. Pages 375–401 in J.T. Hollibaugh, editor. *San Francisco Bay: the ecosystem*.
5350 American Association for the Advancement of Science. San Francisco, CA.
- 5351 Ospina-Alvarez, N. and F. Piferrer, 2008. Temperature-dependent sex determination in fish
5352 revisited: prevalence, a single sex ratio response pattern, and possible effects of climate
5353 change *PloS One* 3:1–11.
- 5354 Ostrach, D.J., J. Groff, K. Springman, F. Loge, K. Eder, H. Haeri, and A. Massoudieh. 2005.
5355 Pathobiological Investigation to determine the condition of field collected 2005 striped bass
5356 and a pilot study to determine the location and effects of bioavailable lipophilic compounds
5357 in the San Francisco Estuary. UC Davis, Davis, California.

- 5358 Ostrach, D.J., J. Groff P. Weber T. Ginn F. Loge. 2009. The role of contaminants, within the
5359 context of multiple stressors, in the collapse of the striped bass population in the San
5360 Francisco Estuary and its watershed. Year 2 Final Report for DWR Agreement No.
5361 4600004664. U.C. Davis, Davis, California.
- 5362 Ostrach, D.J., J.M. Low-Marchelli, K.J. Eder, S.J. Whiteman, and J.G. Zinkl. 2008. Maternal
5363 transfer of xenobiotics and effects on larval striped bass in the San Francisco Estuary.
5364 Proceedings on the National Academy of Sciences 105: 19354–19359.
- 5365 Painter, R.E. 1966. Zoobenthos of San Pablo and Suisun Bays. Pages 40–56 in D.W. Kelley,
5366 editor. Ecological studies of the Sacramento-San Joaquin Estuary. Part 1: Zooplankton,
5367 zoobenthos, and fishes of San Pablo and Suisun Bays, zooplankton and zoobenthos of the
5368 Delta,. California Department of Fish and Game, Fish Bulletin 133.
- 5369 Parsons, J.W. and J.B. Kimsey. 1954. A report on the Mississippi threadfin shad. Progressive
5370 Fish Culturist 16:179–181.
- 5371 Pasternack, G.B. and G.S. Brush, 2001. Seasonal variations in sedimentation and organic content
5372 in five plant associations on a Chesapeake Bay tidal freshwater delta. Estuarine, Coastal and
5373 Shelf Science 53: 93–106.
- 5374 Paerl, H.W. and J. Huisman. 2008. Blooms like it hot. Science 320:57–58.
- 5375 Peterson, H. and M. Vayssières. 2010. Benthic assemblage variability in the upper San Francisco
5376 Estuary: a 27-year retrospective. San Francisco Estuary and Watershed Science Volume 8,
5377 Issue 1, Article 2. <http://www.escholarship.org/uc/item/4d0616c6>.
- 5378 Peterson, M.S. 2003. A conceptual view of environment-habitat-production linkages in tidal
5379 river estuaries. Reviews in Fisheries Science 11:291–313.
- 5380 Reddy, K.R. and J.C. Tucker. 1983. Productivity and nutrient uptake of water hyacinth,
5381 *Eichhornia crassipes* I. Effect on nitrogen sources. Economic Botany 37(2):237–247
- 5382 Rettinghouse, T. 2009. Fish Conservation and Culture Lab (FCCL), winter 2009. IEP Newsletter
5383 22(1):5.
- 5384 Rettinghouse, T. 2010. Fish Conservation and Culture Lab (FCCL), spring 2010. IEP Newsletter
5385 23(2):5–6.
- 5386 Rockström, J., W. Steffen, K. Noone, Å. Persson, F. S. Chapin, III, E. Lambin, T. M. Lenton, M.
5387 Scheffer, C. Folke, H. Schellnhuber, B. Nykvist, C. A. De Wit, T. Hughes, S. van der Leeuw,
5388 H. Rodhe, S. Sörlin, P. K. Snyder, R. Costanza, U. Svedin, M. Falkenmark, L. Karlberg, R.
5389 W. Corell, V. J. Fabry, J. Hansen, B. Walker, D. Liverman, K. Richardson, P. Crutzen, and J.
5390 Foley. 2009. A safe operating space for humanity. Nature 461:472–475.
- 5391 Rodríguez, M.A. and W.M. Lewis, Jr. 1994. Regulation and stability in fish assemblages of
5392 neotropical floodplain lakes. Oecologia 99:166–180.
- 5393 Rohrlack, T., K. Christoffersen, E. Dittmann, I. Nogueira, V. Vasconcelos, and T. Börner. 2005.
5394 Ingestion of microcystins by *Daphnia*: Intestinal uptake and toxic effects. Limnology and
5395 Oceanography 50:440–448.
- 5396 Rollwagen-Bollens, G.C. and D.L. Penry. 2003. Feeding dynamics of *Acartia* spp. copepods in a
5397 large, temperate estuary (San Francisco Bay, CA). Marine Ecology Progress Series 257:139–
5398 158.
- 5399 Roos, M. 1987. Possible changes in California snowmelt patterns. Proceedings, Fourth Pacific
5400 Climate Workshop, Pacific Grove, CA, 141–150.
- 5401 Roos, M. 1991. A trend of decreasing snowmelt runoff in northern California. Proceedings, 59th
5402 Western Snow Conference, Juneau, AK, 29–36.

- 5403 Rose, K.A. 2000. Why are quantitative relationships between environmental quality and fish
5404 populations so elusive? *Ecological Applications* 10:367–385.
- 5405 Rosenfield, J.A. and R.D. Baxter. 2007. Population dynamics and distribution patterns of longfin
5406 smelt in the San Francisco Estuary. *Transactions American Fisheries Society* 136:1577–
5407 1592.
- 5408 Sabo, J.L., J.C. Finlay, T. Kennedy, and D.M. Post. 2010. The role of discharge variation in
5409 scaling of drainage area and food chain length in rivers. *Science* 330:965–967.
- 5410 Saiki, M.K. and M.R. Jennings. 1992. Toxicity of agricultural subsurface drainwater from the
5411 San Joaquin Valley, California, to juvenile chinook salmon and striped bass. *Transactions of*
5412 *the American Fisheries Society* 121:78–93.
- 5413 Santos, M.J., L.W. Anderson, and S.L. Ustin. 2010. Effects of invasive species on plant
5414 communities: an example using submersed aquatic plants at the regional scale. *Biological*
5415 *Invasions*. Published online: <http://www.springerlink.com/content/b883gr221203xr37/>
- 5416 Scheffer, M. 1999. The effects of aquatic vegetation on turbidity; how important are the filter
5417 feeders? *Hydrobiologia* 408–409:307–316.
- 5418 Scheffer M. and S.R. Carpenter. 2003. Catastrophic regime shifts in ecosystems: linking theory
5419 to observation. *Trends in Ecology and Evolution* 18:648–656.
- 5420 Schindler, D.E., J.R. Hodgson, and J.F. Kitchell. 1997. Density-dependent changes in individual
5421 foraging specialization of largemouth bass. *Oecologia* 110:592–600.
- 5422 Schroeter, R. 2008. Biology and long-term trends of alien hydromedusae and striped bass in a
5423 brackish tidal marsh in the San Francisco Estuary. Ph.D. Dissertation. University of
5424 California, Davis.
- 5425 Secor, D.H. 2008. Influence of skipped spawning and misspecified reproductive schedules on
5426 biological reference points in sustainable fisheries. *Transactions of the American Fisheries*
5427 *Society* 137:782–789.
- 5428 Secor, D.H., J.R. Rooker, E. Zlokovitz, and V.S. Zdanowicz. 2001. Identification of riverine,
5429 estuarine, and coastal contingents of Hudson River striped bass based upon otolith elemental
5430 fingerprints. *Marine Ecology Progress Series* 211:245–253.
- 5431 Service, R.F. 2007. Delta blues, California style. *Science* 317:4444.
- 5432 Siegfried, C.A., A.W. Knight, and M.E. Kopache. 1978. Ecological studies on the western
5433 Sacramento-San Joaquin Delta during a dry year. *Water Science and Engineering Papers*
5434 4506, Department of Water Science and Engineering, U.C. Davis, 121 pp.
- 5435 Skinner, J.E. 1962. An historical review of the fish and wildlife resources of the San Francisco
5436 Bay area. California Department of Fish and Game, Water Projects, Branch Report No. 1.
- 5437 Shvidchenko, A.B., R.C. MacArthur, and B.R. Hall. 2004. Historic sedimentation in
5438 Sacramento-San Joaquin Delta. *IEP Newsletter* 17(3):21–30.
- 5439 Sobczak, W.V., J.E. Cloern, A.D. Jassby, and A.B. Muller-Solger. 2002. Bioavailability of
5440 organic matter in a highly disturbed estuary: The role of detrital and algal resources.
5441 *Proceedings of the National Academy of Sciences* 99:8101–8105.
- 5442 Sobczak, W.V., J.E. Cloern, A.D. Jassby, B.E. Cole, T.S. Schraga, and A. Arnsberg. 2005.
5443 Detritus fuels ecosystem metabolism but not metazoan food webs in San Francisco estuary's
5444 freshwater Delta. *Estuaries* 28:124–137.
- 5445 Sommer, T., C. Armor, R. Baxter, R. Breuer, L. Brown, M. Chotkowski, S. Culberson, F. Feyrer,
5446 M. Gingras, B. Herbold, W. Kimmerer, A. Mueller-Solger, M. Nobriga, and K. Souza. 2007.
5447 The collapse of pelagic fishes in the upper San Francisco Estuary. *Fisheries* 32(6):270–277.

- 5448 Sommer, T., R. Baxter, and B. Herbold. 1997. Resilience of splittail in the Sacramento-San
5449 Joaquin Estuary. *Transactions of the American Fisheries Society* 126:961–976.
- 5450 Stacey, M.T. 2003. Hydrodynamics of shallow water habitats in the Sacramento-San Joaquin
5451 Delta. UC Water Resources Center Technical Report, Project No. W-939, 10 pp.
- 5452 State Water Resources Control Board (SWRCB). 2010. DRAFT: Development of Flow Criteria
5453 for the Sacramento-San Joaquin Delta Ecosystem. State Water Resources Control Board,
5454 Sacramento, CA.
- 5455 Stephens, D.W. and J. R.Krebs. 1986. *Foraging Theory*. Princeton University Press, Princeton,
5456 NJ.
- 5457 Stevens, D.E. 1966. Food habits of striped bass, *Roccus saxatilis*, in the Sacramento-San Joaquin
5458 Delta. Pages 97–103 in J.T. Turner and D.W. Kelley, editors. *Ecological studies of the*
5459 *Sacramento-San Joaquin Delta, part II, fishes of the delta*. California Department of Fish and
5460 Game Fish Bulletin 136.
- 5461 Stevens, D.E. and L.W. Miller. 1983. Effects of river flow on abundance of young Chinook
5462 salmon, American shad, longfin smelt, and delta smelt in the Sacramento-San Joaquin River
5463 system. *North American Journal of Fisheries Management* 3:425–437.
- 5464 Stevens, D.E., D.W. Kohlhorst, L.W. Miller, and D.W. Kelley. 1985. The decline of striped bass
5465 in the Sacramento-San Joaquin Estuary, California. *Transactions of the American Fisheries*
5466 *Society* 114:12–30.
- 5467 Stewart, A.R., S.N. Luoma, C.E. Schlekat, M.A. Doblin, and K.A. Hieb. 2004. Food web
5468 pathway determines how selenium affects aquatic ecosystems: A San Francisco Bay case
5469 study. *Environmental Science and Technology* 38:4519–4526.
- 5470 Stoms, D.M. 2010. Change in urban land use and associated attributes in the Upper San
5471 Francisco Estuary, 1990–2006. *San Francisco Estuary and Watershed Science*. Vol. 8, Issue
5472 3.
- 5473 Suddeth, R.J., J. Mount, and J.R. Lund. 2010. Levee decisions and sustainability for the
5474 Sacramento-San Joaquin Delta. *San Francisco Estuary and Watershed Science* Vol. 8. Issue
5475 2. Retrieved from: <http://www.escholarship.org/uc/item/9wr5j84g>
- 5476 Swain, D.P., A.F. Sinclair, and J.M. Hanson. 2007. Evolutionary response to size-selective
5477 mortality in an exploited fish population. *Proceedings of the Royal Society London B*.
5478 274:1–8.
- 5479 Swanson, C., T. Reid, P.S. Young, and J.J. Cech, Jr. 2000. Comparative environmental
5480 tolerances of threatened delta smelt (*Hypomesus transpacificus*) and introduced wakasagi (*H.*
5481 *nipponensis*) in an altered California estuary. *Oecologia* 123:384–390.
- 5482 Sweetnam, D.A. 1999. Status of delta smelt in the Sacramento-San Joaquin Estuary. *California*
5483 *Fish and Game* 85:22–27.
- 5484 Teh, S.J. 2007. Final report of histopathological evaluation of starvation and/or toxic effects on
5485 pelagic fishes: pilot study of the health status of 2005 adult delta smelt in the upper San
5486 Francisco Estuary. Report to the POD Management Team, U.C. Davis, Davis, CA.
- 5487 Teh, S.J., D.V. Baxa, and S. Acuña. 2010. Effects of *Microcystis aeruginosa* in threadfin shad
5488 (*Dorosoma petenense*). Final Report to the Department of Water Resources submitted
5489 November 10, 2010. 87 pp.
- 5490 The Bay Institute. 1998. *From the Sierra to the sea*. The Bay Institute of San Francisco, San
5491 Rafael, CA.
- 5492 The Bay Institute, Center for Biological Diversity, and Natural Resources Defense Council.
5493 2007a. Petition to the State of California Fish and Game Commission and supporting

5494 information for listing the longfin smelt (*Spirinchus thaleichthys*) as an endangered species
5495 under the California Endangered Species Act. Submitted to the California Fish and Game
5496 Commission, 1416 Ninth Street, Sacramento, California.

5497 The Bay Institute, Center for Biological Diversity, and Natural Resources Defense Council.
5498 2007b. Petition to list the San Francisco Bay-Delta population of longfin smelt (*Spirinchus*
5499 *thaleichthys*) as endangered under the Endangered Species Act. Submitted to the U.S. Fish
5500 and Wildlife Service, Sacramento Fish and Wildlife Office, 2800 Cottage Way, Room W-
5501 2605, Sacramento, California.

5502 The Johnson Foundation. 2010. Charting new waters: a call to action to address U.S. freshwater
5503 challenges. Issued by the Participants of The Johnson Foundation Freshwater Summit held
5504 June 9, 2010, see <http://www.johnsonfdn.org/chartingnewwaters/download-reports?sid=1131>

5505 Thomson, J.R., W.J. Kimmerer, L.R. Brown, K.B. Newman, R. Mac Nally, W.A. Bennett, F.
5506 Feyrer, and E. Fleishman. 2010. Bayesian change-point analysis of abundance trends for
5507 pelagic fishes in the upper San Francisco Estuary. *Ecological Applications* 20:181–198.

5508 Trent, L. and W.W. Hassler. 1968. Gill net selection, migration, size and age composition, sex
5509 ratio, harvest efficiency, and management of striped bass in the Roanoke River, North
5510 Carolina. *Chesapeake Science* 9:217–232.

5511 Turner, J.L. 1966. Distribution of threadfin shad, *Dorosoma petenense*; tule perch,
5512 *Hysterochampus traski*; sculpin spp. and crayfish spp. in the Sacramento-San Joaquin Delta.
5513 Pages 160–168 in J.L. Turner and D.W. Kelley, compilers. *Ecological Studies of the*
5514 *Sacramento-San Joaquin Estuary. Part II.* California Department of Fish and Game, Fish
5515 Bulletin 136.

5516 Uphoff, J.H., Jr. 2003. Predator-prey analysis of striped bass and Atlantic menhaden in upper
5517 Chesapeake Bay. *Fisheries Management and Ecology* 10:313–322.

5518 USFWS (U.S. Fish and Wildlife Service). 2008. Formal Endangered Species Act Consultation
5519 on the Proposed Coordinated Operations of the Central Valley Project (CVP) and State
5520 Water Project (SWP). U.S. Fish and Wildlife Service, Sacramento, CA.

5521 USFWS (U.S. Fish and Wildlife Service). 2009. Endangered and threatened wildlife and plants;
5522 12-month finding on a petition to list the San Francisco Bay-Delta population of the longfin
5523 smelt (*Spirinchus thaleichthys*) as endangered. *Federal Register* 74:16169–16175.

5524 USFWS (U.S. Fish and Wildlife Service). 2010. Endangered and threatened wildlife and plants;
5525 12-month finding on a petition to reclassify the delta smelt from threatened to endangered
5526 throughout its range. *Federal Register* 75:17667–17680.

5527 Van Den Avyle, M.J., J. Boxrucker, B. Vondracek, and G.R. Ploskey. 1995. Comparison of catch
5528 rate, length distribution, and precision of six gears used to sample reservoir shad populations.
5529 *North American Journal of Fisheries Management* 15:940–955.

5530 Van Nieuwenhuyse, E.E. 2007. Response of summer chlorophyll concentration to reduced total
5531 phosphorus concentration in the Rhine River (Netherlands) and the Sacramento-San Joaquin
5532 Delta (California, USA). *Canadian Journal of Fisheries and Aquatic Sciences* 64:1529–1542.

5533 Vandeputte, M., M. Dupont-Nivet, H. Chavanne and B. Chatain, 2007. A polygenic hypothesis
5534 for sex determination in the European sea bass *Dicentrarchus labrax* *Genetics* 176:1049–
5535 1057.

5536 Waldman, J.R., R.E. Bender and I.I. Wirgin 1998 Multiple population bottlenecks and DNA
5537 diversity in populations of wild striped bass, *Morone saxatilis*. *Fishery Bulletin* 96:614–620.

5538 Werner, I., L.A. Deanovic, V. Connor, V. De Vlaming, H.C. Bailey, and D.E. Hinton. 2000.
5539 Insecticide-caused toxicity to *Ceriodaphnia dubia* (Cladocera) in the Sacramento-San

5540 Joaquin River Delta, California, USA. *Environmental Toxicology and Chemistry* 19:215–
5541 227.

5542 Werner, I., L. Deanovic, D. Markiewicz, M. Stillway, N. Offer, R. Connon, and S. Brander.
5543 2008. Pelagic Organism Decline (POD): Acute and chronic invertebrate and fish toxicity
5544 testing in the Sacramento-San Joaquin Delta 2006–2007. Final Report. U.C. Davis–Aquatic
5545 Toxicology Laboratory, Davis, California.

5546 Werner, I., L.A. Deanovic, D. Markiewicz, J. Khamphanh, C.K. Reece, M. Stillway, and C.
5547 Reece. 2010a. Monitoring acute and chronic water column toxicity in the northern
5548 Sacramento-San Joaquin Estuary, California, USA, using the euryhaline amphipod, *Hyaella*
5549 *azteca*: 2006–2007. *Environmental Toxicology and Chemistry* 29: 2190–2199.

5550 Werner I., L. Deanovic, M. Stillway, and D. Markiewicz. 2009. Acute toxicity of ammonia/um
5551 and wastewater treatment effluent-associated contaminants on delta smelt, Final Report. U.C.
5552 Davis–Aquatic Toxicology Laboratory, Davis, California.

5553 Werner, I. and J.T. Hollibaugh. 1993. *Potamocorbula amurensis*: comparison of clearance rates
5554 and assimilation efficiencies for phytoplankton and bacterioplankton. *Limnology and*
5555 *Oceanography* 38:949–964.

5556 Werner, I., D. Markiewicz, L. Deanovic, R. Connon, S. Beggel, S. Teh, M. Stillway, C. Reece.
5557 2010b. Pelagic Organism Decline (POD): Acute and chronic invertebrate and fish toxicity
5558 testing in the Sacramento-San Joaquin Delta 2008–2010, Final Report. U.C. Davis–Aquatic
5559 Toxicology Laboratory, Davis, California.

5560 Weston, D.P., R.W. Holmes, and M.J. Lydy. 2009. Residential runoff as a source of pyrethroid
5561 pesticides to urban creeks. *Environmental Pollution* 157:287–294.

5562 Weston, D.P., R.W. Holmes, J. You, and M.J. Lydy. 2005. Aquatic toxicity due to residential use
5563 of pyrethroid insecticides. *Environmental Science and Technology* 39:9778–9784.

5564 Weston, B.P. and M.J. Lydy. 2010. Urban and agricultural sources of pyrethroid insecticides to
5565 the Sacramento-San Joaquin Delta of California. *Environmental Science and Technology*
5566 44:1833–1840.

5567 Weston, D.P., J. You, and M.J. Lydy. 2004. Distribution and toxicity of sediment-associated
5568 pesticides in agriculture-dominated water bodies of California’s Central Valley.
5569 *Environmental Science and Technology* 38:2752–2759.

5570 Whitehead, A., K.M. Kuivila, J.L. Orlando, S. Kotelevtsev, and S.L. Anderson. 2004.
5571 Genotoxicity in native fish associated with agricultural runoff events. *Environmental*
5572 *Toxicology and Chemistry* 23:2868–2877.

5573 Wilcox, R.J., P.D. Champion, J.W. Nagels, and G.F. Croker. 1999. The influence of aquatic
5574 macrophytes on the hydraulic and physico-chemical properties of a New Zealand lowland
5575 stream. *Hydrobiologia* 416:203–214.

5576 Wilkerson F.P., R.C. Dugdale, V.E. Hogue, and A. Marchi. 2006. Phytoplankton blooms and
5577 nitrogen productivity in San Francisco Bay. *Estuaries and Coasts* 29:401–416.

5578 Winder, M. and A.D. Jassby. 2010. Shifts in zooplankton community structure: implications for
5579 food web processes in the upper San Francisco Estuary. *Estuaries and Coasts* DOI
5580 10.1007/s12237-010-9342-x.

5581 Winemiller, K.O. 2005. Life history strategies, population regulation, and implications for
5582 fisheries management. *Canadian Journal of Fisheries and Aquatic Sciences* 62:872–885.

5583 Winemiller, K.O. and K.A. Rose. 1992. Patterns of life-history diversification in North American
5584 fishes: implications for population regulation. *Canadian Journal of Fisheries and Aquatic*
5585 *Sciences* 49:2196–2218.

- 5586 Wright, S.A. and D.H. Schoellhamer. 2004. Trends in the sediment yield of the Sacramento
5587 River, California, 1957–2001. San Francisco Estuary and Watershed Science. Vol 2., Issue 2,
5588 Article 2. Available at: [http://repositories/cdlib.org/jmie/sfews/vol2/iss2/art2.p](http://repositories.cdlib.org/jmie/sfews/vol2/iss2/art2.p)
5589 Yang, S.L. 1998. The role of a *Scirpus* marsh in attenuation of hydrodynamics and retention of
5590 fine sediment in the Yangtze Estuary. Estuarine, Coastal, and Shelf Science 47:227–233.
5591 York, J.K., B.A. Costas and G.B. McManus. 2010. Microzooplankton grazing in green water—
5592 results from two contrasting estuaries. Estuaries and Coasts DOI 10.1007/s12237-010-9336-8
5593

5594 **Manuscripts reviewed by POD-MT or IEP-MT but not yet accepted by a journal:**
5595

- 5596 Brooks, M.L. Brooks, E. Fleishman, L.R. Brown, P.H. Lehman, I. Werner, N. Scholz, C.
5597 Mitchelmore, A.E. Parker, D.M. Stoms, M.L. Johnson, J. Drever, R. Dugdale, D. Schlenk,
5598 S. Teh, and S. van Drunick. In review. Sensitive lifestage and the potential contributions of
5599 contaminants to the decline of pelagic fishes in the San Francisco Estuary, California, USA.
5600 Environmental Management.
5601 Connon R.E., S. Beggel, L.S. D’Abronzio, J. Geist, A.S. Loguinov, C.D. Vulpe, and I. Werner. In
5602 review. Molecular biomarkers in endangered species: neuromuscular impairments following
5603 sublethal copper exposures in the delta smelt (*Hypomesus transpacificus*). Environmental
5604 Toxicology and Chemistry.
5605 Connon, R.E. and I. Werner. In revision. Endocrine, neurological and behavioral responses to
5606 sublethal pyrethroid exposure in the endangered delta smelt, *Hypomesus transpacificus*
5607 (Fam. Osmeridae). Marine Environmental Research.
5608 Hobbs J., L. Lewis, N. Ikemiyagi, R. Baxter, and T. Sommer. In revision. Identifying critical
5609 nursery habitat for an estuarine fish with otolith strontium isotopes. Environmental Biology
5610 of Fishes- Special Issue 4th International Symposium for Fish Otolith Research and
5611 Application.
5612 Loboschfsky, E., G. Benigno, T. Sommer, T. Ginn, A. Massoudieh, K. Rose, and F. Loge. In
5613 review. Bioenergetic Modeling of San Francisco Estuary Striped Bass. San Francisco
5614 Estuary and Watershed Science.
5615 Parker, A.E., R.C. Dugdale, and F. P. Wilkerson. In review. Biogeochemical processing of
5616 anthropogenic ammonium in the Sacramento River and the northern San Francisco Estuary.
5617 Hydrobiologia.
5618 Scholz, N.L., E. Fleishman, L. Brown, I. Werner, M.L. Johnson, M.L. Brooks, C.L.
5619 Mitchelmore, and D. Schlenk. In review. Pesticides and the decline of pelagic fishes in
5620 western North America’s largest estuarine ecosystem. Conservation Letters.
5621 Sommer, T., M. Nobriga, L. Grimaldo, F. Feyrer, and F. Mejia. In review. The spawning
5622 migration of delta smelt in the upper San Francisco Estuary. Marine and Coastal Fisheries.
5623 Sommer, T.R., F. Mejia, K. Hieb, R. Baxter, E.J. Loboschfsky and F.J. Loge. Submitted. Long-
5624 term shifts in the lateral distribution of age-0 striped bass *Morone saxatilis* in the San
5625 Francisco estuary. Transactions of the American Fisheries Society.
5626 Wagner, R.W., M. Stacey, L.R. Brown, and M. Dettinger. Statistical models of temperature in
5627 the Sacramento-San Joaquin Delta under climate-change scenarios and ecological
5628 implications. Estuaries and Coasts.
5629 Werner, I., L.A. Deanovic, D. Markiewicz, J. Khamphanh, C.K. Reece, M. Stillway, C. Reece.
5630 In revision. Monitoring water column toxicity in the Sacramento-San Joaquin Delta,

5631 California, USA, using the euryhaline amphipod, *Hyalella azteca*: 2006–2007. Integrated
5632 Environmental Assessment and Monitoring.
5633

5634
5635

Tables

5636 Table 1. Linkages of POD study elements (PEN = POD element number) to the drivers of the
5637 basic conceptual model (A=previous abundance, H=habitat, B=bottom up, T=top down),
5638 species-specific conceptual models (DS=delta smelt, LFS=longfin smelt, SB=striped bass,
5639 TFS=threadfin shad), season when results of the element apply, and contribution to
5640 understanding the regime shift conceptual model (B=benthic, P=pelagic, L=littoral).
5641

POD study element	PEN	Basic conceptual model	Species-specific models	Seasons	Regime shift conceptual model
2. Ongoing Work					
Development and implementation of IBM of striped bass and longfin smelt	38	All	SB, LFS	All	P
Modeling delta smelt in the S.F. Estuary	41	All	DS	All	P
Estimation of pelagic fish population sizes	43	All	All	All	P
Zooplankton fecundity and population structure	44	B	DS, LFS, SB	All	B, P
Phytoplankton primary production and biomass	45	H, B	All	All	B, P
NCEAS - synthetic analyses of fish and zooplankton	46	All	All	All	All
Evaluate delta smelt otolith microstructure	60	All	DS	All	P
Delta smelt histopathology investigations	61	H, B	DS	All	P
Fish diet and condition	62	B	All	Su	P
Trends in benthic macrofauna abundance and biomass	65	B	All	All	B
Corbula salinity tolerance	76	H, B	All	All	B
Field survey of <i>Microcystis</i> bloom biomass and toxicity	79	H, B	All	Su, F	P
Food web support for delta smelt and other estuarine fishes	82	B	DS	All	P, B
Investigation of power plant impacts	87	A, T	DS, LFS	Sp, Su, F	P
SAV abundance and distribution	102	H	All	Sp, Su, F	L
Fish facility history	107	T, A	All	All	P
Delta smelt culture facility	108a/b	A, H	DS	All	
Striped bass bioenergetics	115	T, A	SB	All	P, L
Long-term sources and early warning signals in turbidity	126	H	DS	W	P
Contaminants and biomarkers work	127	H, B	All	Sp, Su	All
Feasibility of using towed imaging systems	130	A	All	Sp, F, W	P
Use of acoustics to measure trawl openings	131	A	All	F	P
Effects of the Cache Slough complex on north Delta habitat	132	H, B	All	All	P
Impacts of largemouth bass on the Delta	133	T	All	All	L, P
Delta smelt genetics	135	A	DS	All	P
Bioenergetics of zooplankton species	136	B	DS	Spring	P
Population genetics and otolith geochemistry of longfin smelt	137	A	LFS	All	P
Effects of waste water management on primary productivity	138	B	All	All	P

Effects of <i>Microcystis</i> on threadfin shad	139	B, H	TFS	Su	P
Mark-recapture to estimate delta smelt pre-screen loss and salvage efficiency	140	T	DS	Sp	P
3-D modeling of the Delta	141	H	DS	All	
Contaminant synthesis 2 – impacts of contaminants and discharges	146	H	All	All	All
BREACH III: Evaluating and predicting restoration thresholds	147	H, B	DS, SB, TFS	All	L, P
Spatial and temporal variability of Delta water temperatures	148	H	All	All	P
Plankton dynamics in the Delta: trends and interactions	150	B	All	All	P
Environmental controls on the distribution of harmful algae and their toxins in the San Francisco Bay	152	H, B	All	All	P
Comparison of nutrient sources and phytoplankton growth and species composition	153	H,B	All	All	P
Spatial and temporal quantification of pesticide loadings	154	H	All	All	All
3. New IEP, new CALFED/Delta Science Program or expanded IEP work					
Acute and chronic toxicity of contaminant mixtures and multiple stressors	157	H, B	All		P
Advancing procedures for extracting and recovering chemicals of concern from sediment interstitial water	158	H	All		B
Investigation of pyrethroid pesticides in the American River	159	H, B		W	P, B
Full life-cycle bioassay approach to assess chronic exposure of <i>Pseudodiaptomus forbesi</i> to ammonia	160	H, B	DS	Su, F	P
SRWTP effluent toxicity testing with delta smelt and rainbow trout	161	H	DS	All	P
Potential loss of life history variation and the decline of delta smelt	162	A	DS	All	P
Comparison of 1- and 2-D hydrodynamic and water quality models of the Delta	163	H	All	All	P
Spatial and temporal variability in nutrients in Suisun Bay in relation to spring phytoplankton blooms	164	B	DS, LFS, SB	Sp	P
Ammonia sampling for the Sacramento-San Joaquin Delta and Estuary	165	H	DS	Sp	P
Using PCR to detect sliverside predation on larval delta smelt	166	T	DS	Sp	P, L
Investigation of presence, migration patterns and site fidelity of sub-adult striped bass	167	T	All	All	P, L
Monitoring inter-annual variability of delta smelt population contingents and growth	168	A, T	DS	All	P
Delta smelt feeding and foodweb interactions	169	B	DS	Fall	P
Experimentally determining early life-stage sensitivity to salinity for longfin smelt	170	H	LFS	All	P
Remote sensing mapping and monitoring of <i>Microcystis</i> and turbidity in the upper SFE	171	H	All	Su	P
The role of pyrethroid insecticides in limiting prey availability for delta smelt in the north Delta	172	B, H	DS	W, Sp	P
Distribution, concentrations and fate of ammonium in the Sacramento River and the low salinity zone	173	B, H	DS, SB	Sp, Su, F	P
Influence of elevated ammonium (NH ₄) on phytoplankton physiology in the SFE during fall	174	B, H	DS	F	P
Effect of seasonal variations in flow on the spatial and temporal variations of nutrients, organic matter and	175	B, H	DS	F	P

phytoplankton					
Influence of water quality and SAV on largemouth bass distribution, diet composition and predation on delta smelt	176	T, H	DS, SB, TFS	All	L
Metabolic responses to variable salinity environments in field acclimatized <i>Corbula amurensis</i>	177	B, H	DS, SB, LFS	All	B, P
Bivalve effects on the food web supporting delta smelt	178	B	All	All	B, P
Causes of seasonal and spatial variations in NH ₄ sources, sinks, and contributions to algal productivity using a multi-isotopic approach	179	B, H	All	All	P
Hydrodynamic and particle tracking modeling of delta smelt habitat and prey	180	H, B	All	All	
Longfin smelt bioenergetics	181	A, H	LFS	All	P
Development of an acoustic transmitter suitable for use in delta smelt	182		DS	F, W, Sp	P
Novel molecular and biochemical biomarker work	183	H	DS	All	All
Disease and physiology monitoring in wild delta smelt adults	184	H	DS	Winter	P
Physical processes influencing spawning migrations of delta smelt	187	H	DS	Winter	P
OP and pyrethroid use in the Sacramento River and Delta	188	H	All	Spring, summer	All
Ammonia literature review	189	H, B	All	All	All

5642

5643

5644

5645 Table 2. Costs and funding sources for individual POD study elements.
 5646

Element name, program element number (PEN) and estimated budget amounts by funding source for 2010 POD work plan elements (amounts are in \$1,000).

1. Expanded Monitoring							
	PEN	POD Total	DWR POD	USBR POD	CALFED ERP or Delta Science	SWRCB POD	Other
Fall midwater trawl	3	\$22		\$22			
Summer townet survey	7	\$24		\$24			
20-mm survey	33						
Spring Kodiak trawl	88						
Directed field collections	89	\$68		\$68			
Smelt larvae survey	96	\$269		\$269			
TOTAL for EXPANDED MONITORING		\$383		\$383			

2. Ongoing Work							
	PEN	POD Total	DWR POD	USBR POD	CALFED ERP or Delta Science	SWRCB POD	Other
Development and implementation of IBM of striped bass and longfin smelt	38	\$125	\$125				
Modeling delta smelt in the S.F. Estuary	41						
Estimation of pelagic fish population sizes	43	\$188	\$94	\$94			
Zooplankton fecundity and population structure	44	\$41	\$41				
Phytoplankton primary production and biomass	45	\$35	\$35				
NCEAS - synthetic analyses of fish and zooplankton	46	\$751	\$302	\$449			
Evaluate delta smelt otolith microstructure	60	\$292			\$292		
Delta smelt histopathology investigations	61	\$292			\$292		
Fish diet and condition	62	\$40		\$40			
Trends in benthic macrofauna abundance and biomass	65	\$40			\$40		
Corbula salinity tolerance	76						
Field survey of <i>Microcystis</i> bloom biomass and toxicity	79	\$144			\$144		
Food web support for delta smelt and other estuarine fishes	82	\$162			\$162		
Investigation of power plant impacts	87	\$25	\$25				
SAV abundance and distribution	102						
Fish facility history	107	\$50					\$50
Delta smelt culture facility	108a/b	\$2,284	\$48	\$592			\$1,644
Striped bass bioenergetics	115	\$30	\$30				
Long-term sources and early warning signals in turbidity	126						
Contaminants and biomarkers work	127	\$453	\$453				
Feasibility of using towed imaging systems	130	\$201	\$41	\$160			
Use of acoustics to measure trawl openings	131	\$13		\$13			

Effects of the Cache Slough complex on north Delta habitat	132	\$334	\$116	\$218			
2. Ongoing Work (continued)							
	PEN	POD Total	DWR POD	USBR POD	CALFED ERP or Delta Science	SWRCB POD	Other
Impacts of largemouth bass on the Delta	133	\$239	\$61	\$178			
Delta smelt genetics	135						
Feeding and Growth of Delta Smelt	136	\$17	\$17				
Population genetics and otolith geochemistry of longfin smelt	137	\$113		\$113			
Effects of waste water management on primary productivity	138	\$119				\$119	
Effects of <i>Microcystis</i> on threadfin shad	139	\$178	\$178				
Mark-recapture to estimate delta smelt pre-screen loss and salvage efficiency	140	\$15			\$15		
3-D modeling of the Delta	141						
Contaminant synthesis 2 – impacts of contaminants and discharges	146	\$141				\$141	
BREACH III: Evaluating and predicting restoration thresholds	147	\$1,100			\$1,100		
Spatial and temporal variability of Delta water temperatures	148						
Plankton dynamics in the Delta: trends and interactions	150	\$83			\$83		
Environmental controls on the distribution of harmful algae and their toxins in the San Francisco Bay	152	\$82			\$82		
Comparison of nutrient sources and phytoplankton growth and species composition	153	\$338			\$338		
Spatial and temporal quantification of pesticide loadings	154	\$395			\$395		
Contaminants support	186	\$50	\$50				
TOTAL for ONGOING WORK		\$8,370	\$1,616	\$1,857	\$2,943	\$260	\$1,694

3. New IEP, new CALFED/Delta Science Program or expanded IEP work							
	PEN	POD Total	DWR POD	USBR POD	CALFED ERP or Delta Science	SWRCB POD	Other
Acute and chronic toxicity of contaminant mixtures and multiple stressors	157	\$40				\$40	
Advancing procedures for extracting and recovering chemicals of concern from sediment interstitial water	158	\$40				\$40	
Investigation of pyrethroid pesticides in the American River	159	\$100				\$100	
Full life-cycle bioassay approach to assess chronic exposure of <i>Pseudodiaptomus forbesi</i> to ammonia	160	\$77				\$77	
SRWTP effluent toxicity testing with delta smelt and rainbow trout	161	\$65				\$65	
Potential loss of life history variation and the decline of delta smelt	162	\$32				\$32	
Comparison of 1- and 2-D hydrodynamic and water quality models of the Delta	163	\$59				\$59	
Spatial and temporal variability in nutrients in Suisun Bay in relation to spring phytoplankton blooms	164	\$25				\$25	
Ammonia sampling for the Sacramento-San Joaquin Delta and Estuary	165	\$68				\$68	

3. New IEP, new CALFED/Delta Science Program or expanded IEP work (continued)

	PEN	POD Total	DWR POD	USBR POD	CALFED ERP or Delta Science	SWRCB POD	Other
Using PCR to detect sliverside predation on larval delta smelt	166	\$68	\$68				
Investigation of presence, migration patterns and site fidelity of sub-adult striped bass	167	\$75	\$75				
Monitoring inter-annual variability of delta smelt population contingents and growth	168	\$98		\$98			
Delta smelt feeding and foodweb interactions	169	\$400		\$400			
Experimentally determining early life-stage sensitivity to salinity for longfin smelt	170	\$70		\$70			
Remote sensing mapping and monitoring of Microcystis and turbidity in the upper SFE	171	\$134		\$134			
The role of pyrethroid insecticides in limiting prey availability for delta smelt in the north Delta	172	\$158		\$158			
Distribution, concentrations and fate of ammonium in the Sacramento River and the low salinity zone	173	\$77		\$77			
Influence of elevated ammonium (NH4) on phytoplankton physiology in the SFE during fall	174	\$114		\$114			
Effect of seasonal variations in flow on the spatial and temporal variations of nutrients, organic matter and phytoplankton	175	\$42		\$42			
Influence of water quality and SAV on LMB distribution, diet composition and predation on delta smelt	176	\$173		\$173			
Metabolic responses to variable salinity environments in filed acclimatized Corbula amurensis	177	\$137		\$137			
Bivalve effects on the food web supporting delta smelt	178	\$89		\$89			
Causes of seasonal and spatial variations in NH4 sources, sinks, and contributions to algal productivity using a multi-isotopic approach	179	\$242		\$242			
Hydrodynamic and particle tracking modeling of delta smelt habitat and prey	180	\$339		\$339			
Longfin smelt bioenergetics	181	\$128		\$128			
Development of an acoustic transmitter suitable for use in delta smelt	182	\$178		\$178			
Novel molecular and biochemical biomarker work	183	\$220		\$220			
Disease and physiology monitoring in wild delta smelt adults	184						
Physical processes influencing spawning migrations of delta smelt	187						
OP and pyrethroid use in the Sacramento River and Delta	188	\$175				\$175	
Ammonia literature review	189	\$69				\$69	
TOTAL for NEW WORK		\$3,248	\$143	\$2,599		\$750	

2010 POD TOTAL		\$12,025	\$1,698	\$4,839	\$3,019	\$1,010	\$1694
-----------------------	--	-----------------	----------------	----------------	----------------	----------------	---------------

5647

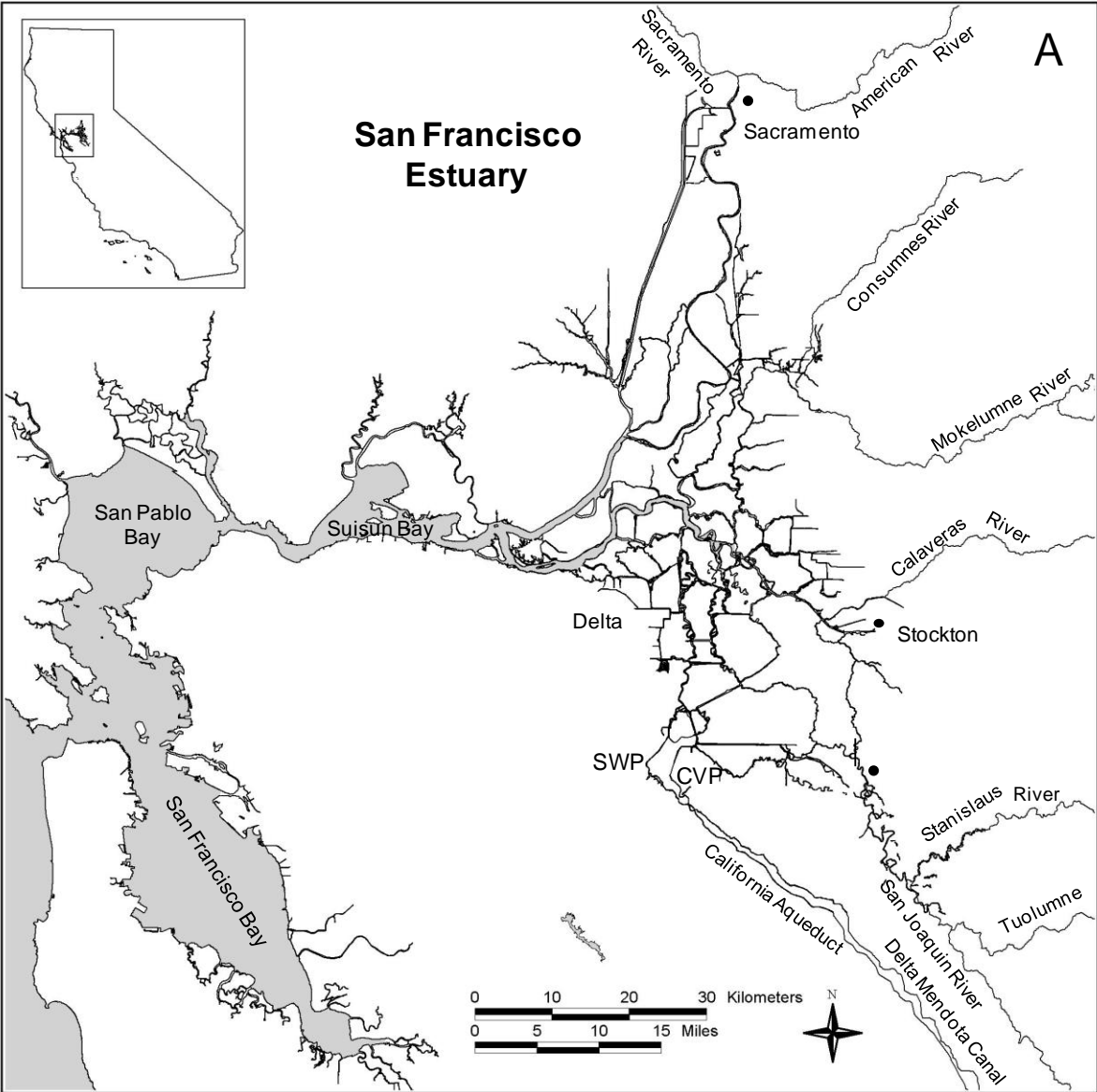
5648

5649
5650

Figures

A

San Francisco Estuary



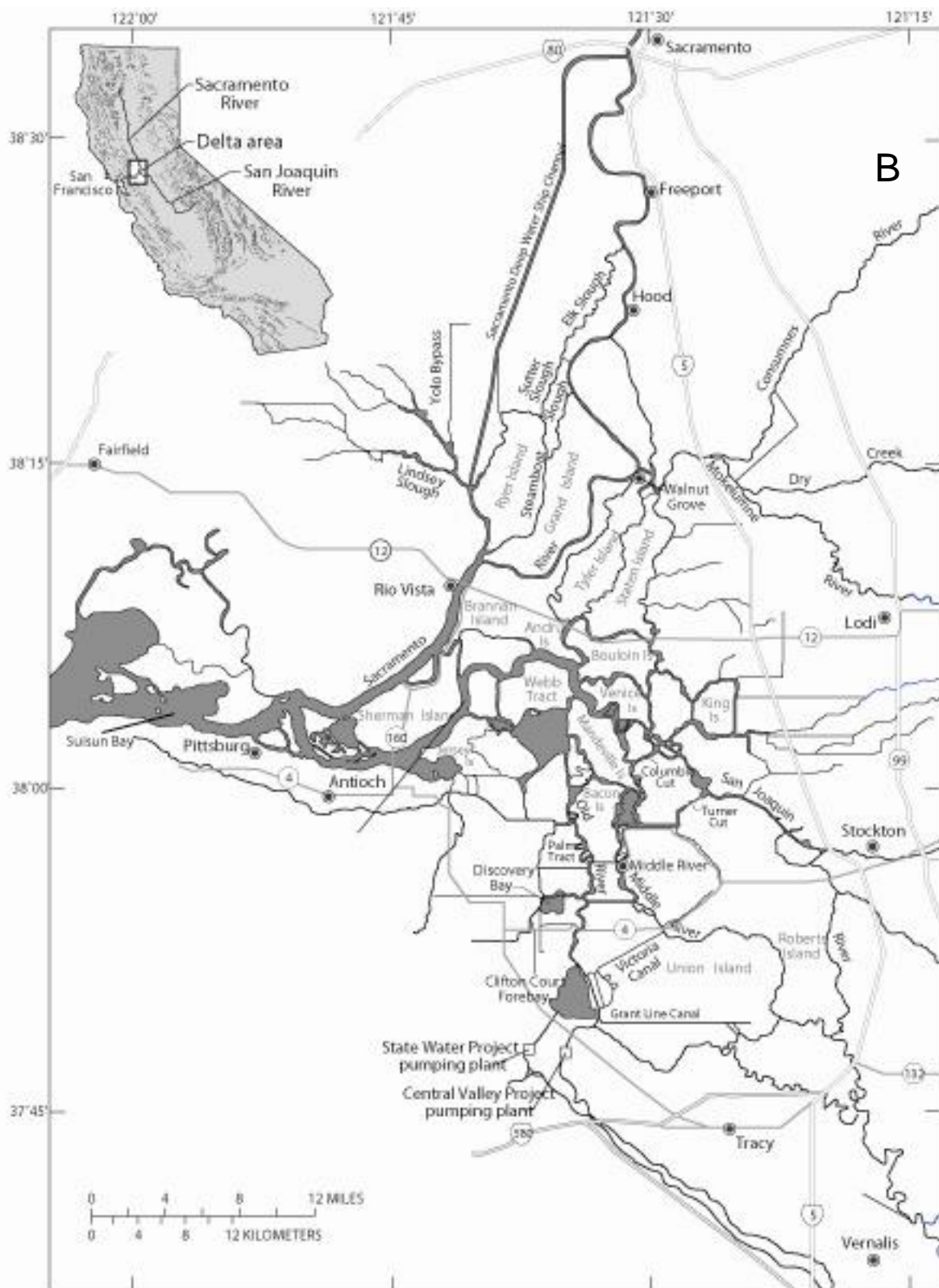
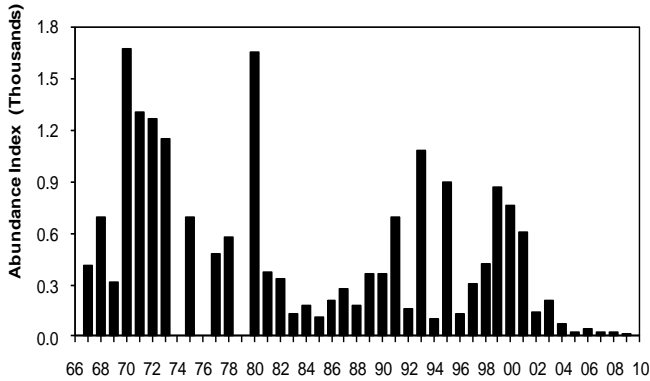
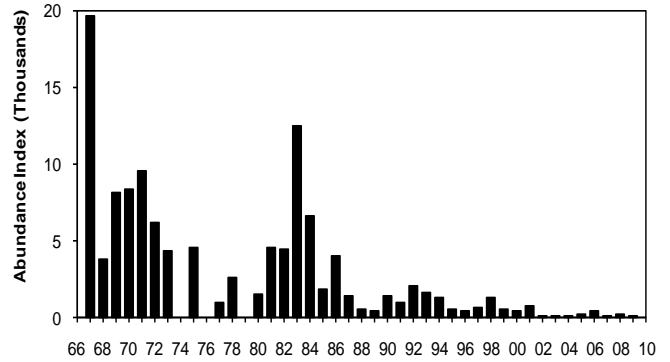


Figure 1. Map of the (A) the entire San Francisco Estuary and (B) the Sacramento-San Joaquin Delta.

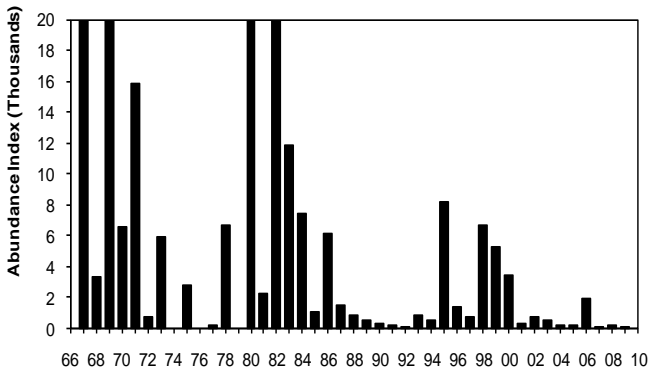
Delta smelt



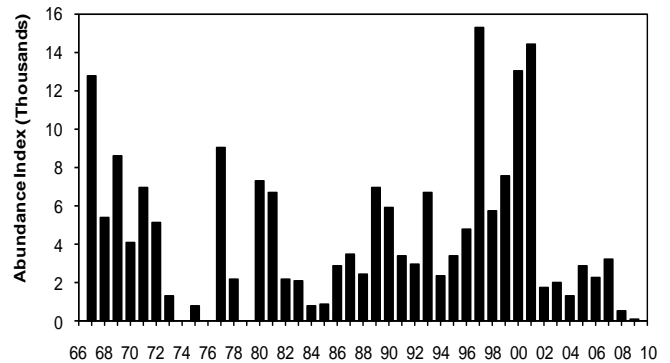
Age-0 striped Bass



Longfin smelt



Threadfin shad



2. Trends in abundance indices for four pelagic fishes from 1967 to 2009 based on the Fall Midwater Trawl, a California Department of Fish and Game survey that samples the upper San Francisco Estuary. No sampling occurred in 1974 or 1979 and no index was calculated for 1976. Development of abundance indices from catch data is described by Stevens and Miller (1983). Note that the y-axis for longfin smelt represents only the lower 25% of its abundance range to more clearly portray the lower abundance range.

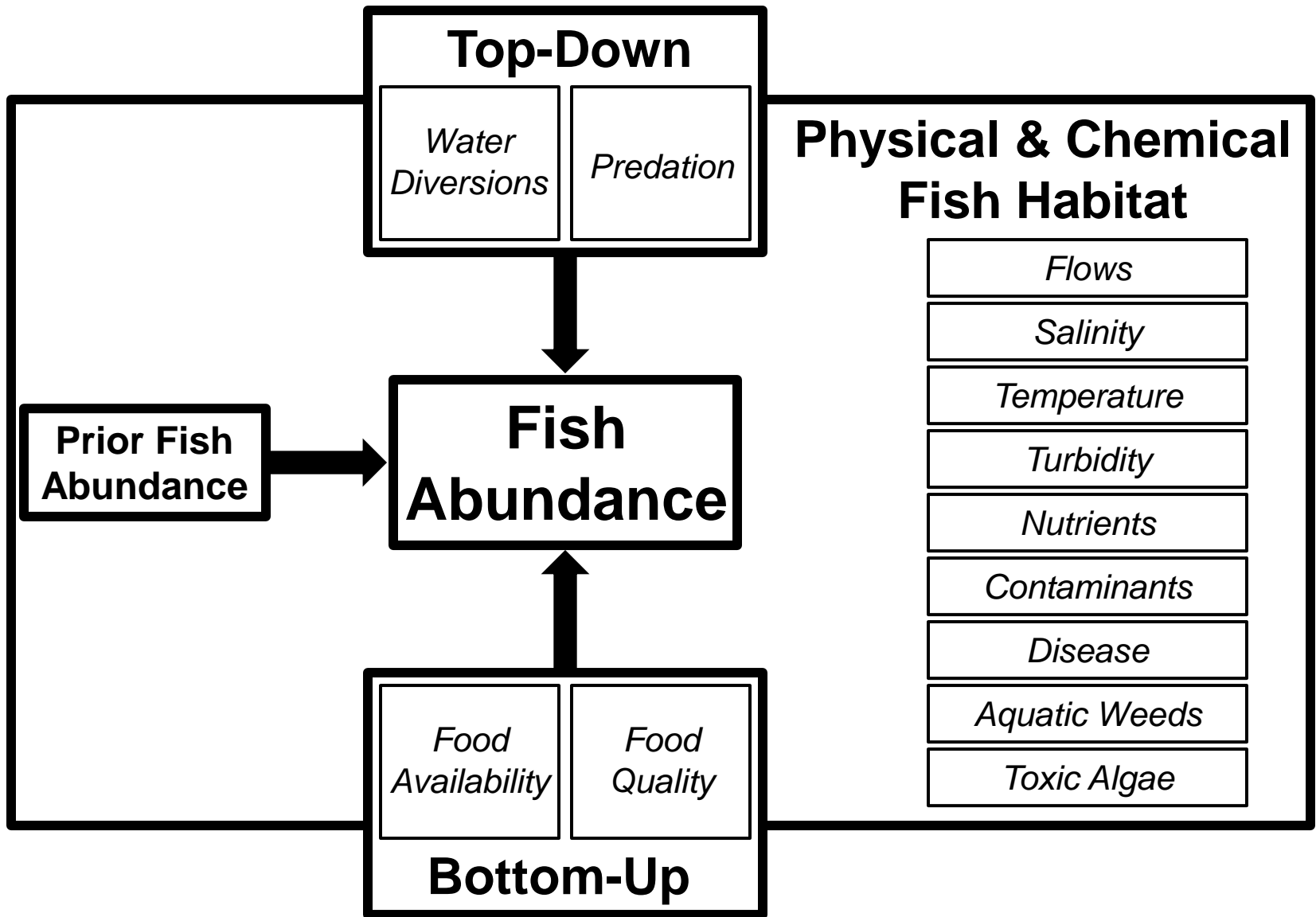


Figure 3. The basic conceptual model for the pelagic organism decline (updated from Sommer et al. 2007).

Figure 4. Delta smelt species model.

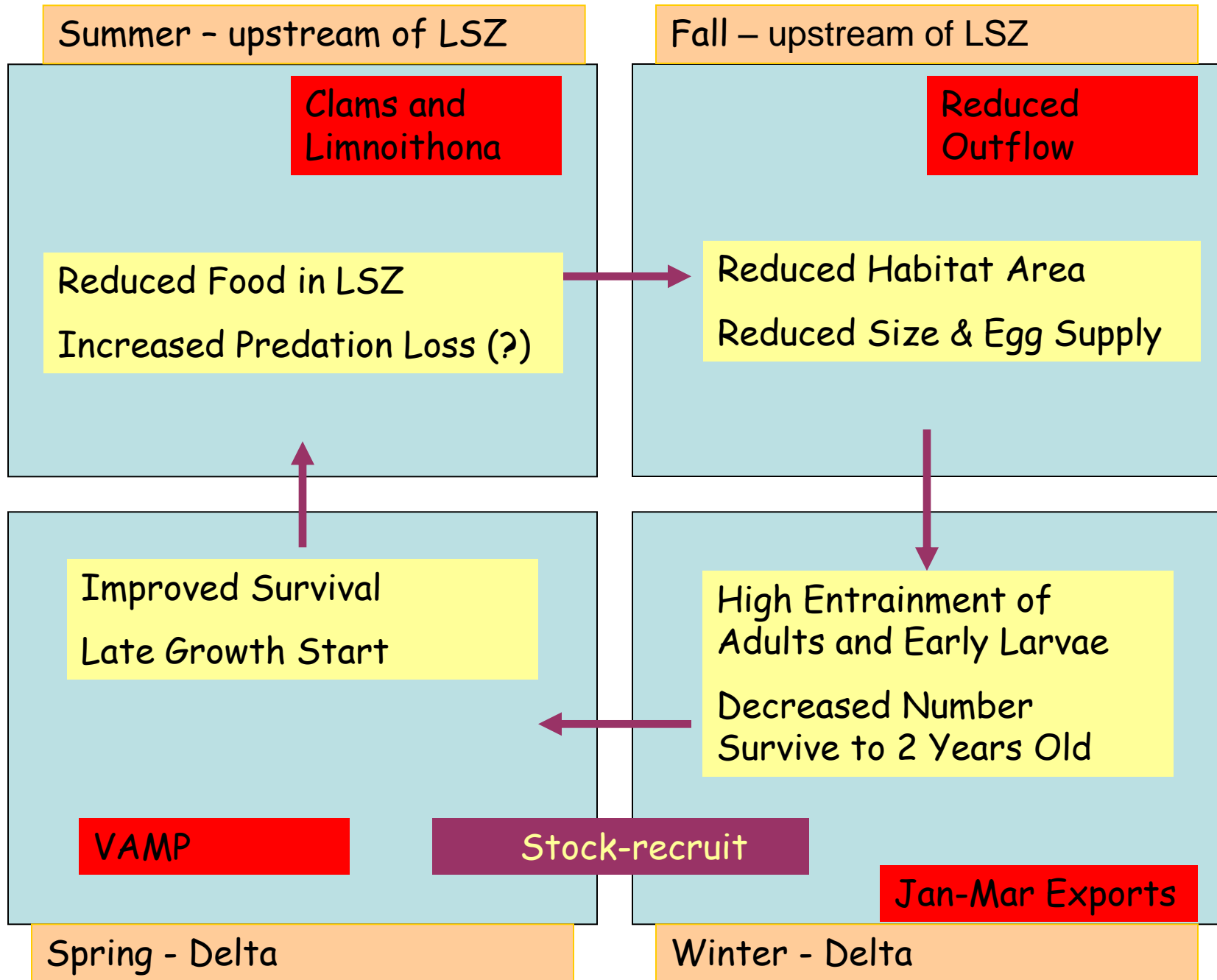


Figure 5. Longfin smelt species model. The dotted line indicates that the importance of a stock recruitment relationship is unclear. The stage recruitment loop illustrates that both survival from age-0 to age-1 and from age-1 to age-2 are important.

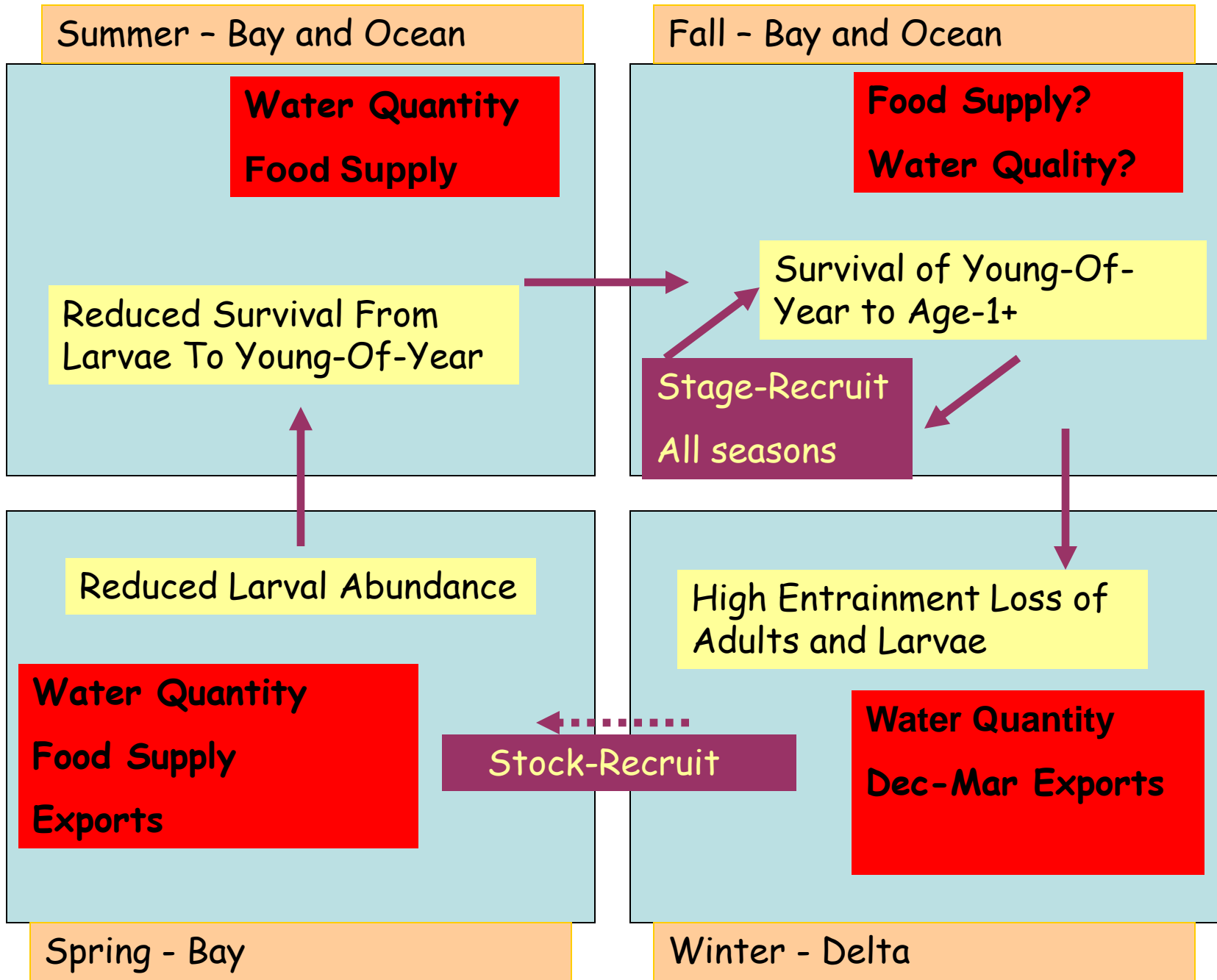


Figure 6. Striped bass species model. The dotted line indicates that the form of the stock recruitment relationship has changed and the present stock-recruitment relationship is unclear.

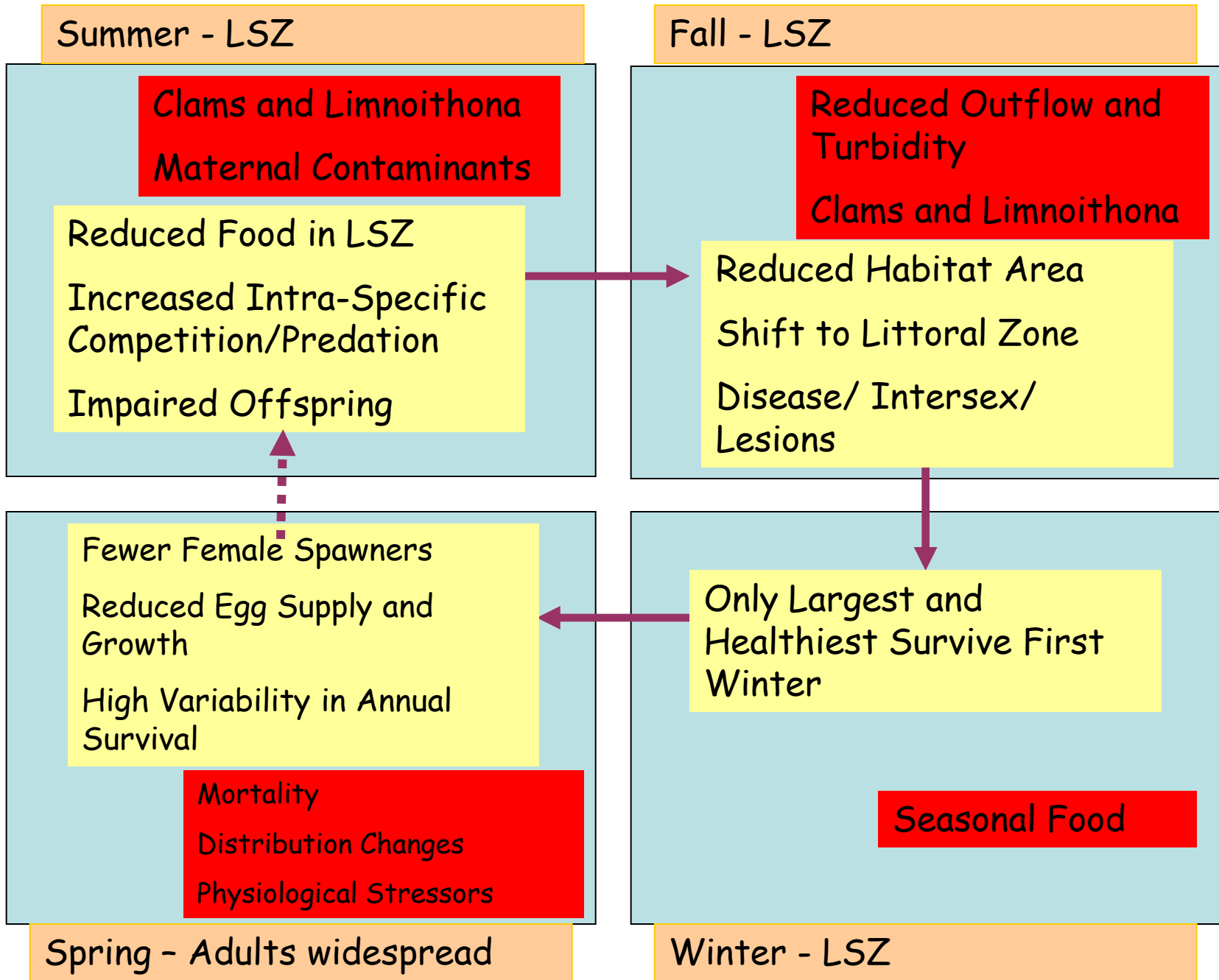
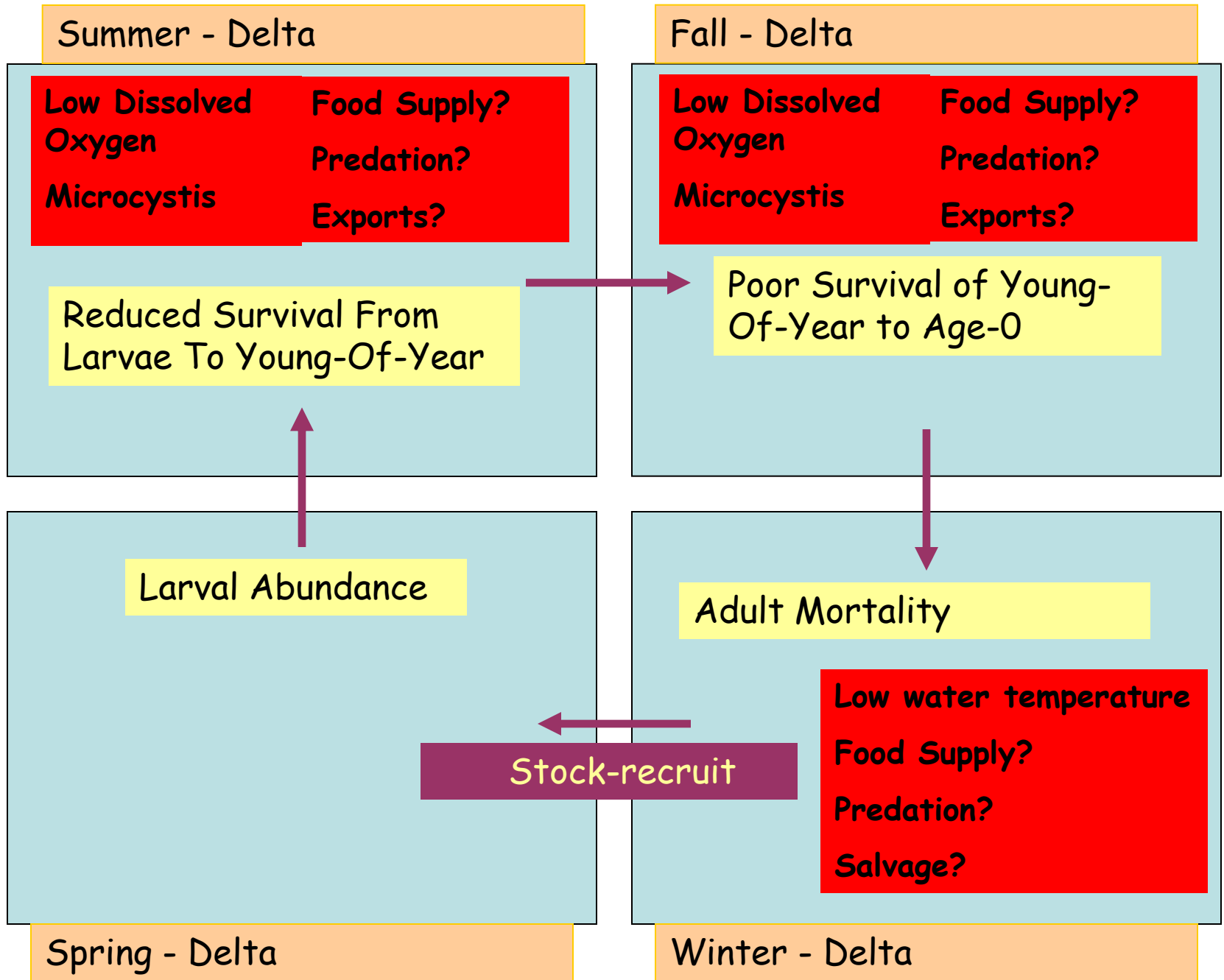


Figure 7. Threadfin shad species model. The dotted line indicates that the importance of a stock-recruitment relationship is unclear.



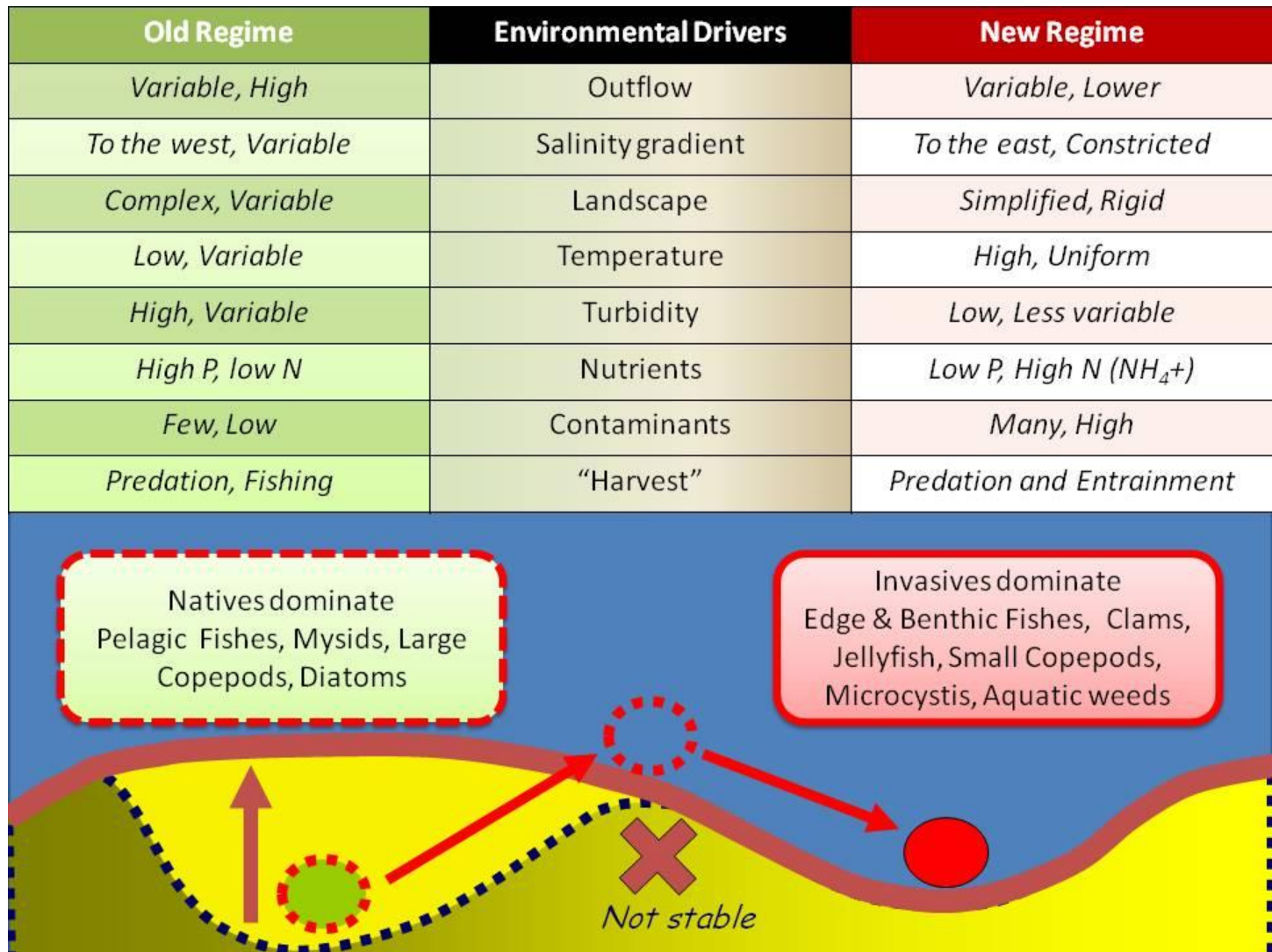


Figure 8. The ecological regime shift in the Delta results from changes in (slow) environmental drivers that lead to profoundly altered biological communities and, as soon as an unstable threshold region is passed, a new relatively stable ecosystem regime.

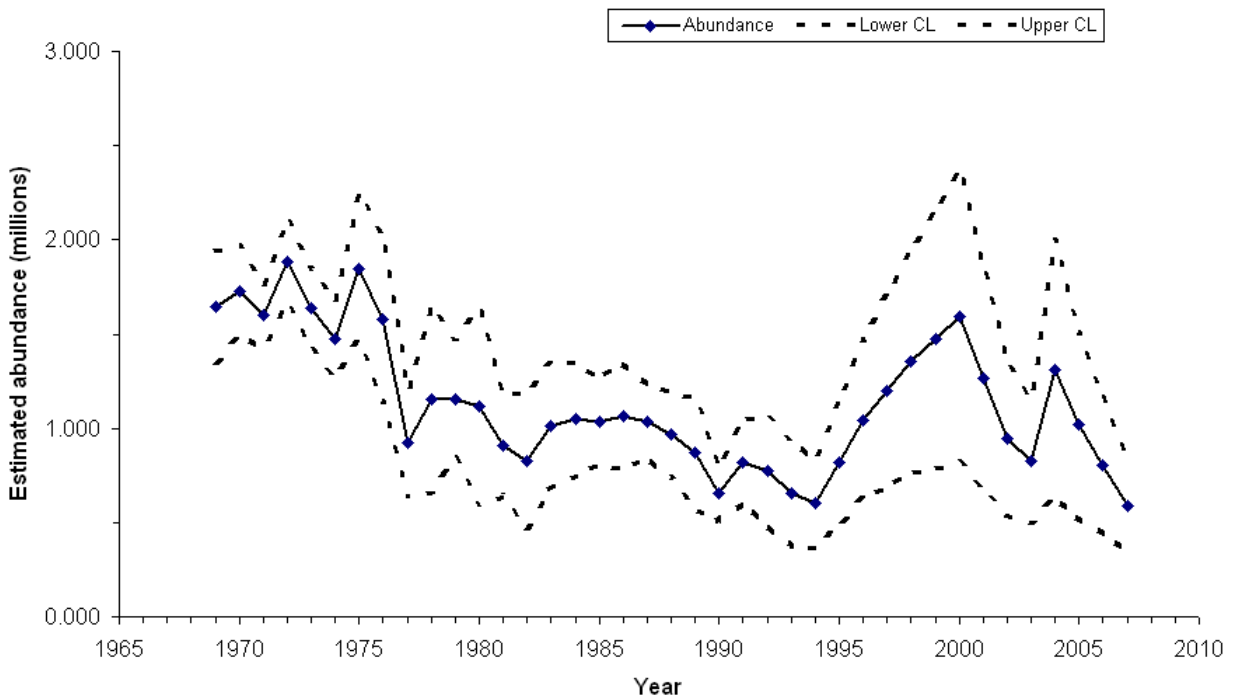


Figure 9. Peterson population estimates of the abundance of adult (3+) striped bass > 460 mm total length from 1969 to 2008, with 95% confidence limits (California Department of Fish and Game, unpublished data). Striped bass were only tagged during even years from 1994 to 2002, so no estimates are available for odd years during that period.

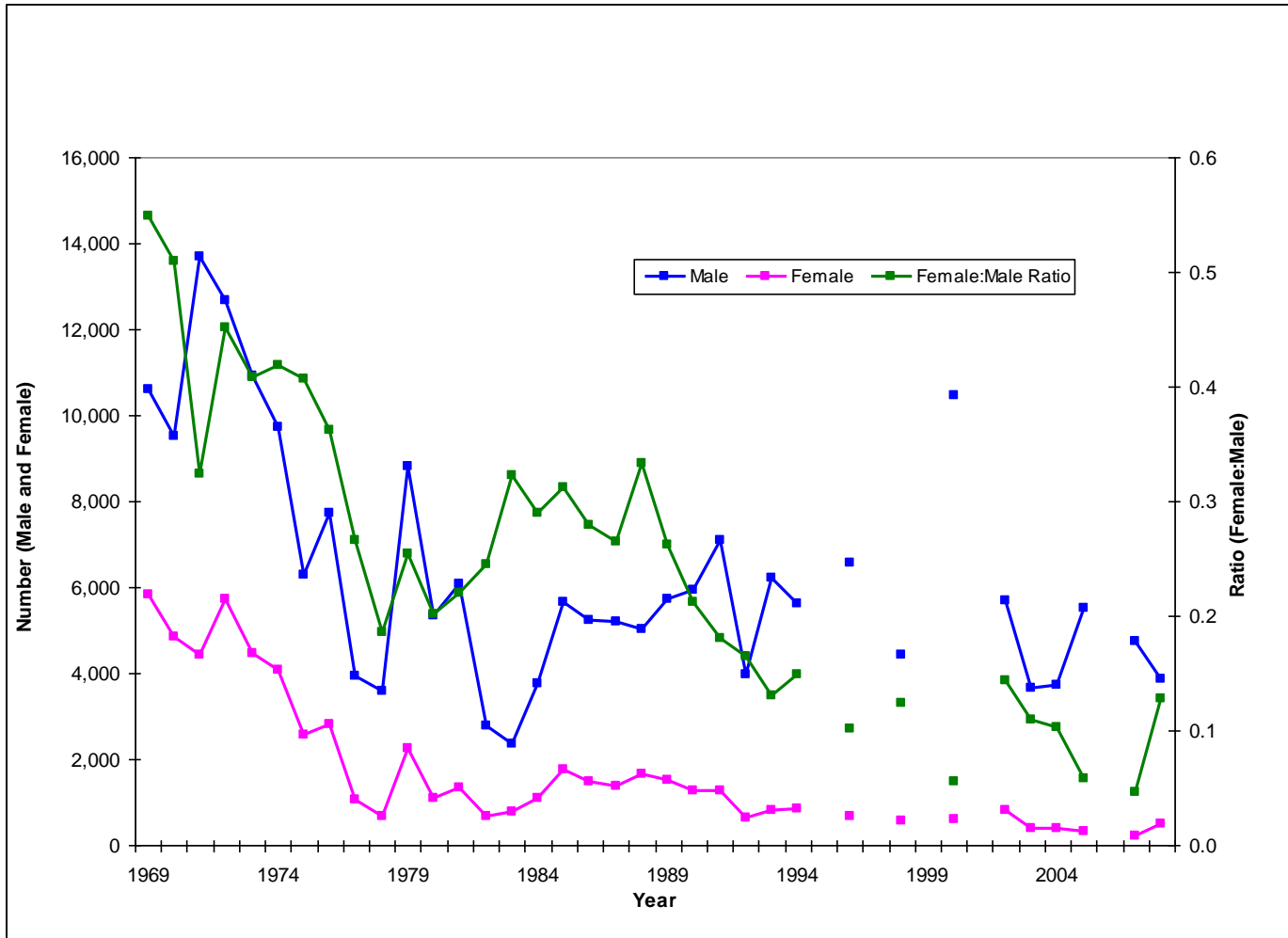


Figure 10. Number of male and female striped bass collected and corresponding female:male ratio from the Adult Striped Bass Tagging Program, 1969 – 2008. No estimates available for odd years between 1995 and 2001 or 2006.

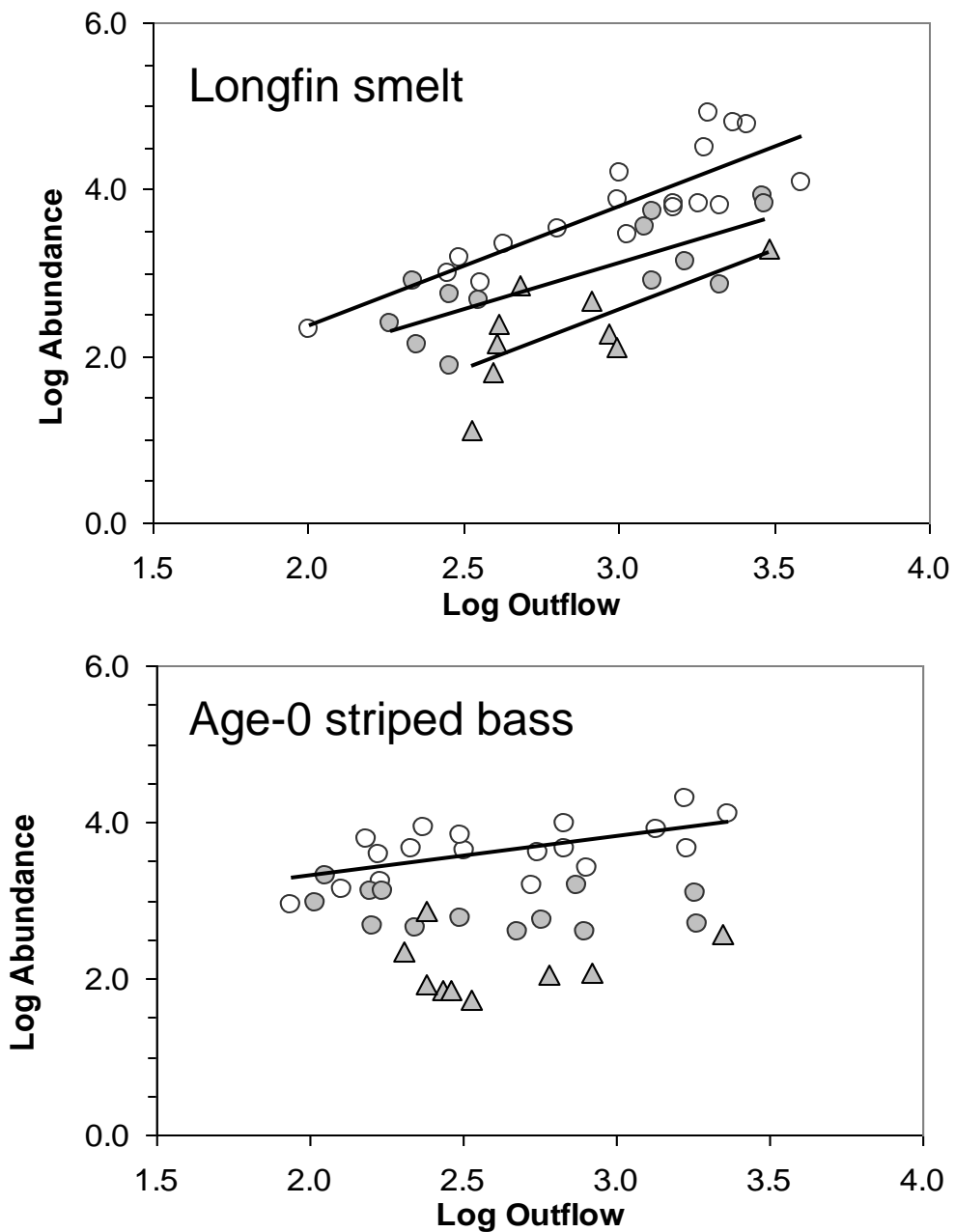
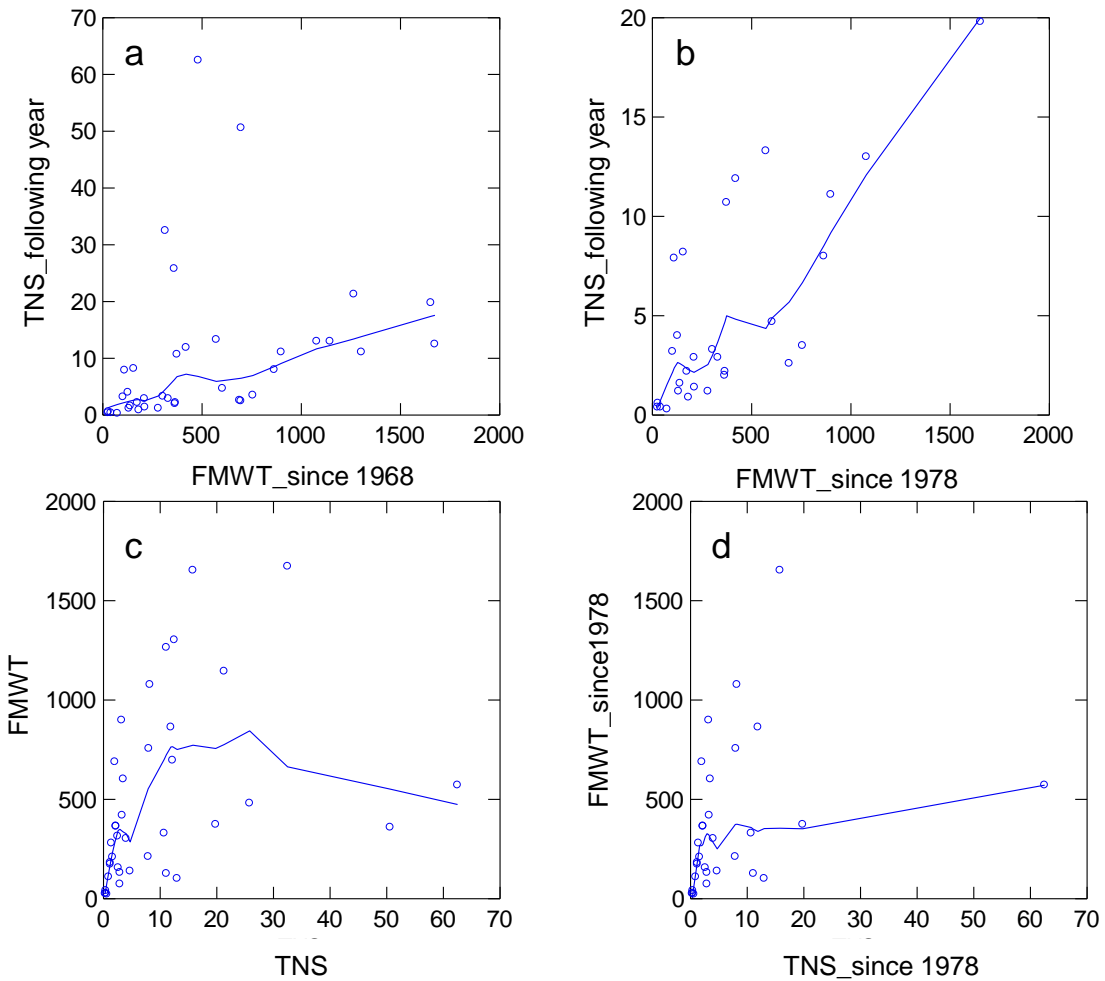


Figure 11. Log-log relationships between Fall Midwater Trawl abundance and delta outflow for longfin smelt (all ages) and age-0 striped bass. Delta outflow (m^3/s) values represent mean levels during January through June for longfin smelt and during April through July for age-0 striped bass. The data are compared for pre-*Corbula* invasion years (1967–1987; white circles), post-*Corbula* invasion (1988–2000; filled circles), and during the POD years (2001–2009; triangles). Lines depict significant linear regression relationships ($p < 0.05$).



12. Scatterplots and LOWESS splines depicting the relationship between a) the Fall Midwater Trawl index of delta smelt relative abundance (FMWT) versus the following calendar year's Summer Towntet Survey index of delta smelt relative abundance (TNS) for FMWT years 1968–2007; b) the FMWT versus the following calendar year's TNS for FMWT years 1978–2007; c) the TNS and the subsequent FMWT for 1969–2008; and d) the TNS and the subsequent FMWT for 1978–2008.

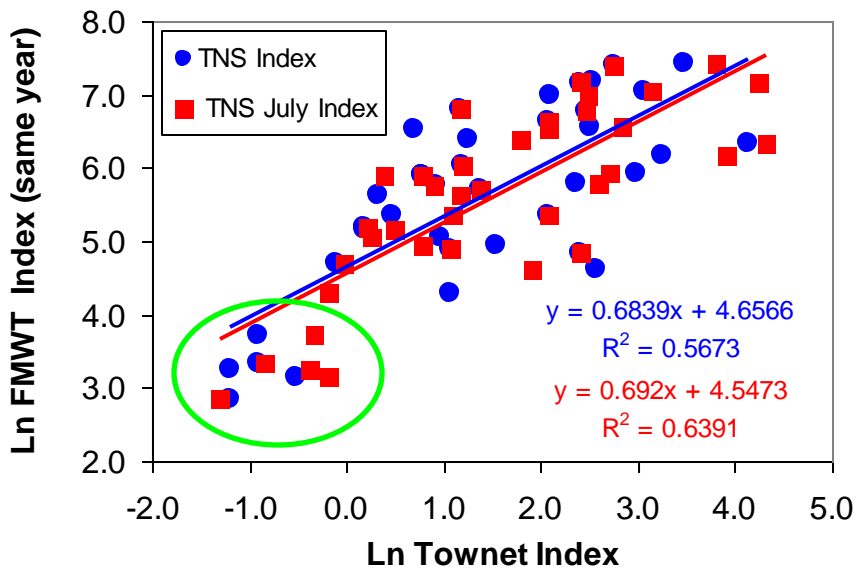


Figure 13. Ln–ln relationships between juvenile and adult lifestages of delta smelt since 1969. The Towntet Survey is a measure of summer juvenile abundance. The Fall Midwater Trawl is a measure of fall pre-spawning adult abundance. The blue circles represent the data from the first 2 Towntet Surveys which typically begins in June or July and are used to calculate the Towntet Survey delta smelt index. The red squares represent data from July only and represent an alternate index. Regression equations and coefficients are given in blue font for the full Towntet Survey data and in red font for the July Towntet Survey data. Data from 2005 to 2009 are contained in the green ellipse.

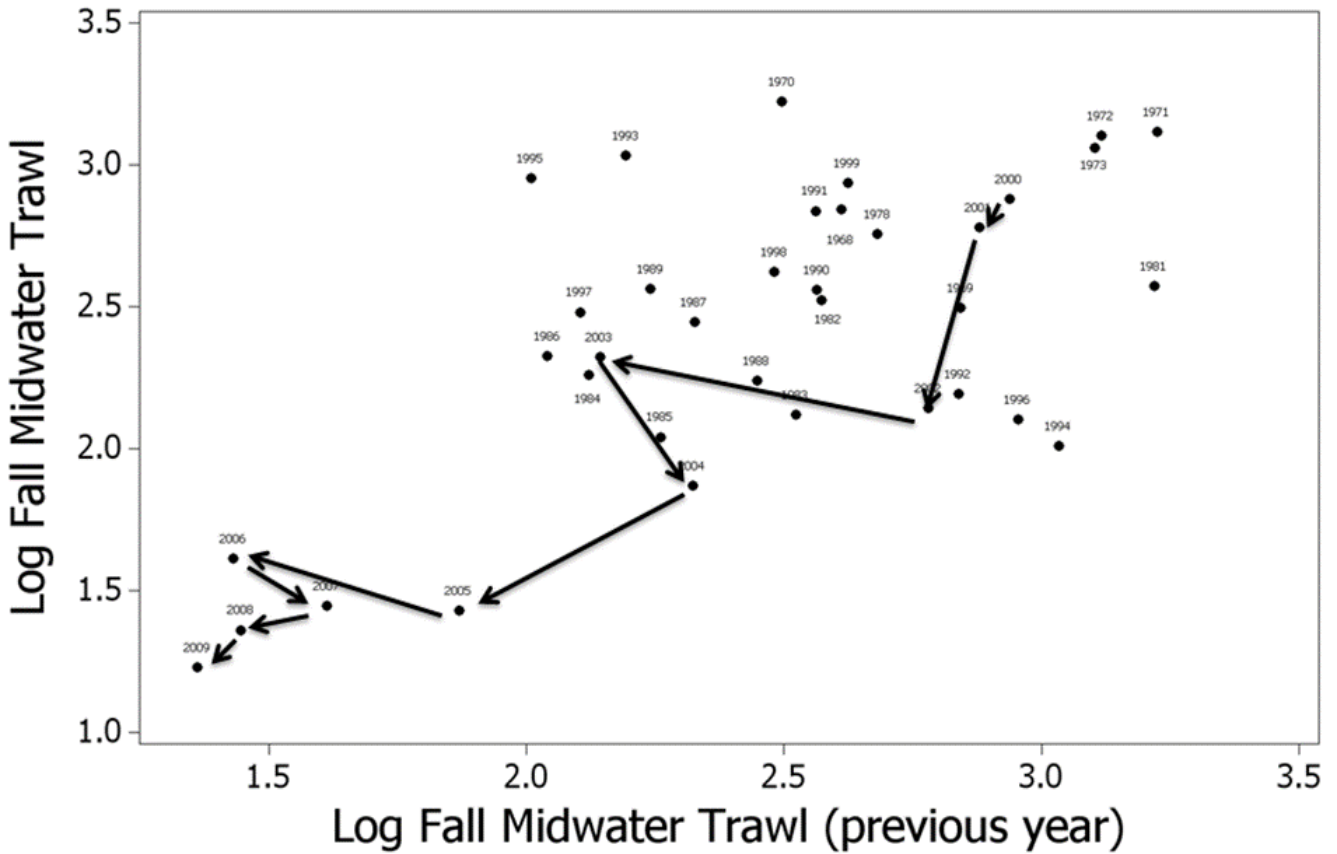
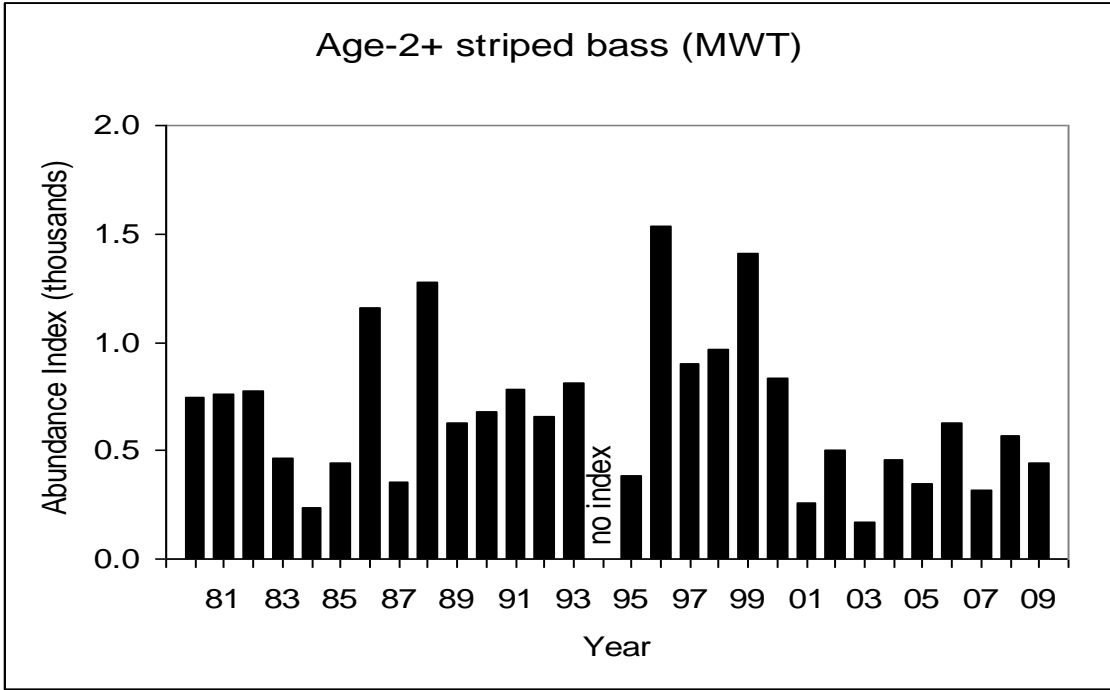
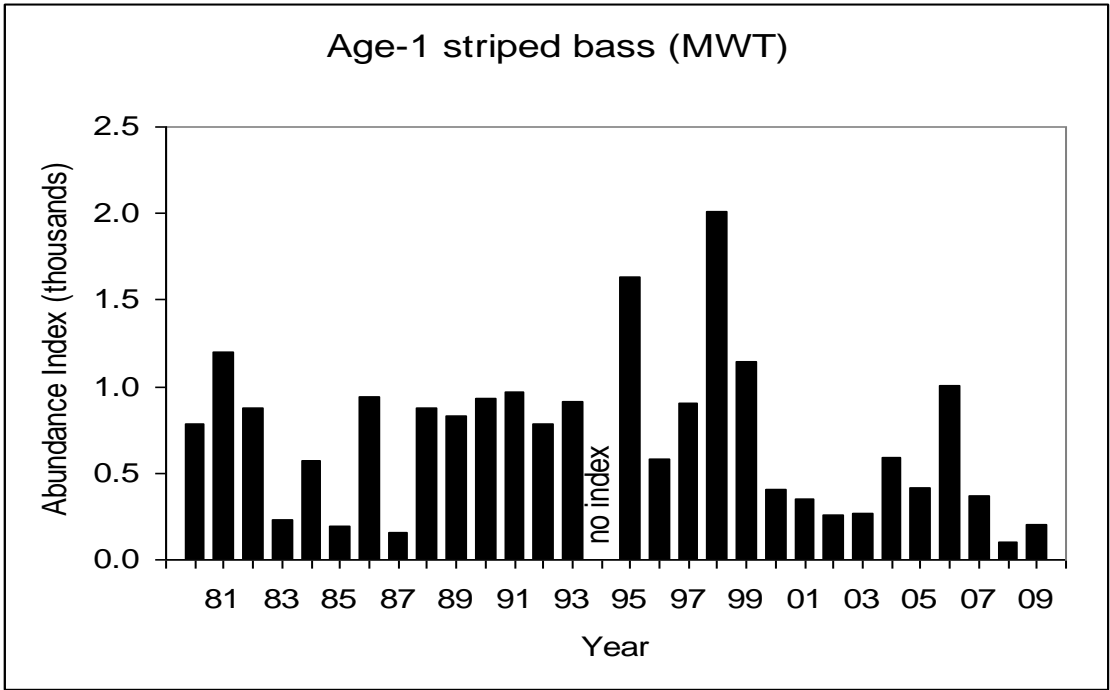
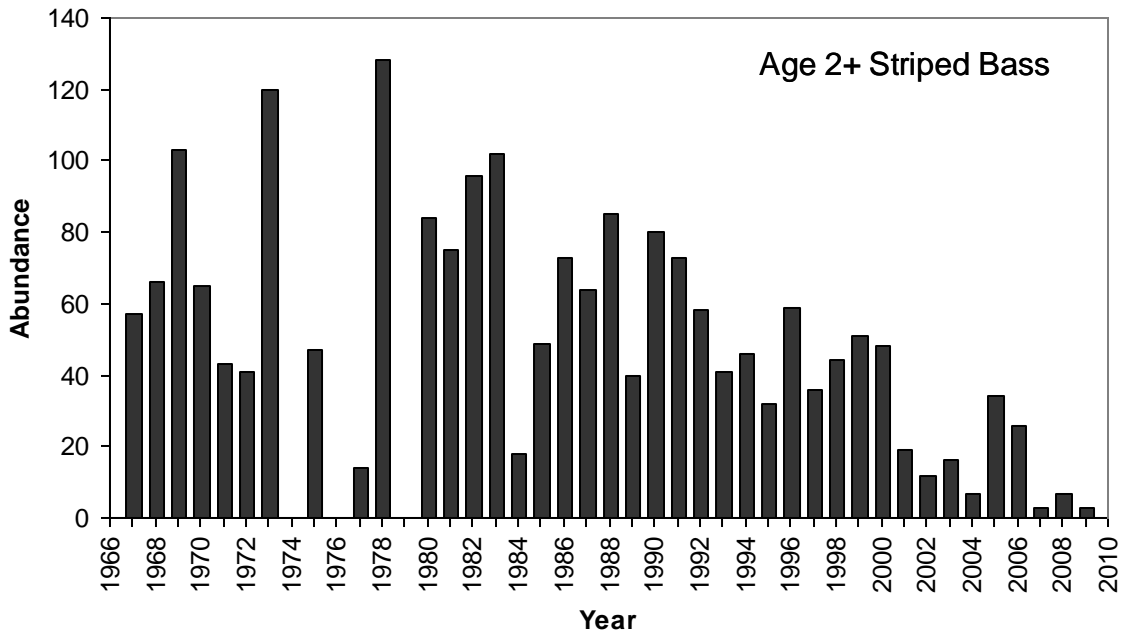
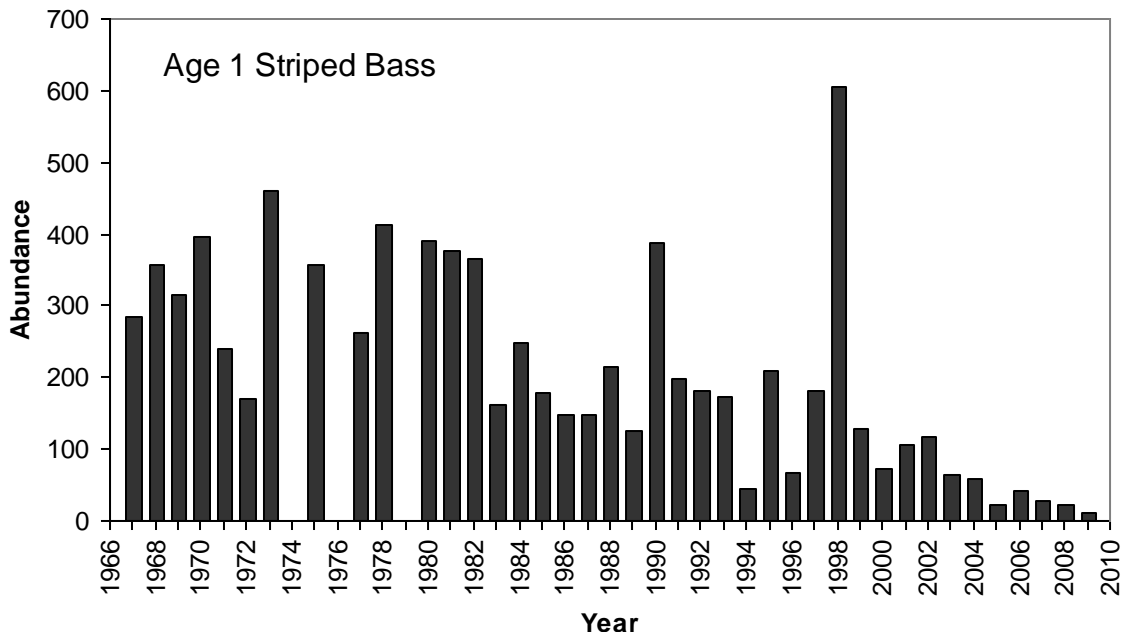


Figure14. Log–log relationships between Fall Midwater Trawl abundance and Fall Midwater Trawl abundance in the previous year. The arrows connect consecutive years from 2000 to 2009.



A



B

Figure 15. Abundance of age-1 and age-2+ striped bass in midwater trawls in A) San Francisco Bay based on the California Department of Fish and Game Bay study (Bay Study) and B) in the Delta from the Fall Midwater Trawl. No Fall Midwater Trawl indices were calculated for years 1974, 1976 and 1979.

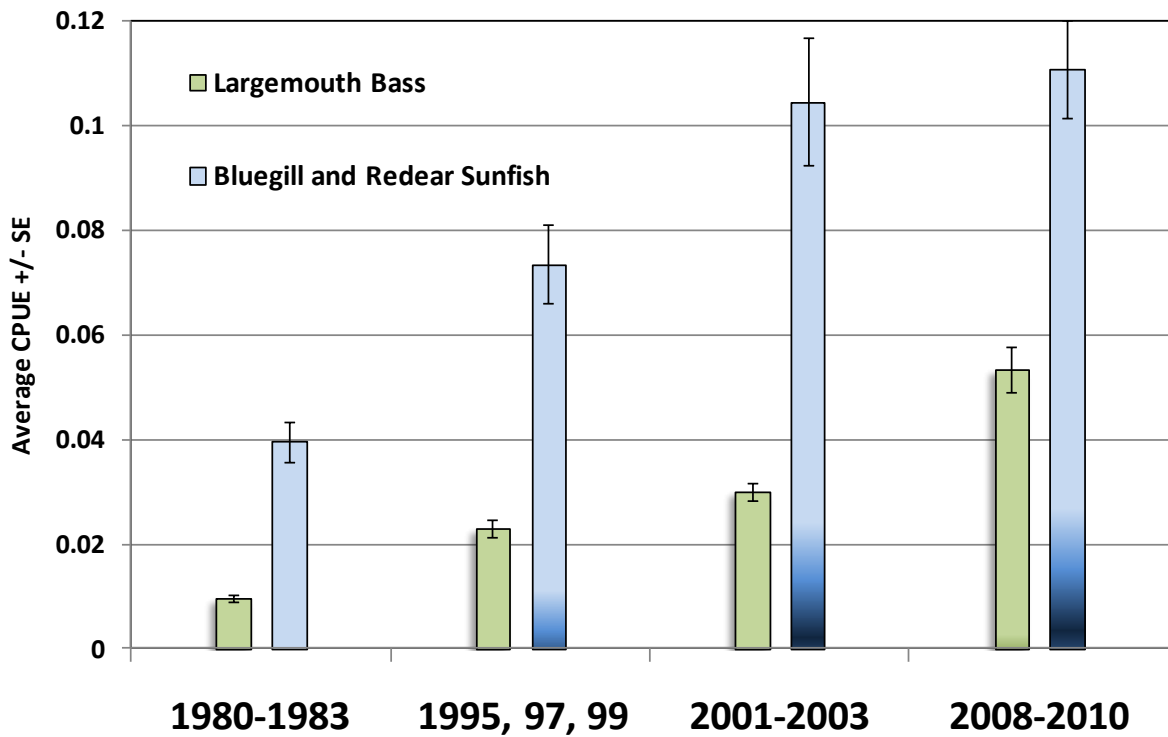


Figure 16. Average catch per unit effort (CPUE) for largemouth bass, bluegill, and redear sunfish from successive electrofishing efforts in the Delta from 1980 to 2010. Data from 1980–2003 is from CDFG (unpublished data) and data from 2008–2010 is from L. Conrad (DWR, unpublished data).

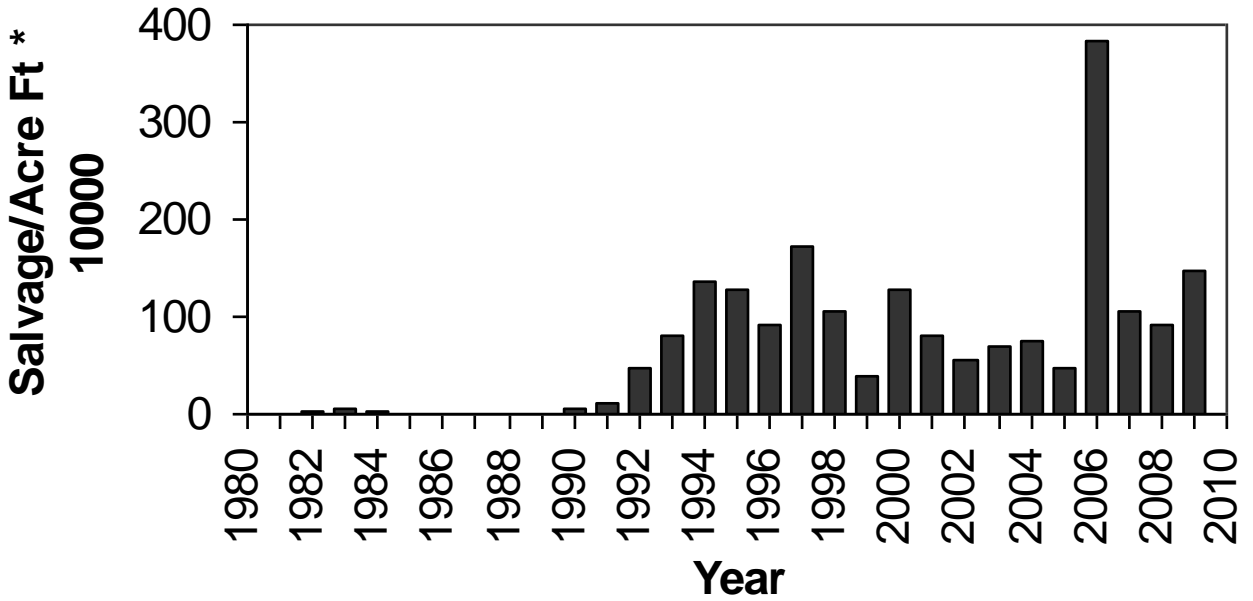


Figure 17. Annual salvage density (fish per 10,000 acre feet) of largemouth bass at the CVP and SWP combined from 1979 to 2009 (California Department of Fish and Game, unpublished data).

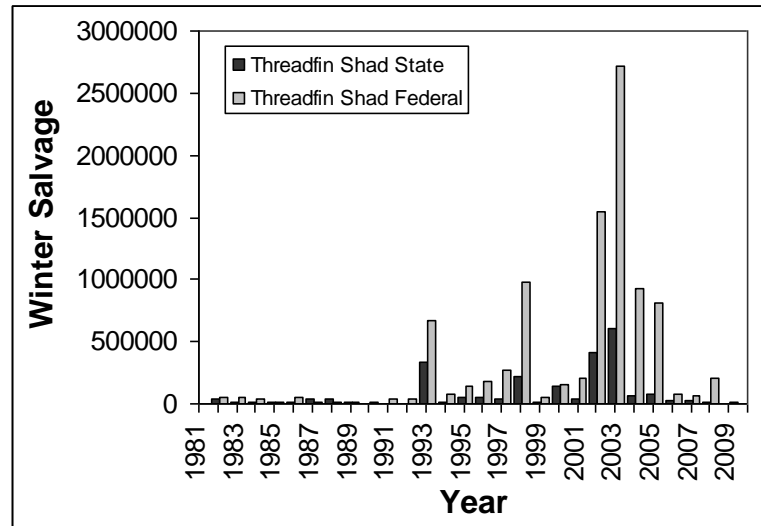
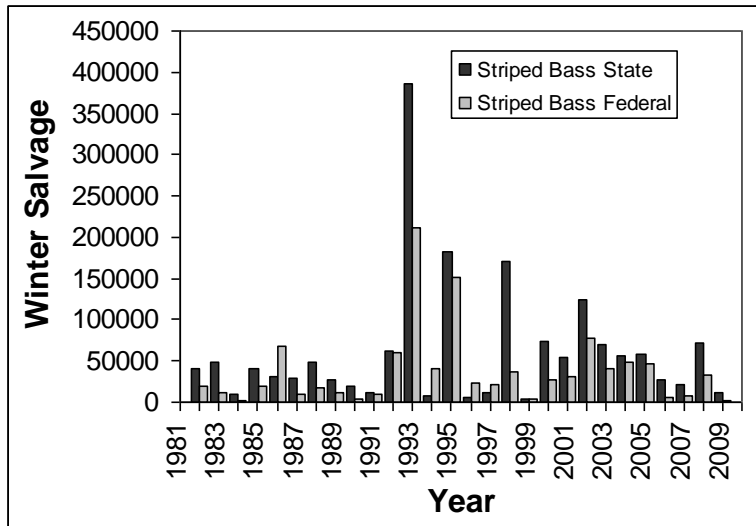
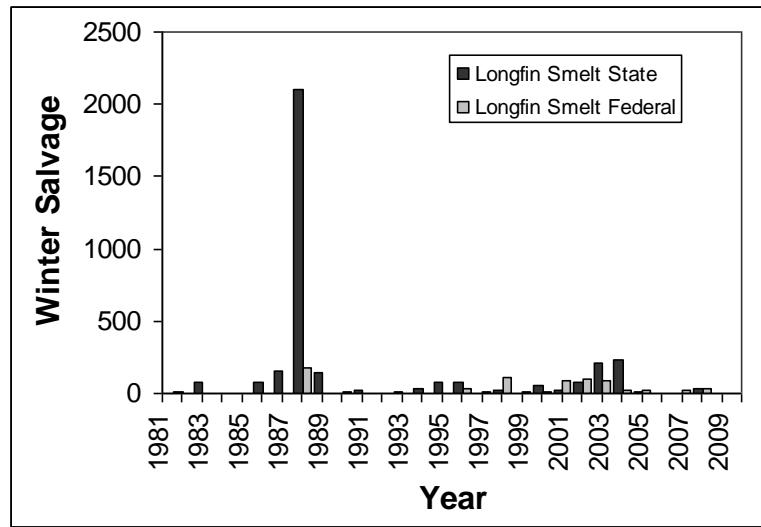
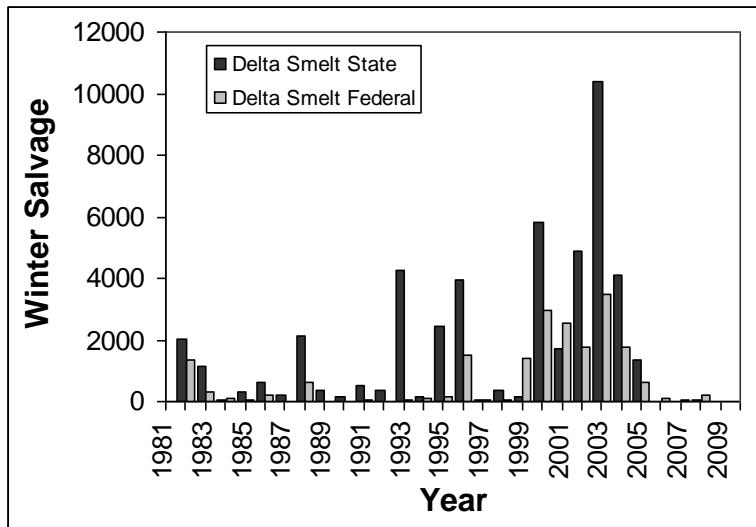
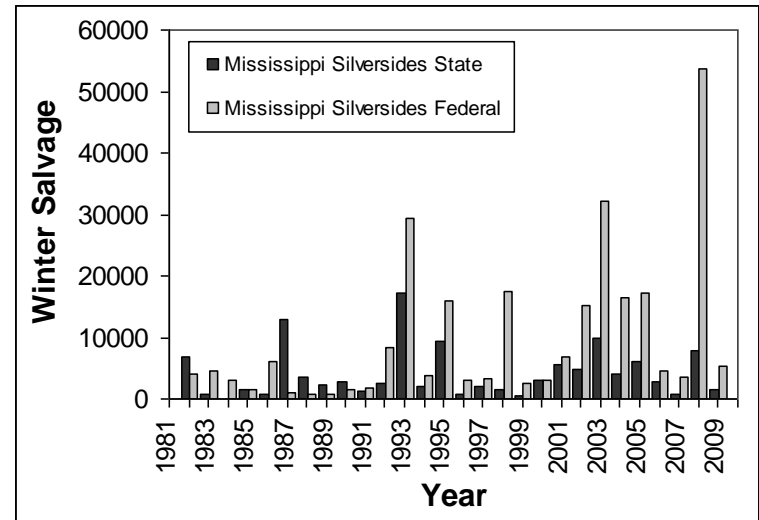
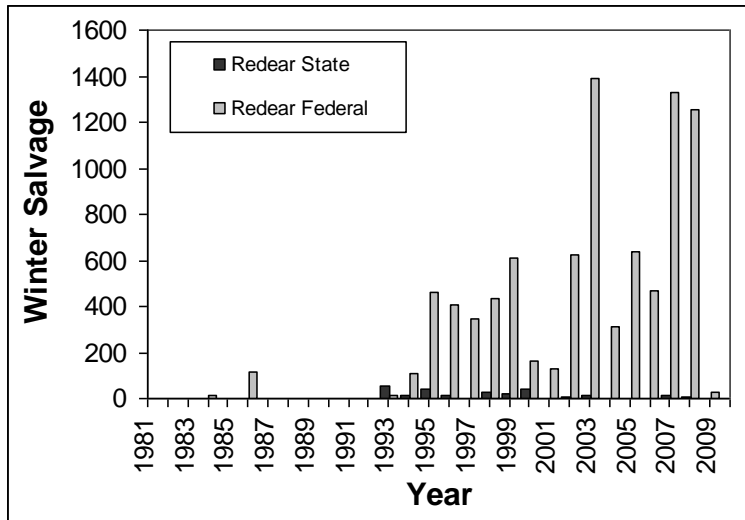
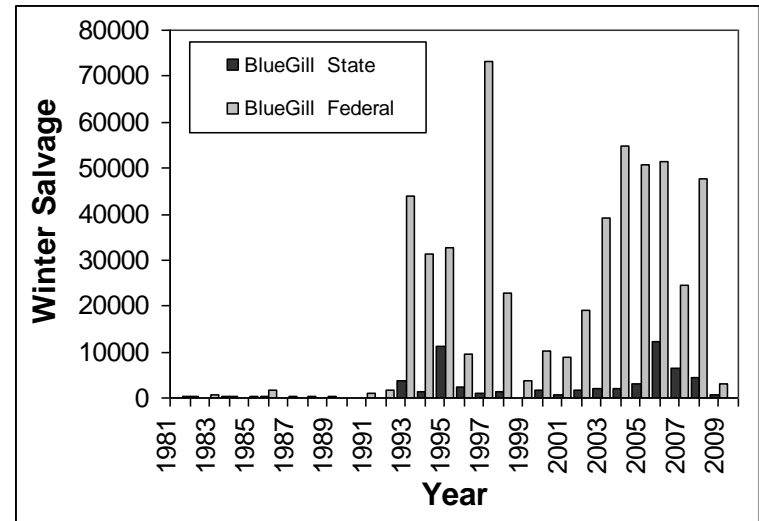
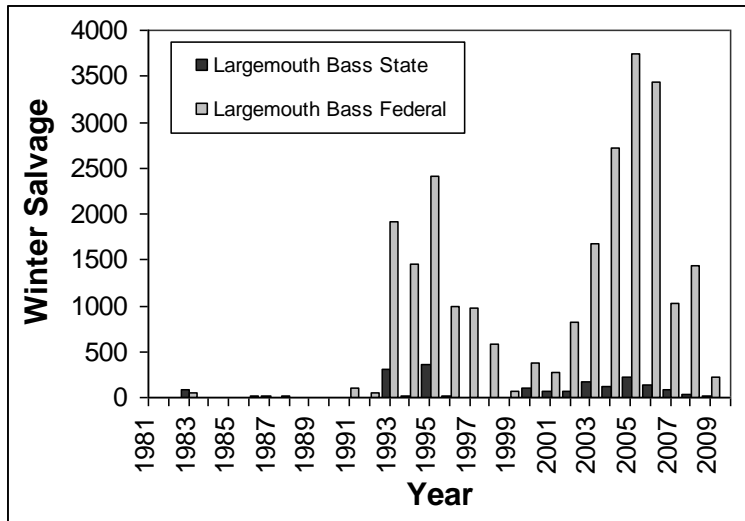


Figure 18. Winter salvage data for striped bass, delta smelt, longfin smelt, and threadfin shad for the federal Central Valley Project (Federal) and State Water Project (State) from 1981 to 2009. Salvage for delta smelt and longfin smelt before 1993 should be interpreted with caution because of variable degrees of training among personnel identifying fishes.



19. Winter salvage for largemouth bass, inland silversides, bluegill, and redear sunfish for the federal Central Valley Project (Federal) and State Water Project (State) from 1981 to 2009.

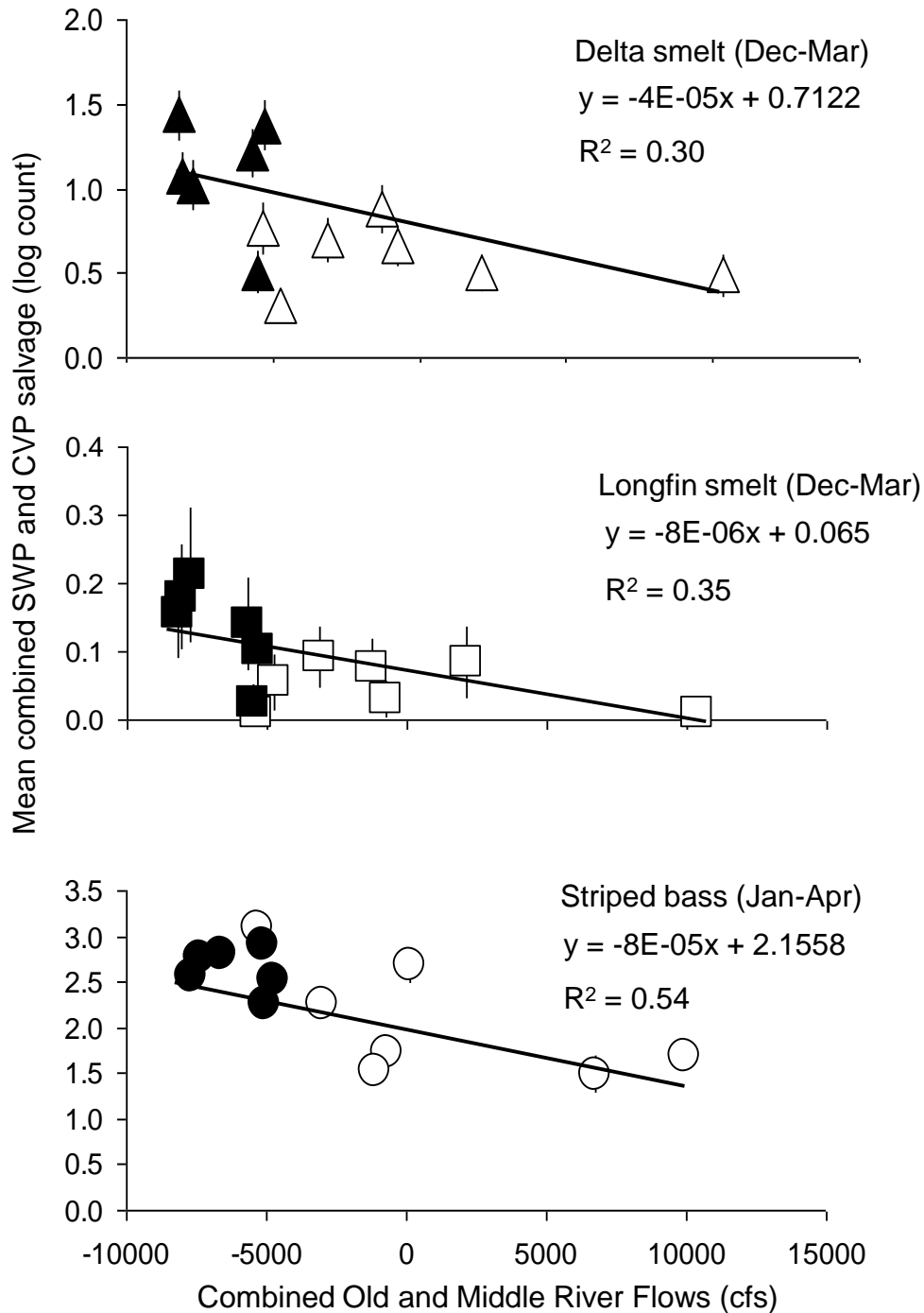


Figure 20. Relationship of mean combined salvage of delta smelt, longfin smelt, and striped bass at the State Water Project (SWP) and Central Valley Project (CVP) to combined Old and Middle rivers flow (cubic feet per second). Open symbols denote pre-POD years (1993–1999) and filled symbols represent post-POD years (2000–2005) (Grimaldo et al. 2009).

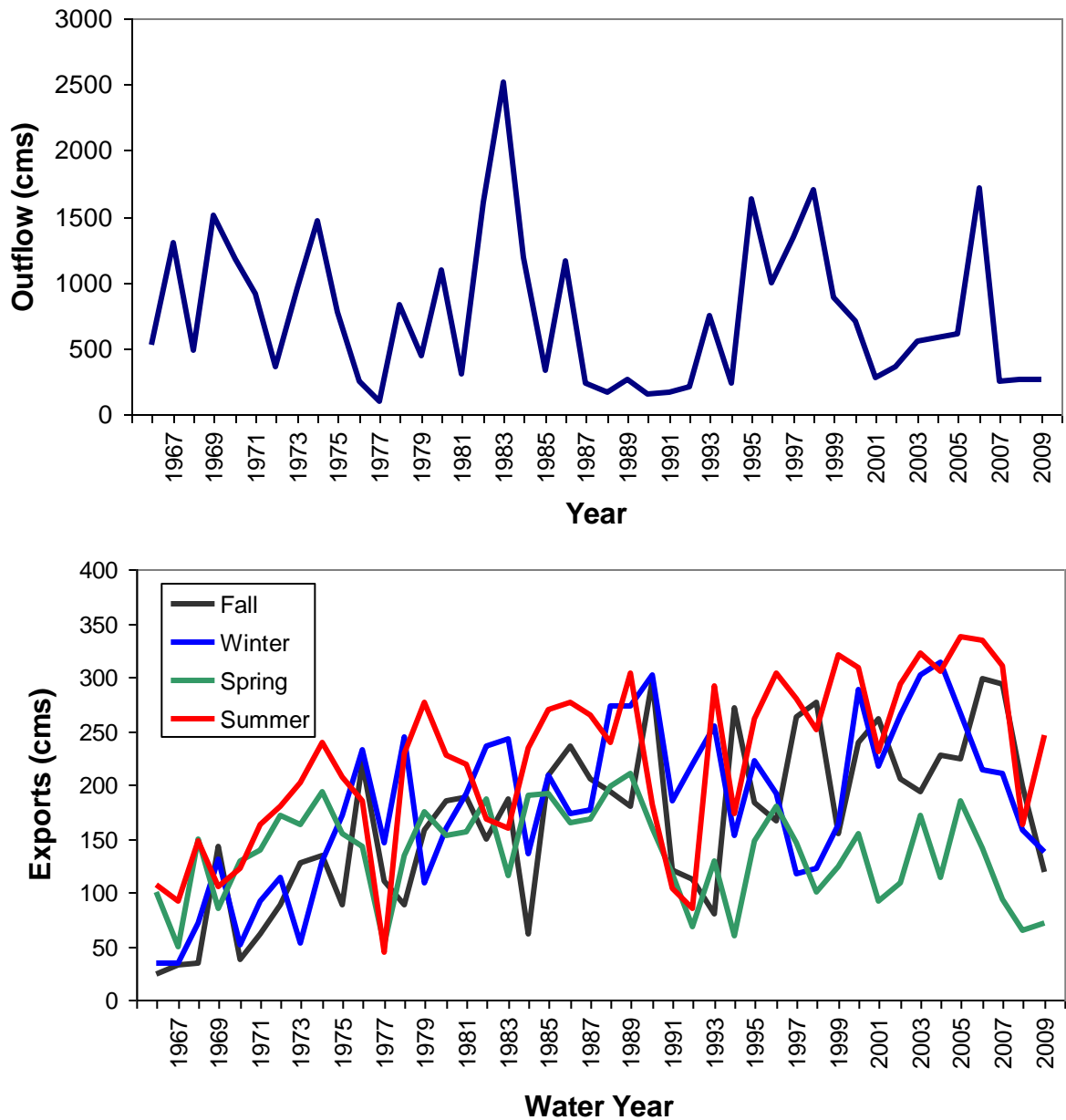


Figure 21. Delta outflow (m^3/s) averaged over water years (top) and export flow (m^3/s) averaged over seasons (bottom). Water years begin on 1 October of the previous calendar year. Seasons are in 3-month increments starting in October. Export flows are the sum of diversions to the federal Central Valley Project and State Water Project pumping plants. The outflow and export data are from California Department of Water Resources (<http://iep.water.ca.gov/dayflow>).

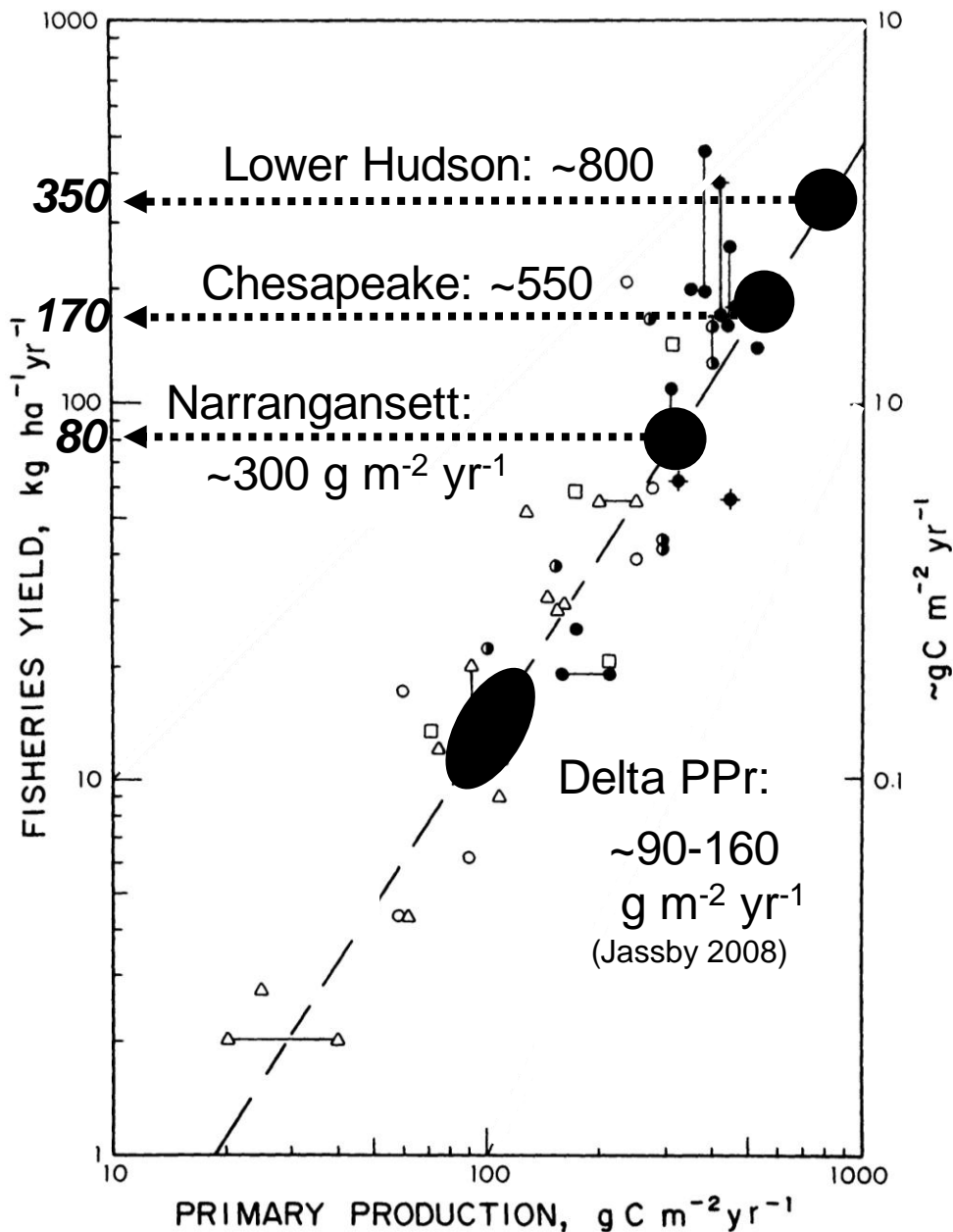


Figure 22. Range in primary production in Suisun Bay and the Delta since 1975 plotted on the relationship of fishery yield to primary production from other estuaries around the world (modified from Nixon 1988, using results in Jassby 2008 and Jassby et al 2002 and data provided by James Cloern, U.S. Geological Survey).

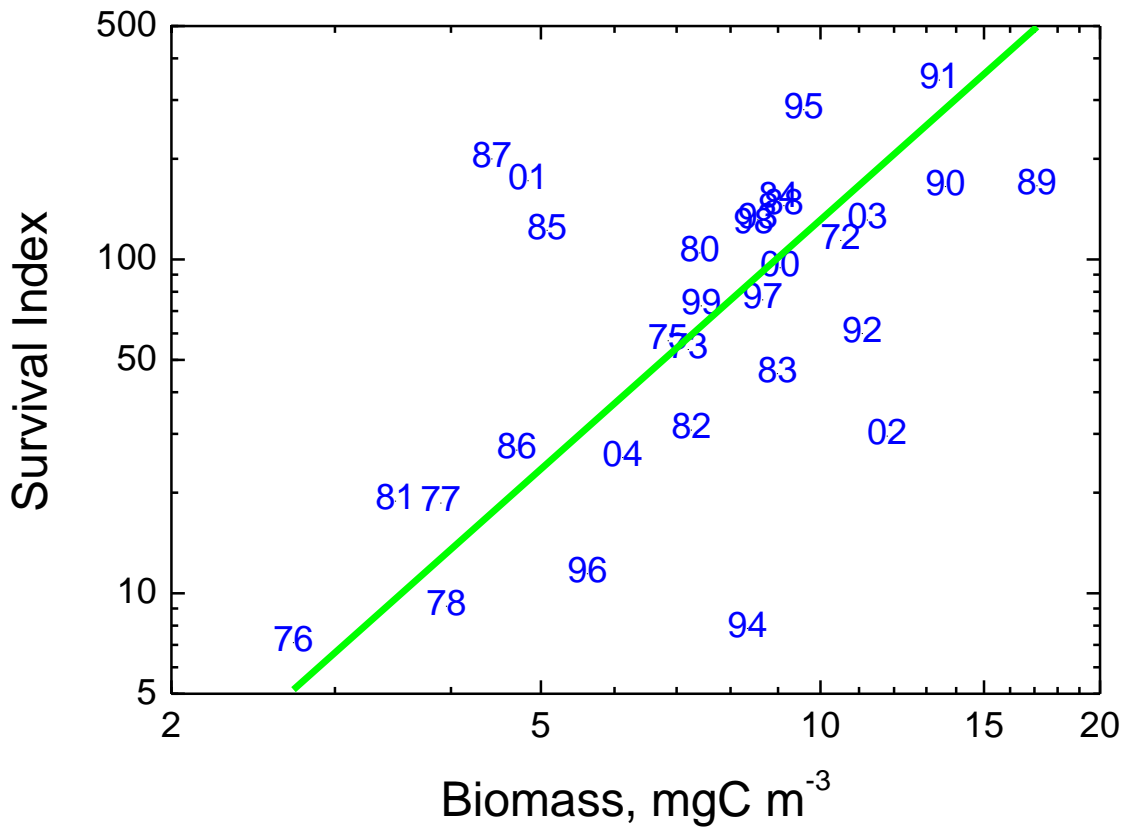
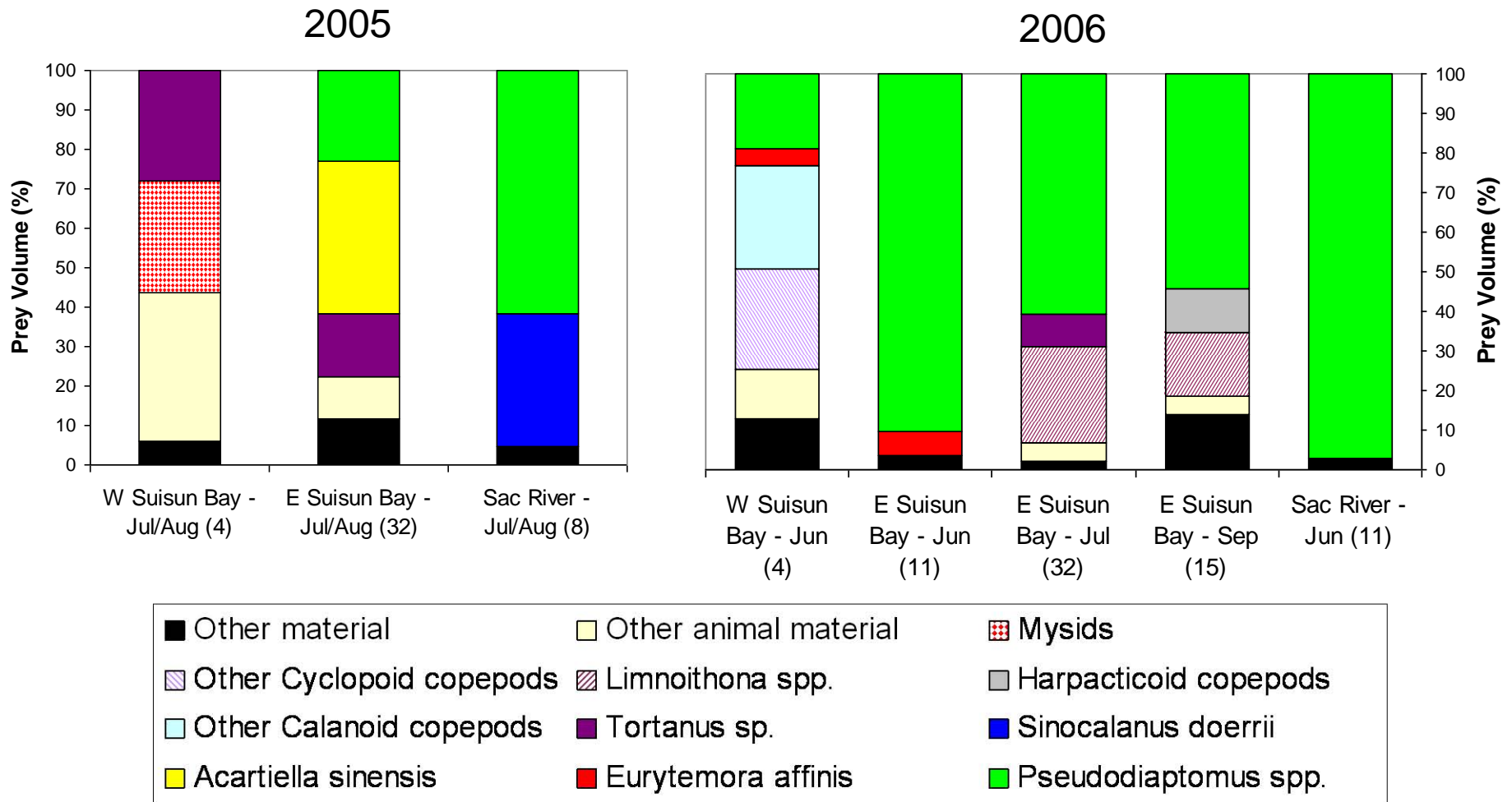


Figure 23. Summer to fall survival index of delta smelt in relation to zooplankton biomass in the low salinity zone (0.15 – 2.09 psu) of the estuary. The survival index is the log ratio of the Fall Midwater Trawl index to the Summer Townet Survey index. The line is the geometric mean regression for log(10)-transformed data, $y = 2.48x - 0.36$. The correlation coefficient for the log-transformed data is 0.58 with a 95% confidence interval of (0.26, 0.78) (Kimmerer, 2008).



24. Prey volume in guts of delta smelt collected during summer 2005 and 2006. Sample size appears in parentheses (S. Slater, California Department of Fish and Game, unpublished data).

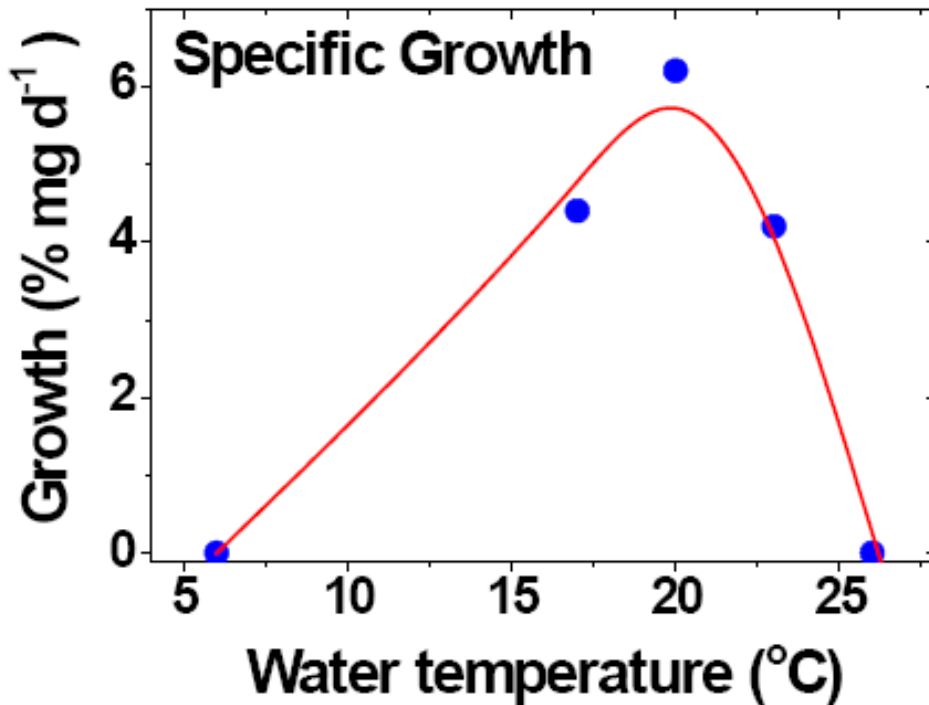


Figure 25. Water temperature influences on growth in the juvenile stage (i.e. specific growth) from studies by Baskerville et al. (2004) during aquaculture. A cubic spline model was fitted to approximate the functional form of specific growth with respect to water temperature (from Bennett et al. 2008).

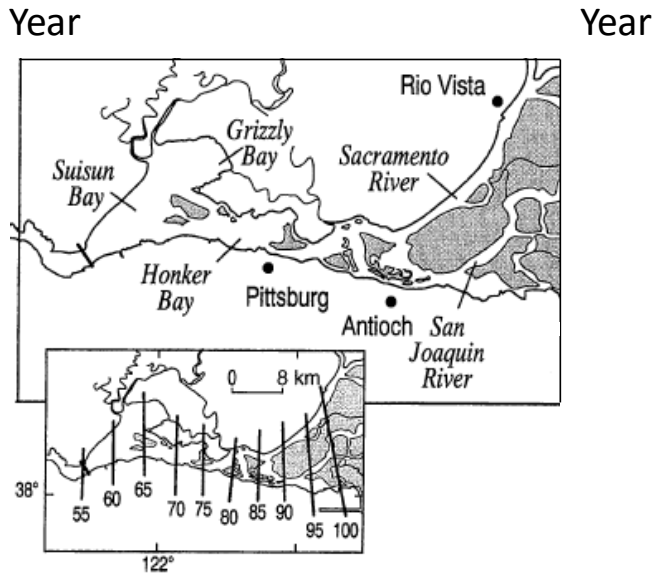
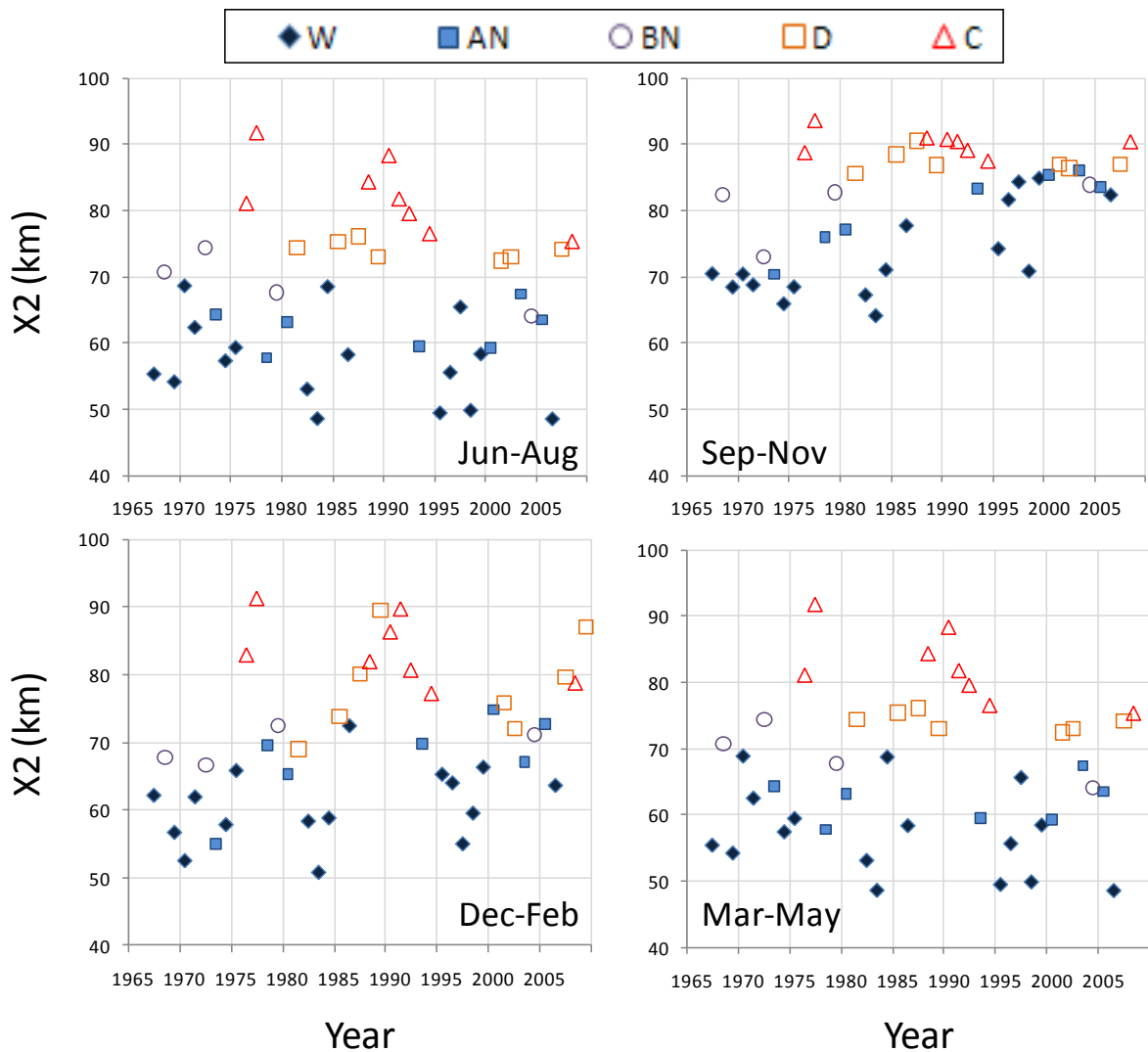


Figure 26. Seasonal means of X2, the location of the 2‰ bottom isohaline along the axis of the estuary (distance from the Golden Gate in km) from 1967–2009. Symbols indicate water year types for the Sacramento Valley (W: Wet, AN: Above Normal, BN: Below Normal, D: Dry, C: Critically Dry). Map (modified from Jassby et al 1995): Suisun Bay and the western Delta. Insert shows lines positioned along nominal distances (in km) from the Golden Gate along the axis of the estuary.

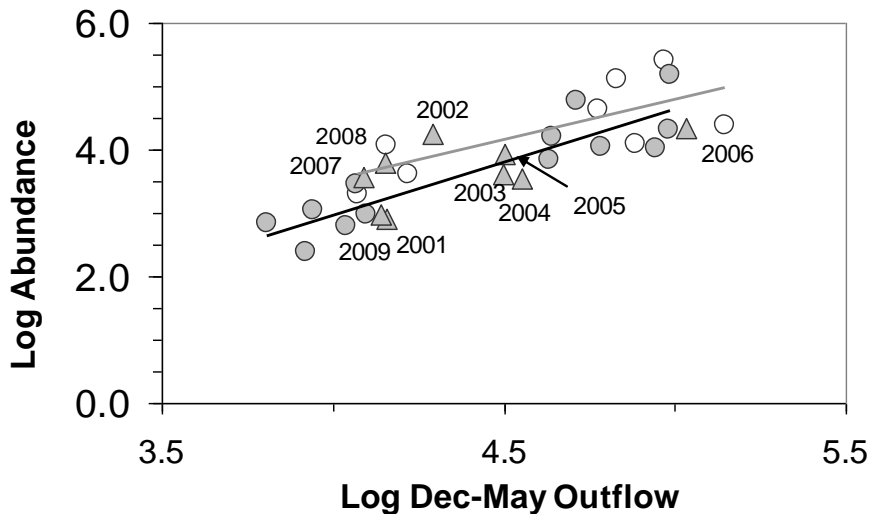
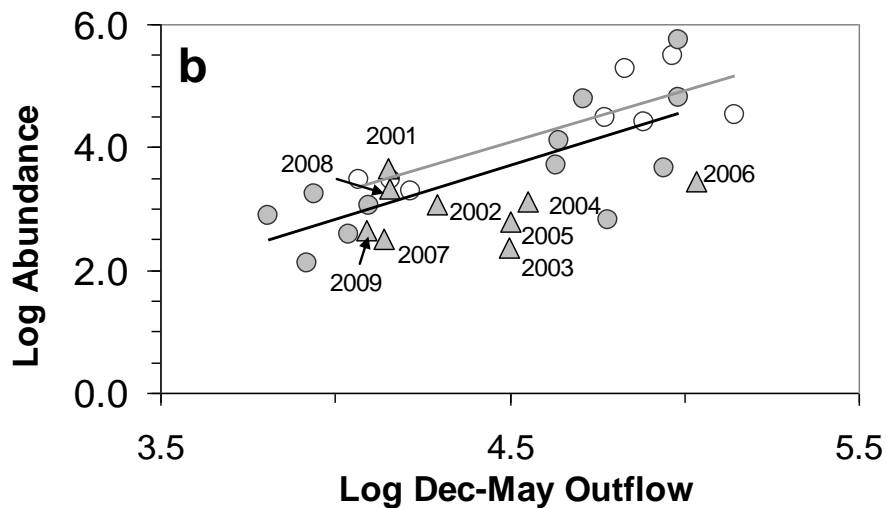
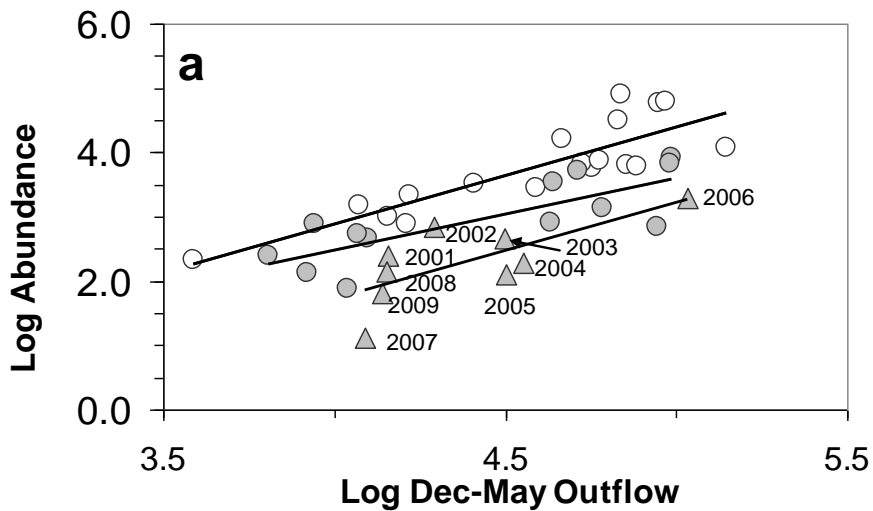


Figure 27. Longfin smelt outflow abundance relationships based on December through May mean outflow (cfs) and (a) Fall Midwater Trawl annual abundance, 1967–2009, all ages, (b) Bay Study Age-0 midwater trawl abundance, 1980–2009, and (c) Bay Study age-0 otter trawl abundance, 1980–2009. Abundance data are compared for pre-*Corbula* invasion years (survey start – 1987, open circles), post-*Corbula* invasion (1988–2000, filled circles) and POD years. Fitted lines indicate linear regression relationships that are significant at the $p < 0.05$ level.

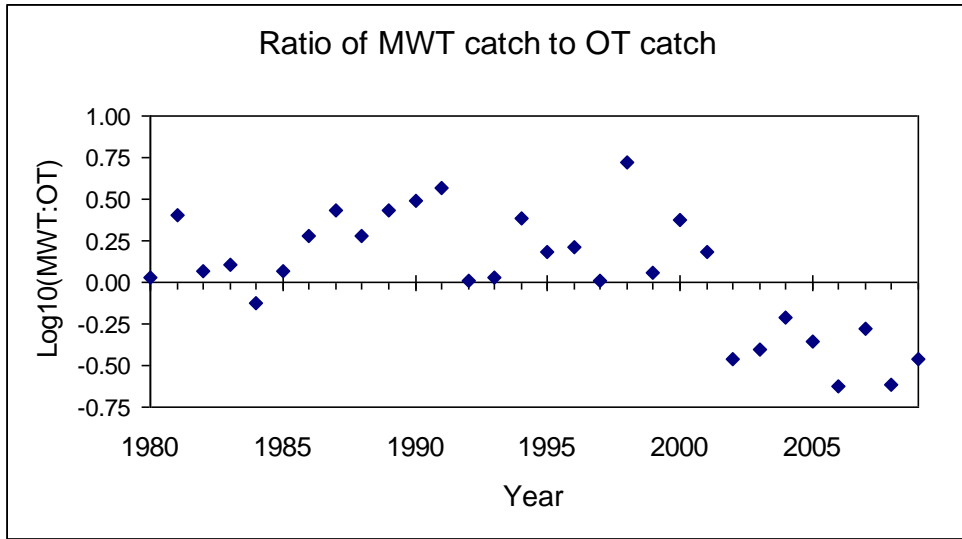


Figure 28. Ratio of annual longfin smelt catch (all ages) in the Bay Study midwater (MWT) and otter (OT) trawls (log 10 transformed), 1980–2009. Catch data are from all months of the year sampled, which varied by year, and all 35 core stations where valid tows were completed for both nets.

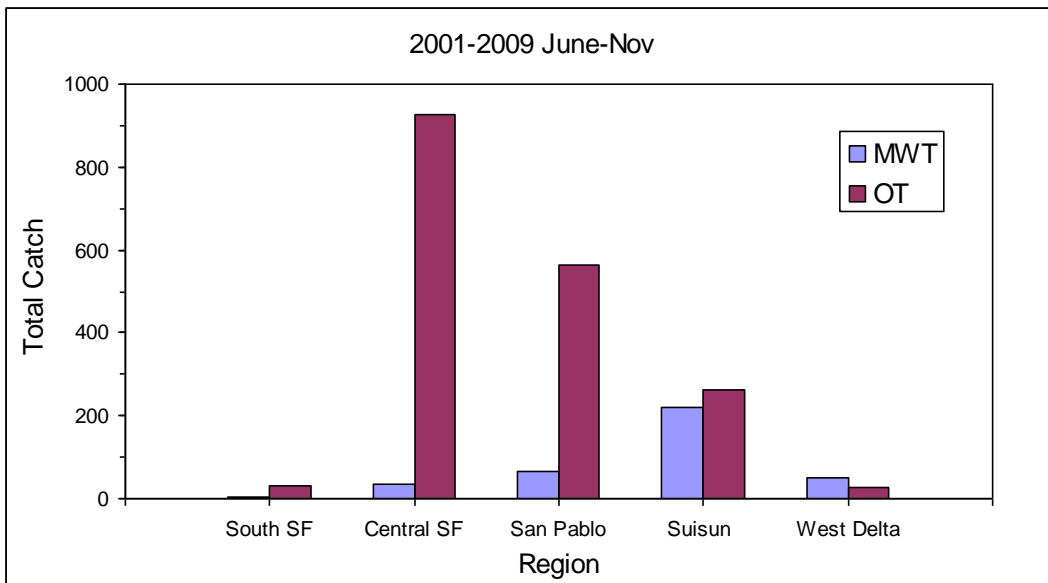
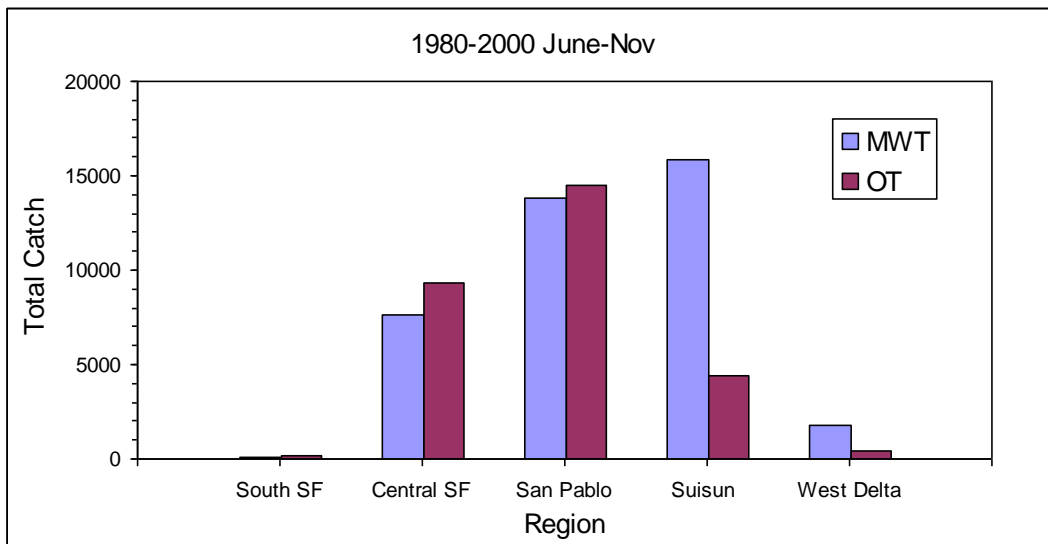


Figure 29. Total June through November longfin smelt catch (all ages) in the Bay Study midwater (MWT) and otter (OT) trawls, 1980–2000 for 35 core stations where valid tows were completed for (a) both the nets, and (b) the same data for POD years 2001–2009.

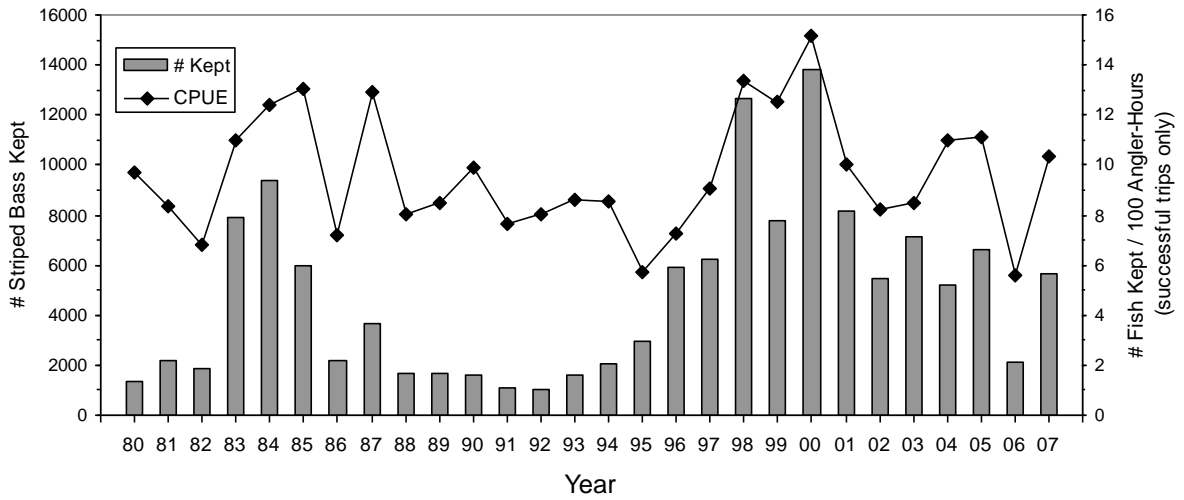


Figure 30. Harvest and CPUE for Striped Bass by Commercial Passenger Fishing Vessels in the San Francisco Bay and Delta, 1980–2009.

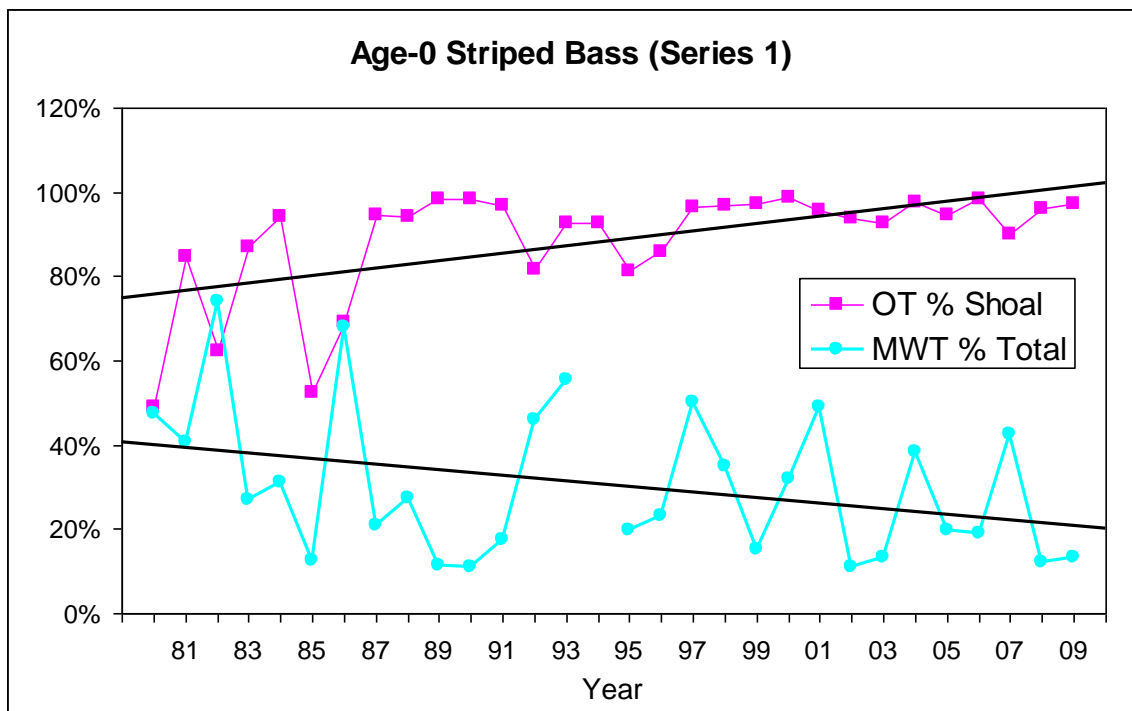


Figure 31. San Francisco Bay Study age-0 striped bass catch May–October. The blue line and circles represent the proportion of the total catch of the combined midwater trawl (MWT) and otter trawl catch (OT) that was taken in the midwater trawl; data for 1994 are incomplete and were not plotted. The pink line and solid squares represent the proportion of the total otter trawl catch taken at shoal stations (<7m deep).

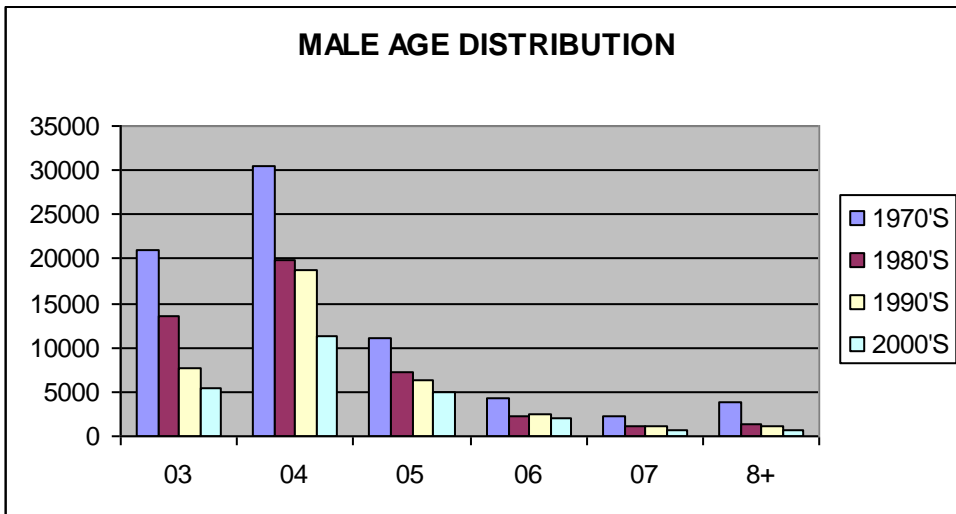
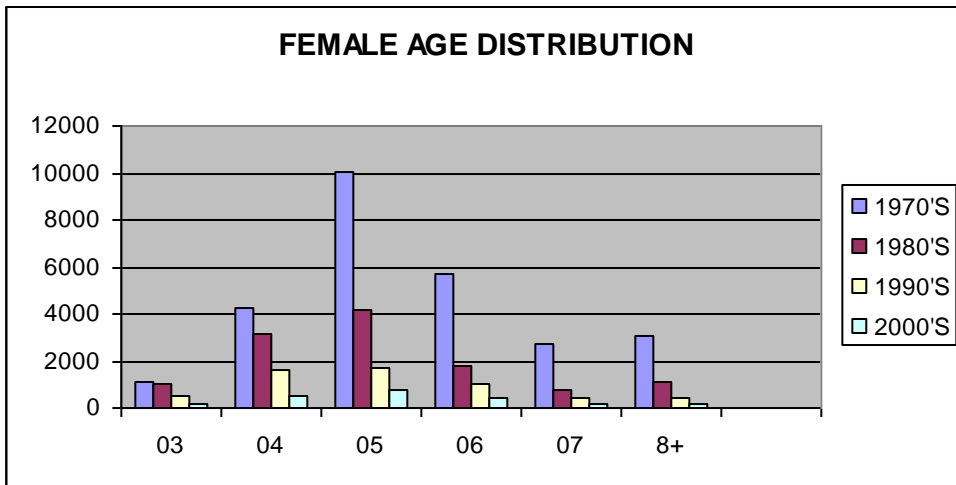


Figure 32. Male and female age distribution by decade for striped bass ages 3 – 8+ in fyke net collections. Note that age distribution is the same for gill-net collected fish (graphic not displayed).

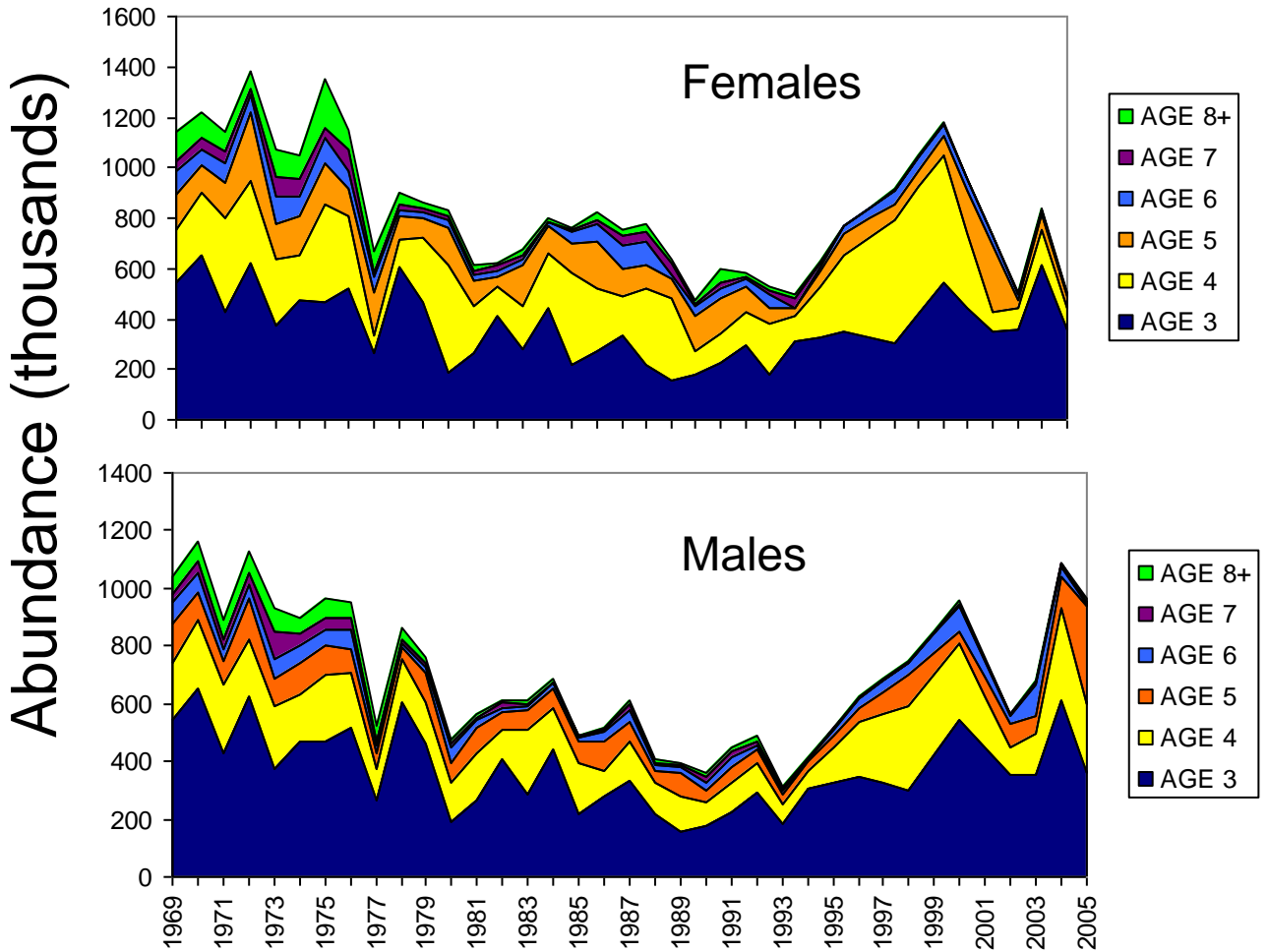


Figure 33. Abundance estimate (in thousands) of male and female striped bass $\geq 3 - 8+$ collected in the Adult Striped Bass Tagging Program, 1969 – 2005.

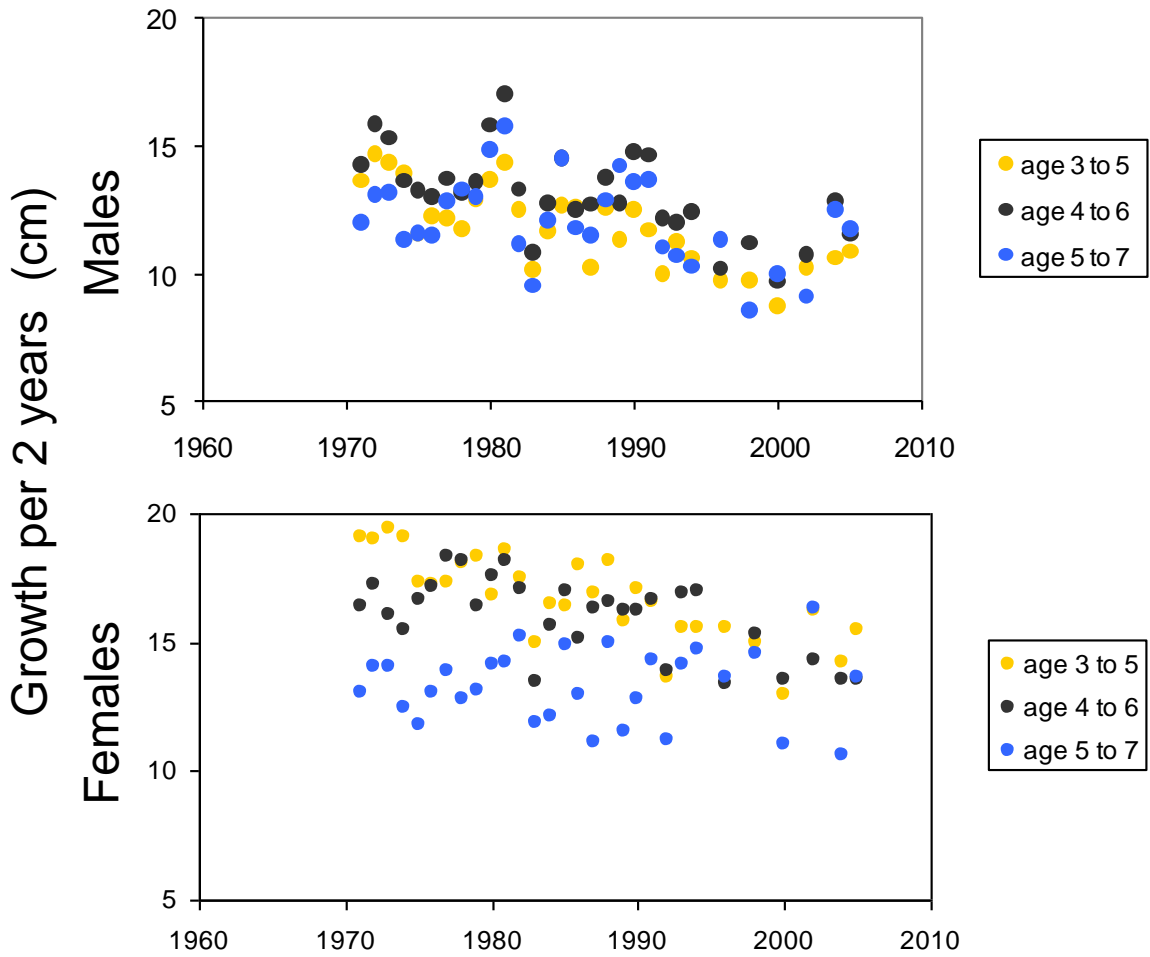
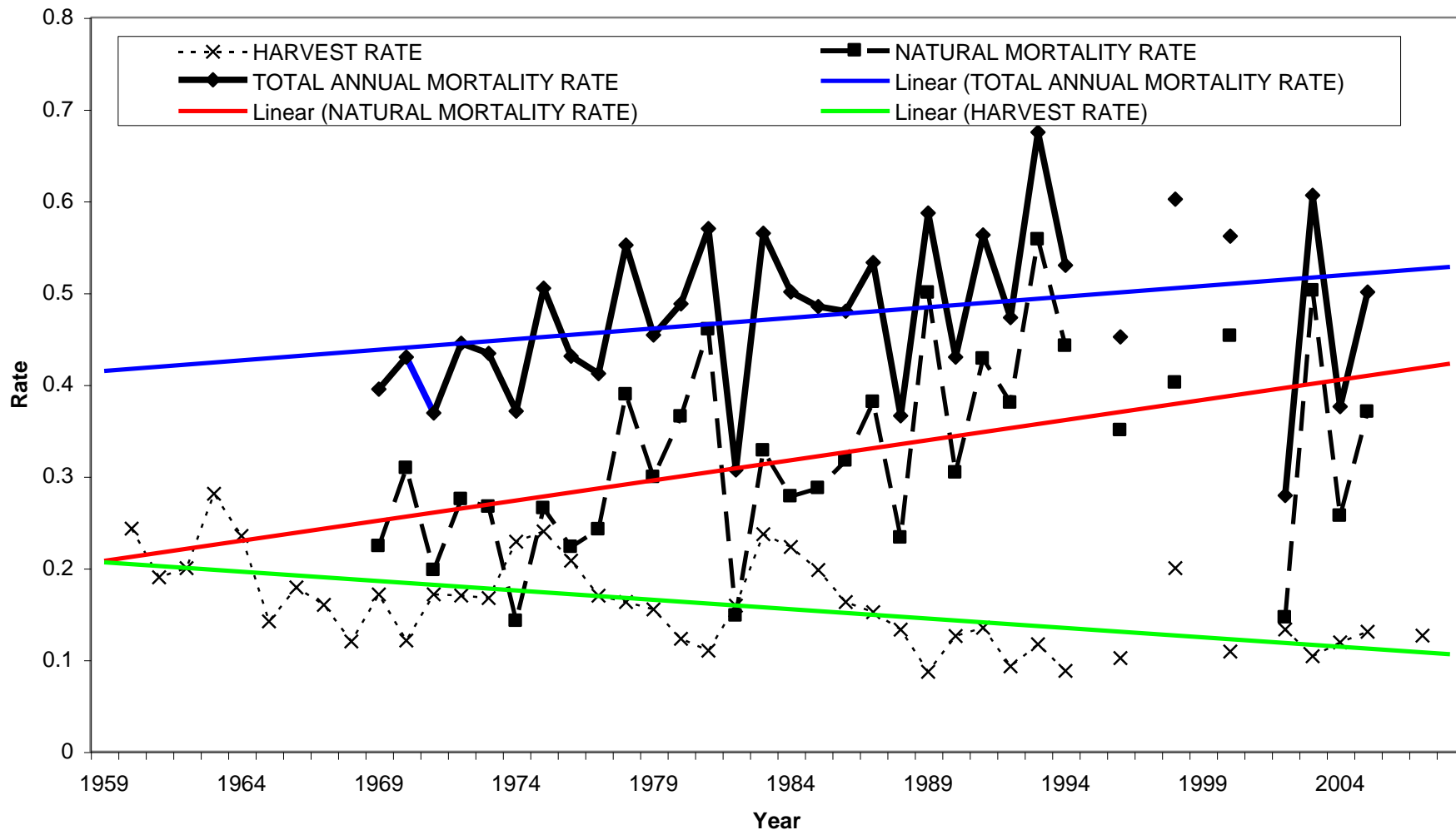
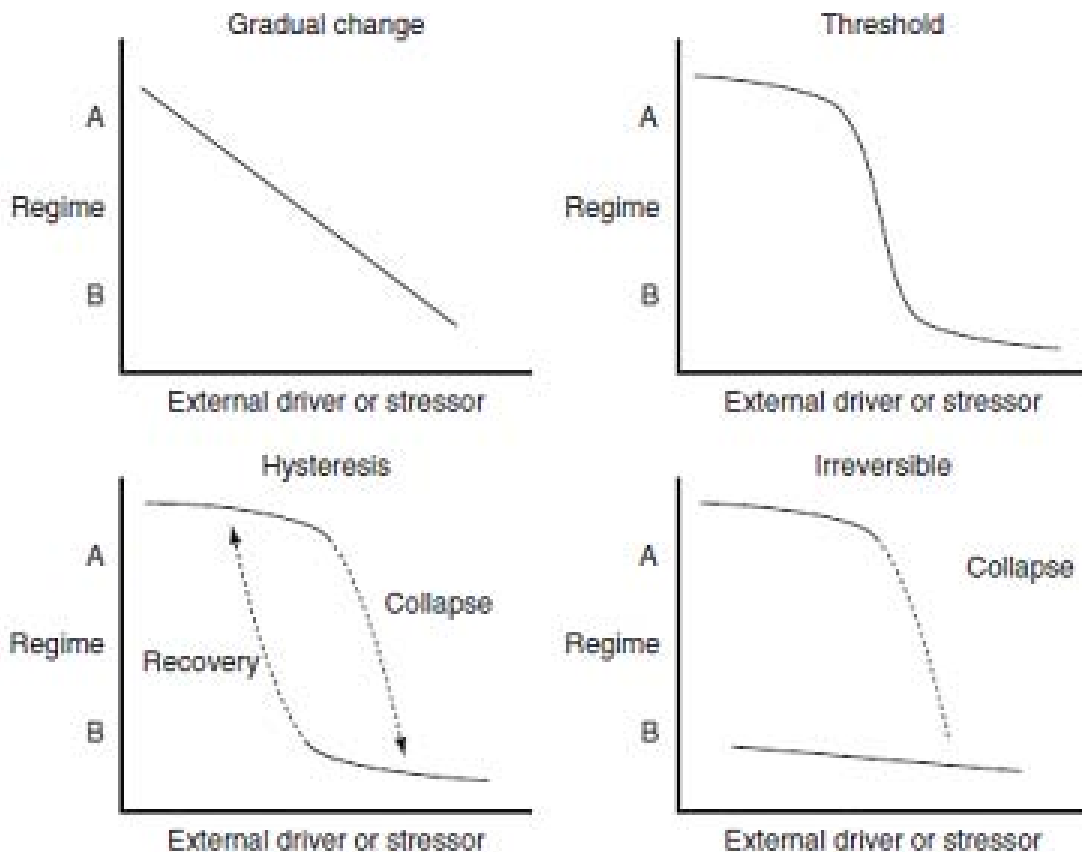


Figure 34. Apparent growth of male and female striped bass ages 3 – 7, calculated from data collected from the Adult Striped Bass Tagging Program, 1970 – 2005.



35. Estimates of harvest rate (1960 – 2007), and natural and total annual mortality rate (1969 – 2007) of legal-sized striped bass $\geq 3 - 8+$, Adult Striped Bass Tagging Program.



36. Four models of ecological change. The gradual change model shows a gradual linear change from regime A to regime B in response to a gradually changing environmental driver. The transition can be reversed by reversing the driver. The threshold model shows a nonlinear abrupt change from regime A to regime B occurring in response to a relatively small change in the environmental driver. Reducing the environmental driver to a value below the threshold results in a change back to the original regime. The hysteresis model describes a nonlinear response to the environmental driver but the threshold for recovery differs from the original threshold causing the collapse. The fourth model describes an irreversible change (adapted from Davis et al. 2010).

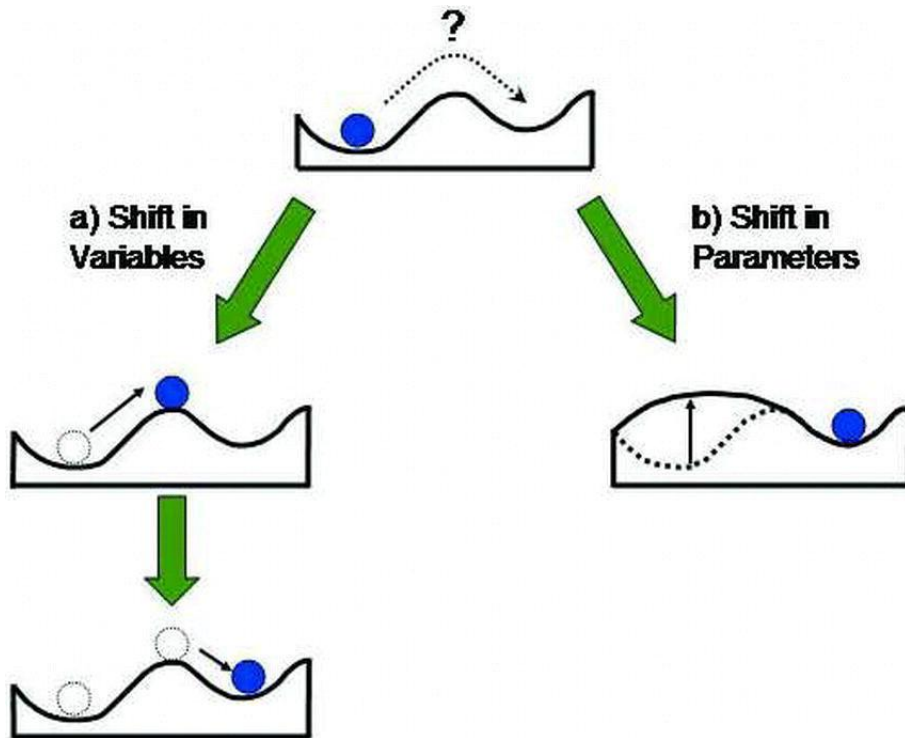
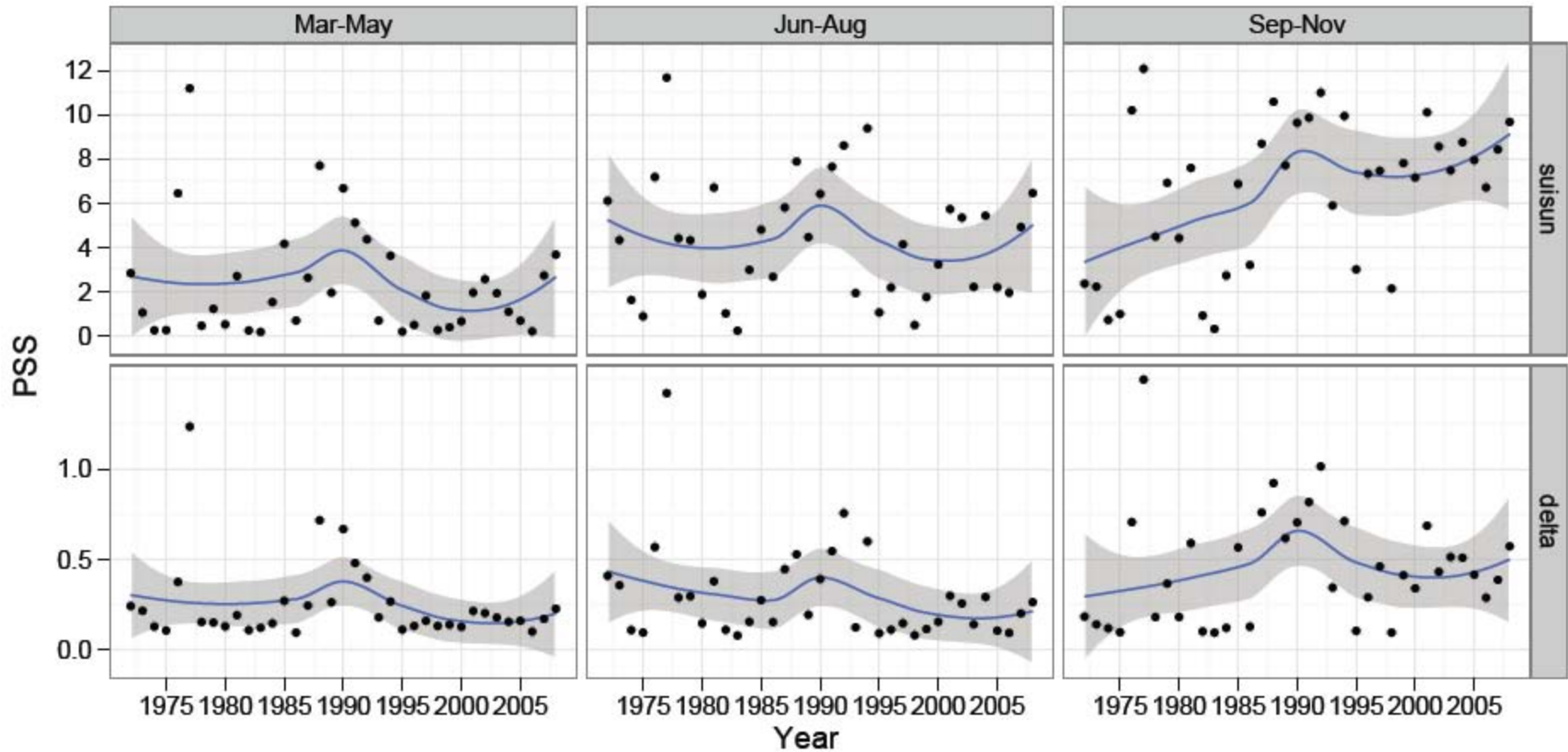


Figure 37. Two-dimensional ball-in-cup diagrams showing (left) the way in which a shift in state variables causes the ball to move, and (right) the way a shift in parameters causes the landscape itself to change, resulting in movement of the ball (from Beisner et al. 2003).



38. Annual average salinity by season (spring, summer, fall) and subregion (see Fig. 1) in the upper San Francisco Estuary between 1972 and 2008. Top panels represent data from stations in the “suisun” subregion, which include Suisun Bay and the western Delta. Bottom panels represent data from stations in the “delta” region, which includes the remaining areas of the Delta. Blue line displays a loess fit standard error (shaded area) (from Winder and Jassby 2010).

APPENDIX 1

2010 Element Descriptions of Ongoing and New Interagency Ecological Program (IEP) Studies for the Pelagic Organism Decline (POD)

6 December, 2010

I. Expanded Monitoring

Fall Midwater Trawl (FMWT)

IEP 2010-03

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: What is the relative abundance (via abundance index) of striped bass, delta smelt and other pelagic fishes of the upper estuary? How are juvenile delta smelt, longfin smelt and striped bass distributed in relation to potential food items?

Description: This survey targets age-0 striped bass and other pelagic species 30-150 mm in length using a midwater trawl towed through the water column for 12 minutes in a stepped oblique manner (Stevens and Miller, 1983). Sampling takes place at 116 stations located from the upstream of San Pablo Bay through the Suisun Marsh and Bay and into the Delta. This survey historically produced annual abundance indices of upper estuary pelagic fishes, which were used to identify the decline in pelagic organisms (Sommer et al., 2007); more recently the survey has provided fish samples for otolith, diet, condition and histopathology studies, and collected zooplankton and mysid samples to investigate food organism presence and diet, and ranked *Microcystis* apparent densities (see below).

During September 2005 and September and October 2006 and 2007, zooplankton tows were made at a subset of stations and mysid samples were collected starting in 2007 (32 sites in 2005, 18 sites in 2007). Beginning in 2005, the heads and bodies of delta smelt and striped bass were preserved separately for otolith and histopathological analyses (Bennett, 2005, section 3a; Teh, 2005, section 3b) or were preserved intact for diet and condition analyses (Gartz and Slater, 2005, section 3c). In 2006-2007, sampling continued as described above and longfin smelt, threadfin shad, and some inland silversides were collected for diet and condition. In 2008, fish preservation remained the same as 2006 and 2007, except longfin smelt heads and bodies were preserved separately for otolith, genetics, diet and sex ratio analyses.

In 2009, only striped bass heads and bodies were preserved separately; longfin smelt and wakasagi were preserved in ethanol, threadfin shad and some American shad were preserved in formalin and all delta smelt and most American shad were immediately frozen in liquid nitrogen. Due to furloughs and sample backlogs, mysid and zooplankton samples were only taken when delta smelt were caught and these samples were supplied to Wim Kimmerer along with the frozen delta smelt. Furthermore, turbidity measurements (Hach Model # 2100P Turbidimeter) were initiated at all stations and 6 supplemental locations in the Sacramento Deepwater Ship Channel (SDWSC) were added to the survey to target delta smelt.

In 2010, sampling will continue to separate heads and bodies of a subset of delta and longfin smelt. Up to 20 striped bass, Mississippi silverside and threadfin shad per station will be preserved in formalin for diet and condition analyses. In addition, the combined mysid and

Clarke-Bumpus (CB) tows will again be conducted at a subset of 10 stations positioned along a transect from the lower Sacramento River through Suisun Bay. These stations overlap the fall distributions of delta smelt and age-0 striped bass. Additionally, we will continue the visual survey technique started in 2007, ranking the density of *Microcystis* observed at each station. The same technique is also used by the Environmental Monitoring Program (EMP).

Time period: Sampling is conducted monthly from September through December and takes 2 weeks to complete.

Resources required:

Cost: The 2010 FMWT budget is \$22,000 from POD sources.

Principal Investigator (PI)(s): Randy Baxter, Steven Slater and Dave Contreras (DFG)

Contract needed / in place: Reimbursable contracts with DWR and USBR in place.

Contract manager(s): Rich Breuer (DWR), Erwin Van Nieuwenhuysse (USBR) and Kelly Souza (DFG)

Term of contract: July 1, 2007–December 31, 2010 (DWR) and through December 31, 2011(USBR).

Personnel: The field component of this project requires 1 boat operator, 1 biologist, and 1 scientific aide. The laboratory component requires numerous personnel for pre-season preparation, fish identification, data validation, diet and condition procedures, stomach content analysis and zooplankton processing.

Equipment: A boat with davits and hydraulics appropriate to pull a midwater trawl net (such as the *R/V Scrutiny*), laboratory facilities, warehouse space, formalin, ethanol, and suitable containers for sample collection and preservation.

Deliverables and dates:

- Monthly survey indices will be calculated and checked by the end of each month.
- Annual indices (January 2011).
- Database and flat file of species catch per tow (February 2011).
- Status and trends article for the IEP Newsletter (spring 2011).
- Trends in distribution and abundance of jellyfish will be examined for data collected since 2001.
- Zooplankton and mysid identification has been delayed by lab staff shortages. Once identification is complete, catch per unit effort (CPUE) calculations will be completed and the data will contribute to the fish and food item match–mismatch analysis, which has been temporarily removed from the work plan.

Comments: The FMWT Survey collected delta smelt and striped bass for otolith and histopathology investigations, but did not collect sufficient numbers of either to support all the projects in 2005 - 2007, so additional field collections were necessary. Targeted supplemental sampling for fishes is not planned in 2010 due to smelt collection restrictions. This survey currently reports annual abundance indices for 6 fishes and has collected count data on jellyfish since 2001. The ratio of same-year FMWT to the Summer Townet Survey (TNS) indices for age-0 striped bass is used as an index of summer survival (Stevens et al., 1985).

Summer Townet Survey

IEP 2010-007

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: What is the relative abundance of striped bass and delta smelt? Can these data be used to estimate apparent mortality? How are juvenile striped bass and delta smelt distributed in relation to potential food items? Is the density of food items related to fish condition, growth rate or health indices?

Description: The TNS has collected juvenile fishes in the range of 20 to 50 mm since 1959 (Turner and Chadwick, 1972) and currently provides indices of abundance for age-0 striped bass (38 mm index) and delta smelt. Samples are collected using a conical net with a 1.5 m² mouth and 12.7 mm (½ in.) stretched mesh nylon lashed to a hoop frame and mounted on skis. Three, 10-minute oblique tows are made against the current at each of the 32 stations located from eastern San Pablo Bay to Rio Vista on the Sacramento River and Stockton on the San Joaquin River.

This survey was expanded in 2005 to include simultaneous zooplankton sampling at each station, water collections for invertebrate toxicity tests (Werner, 2005, section 3e) at a subset of 10 stations, and a water quality profile at every station. Since 2006, water sample collections for invertebrate toxicity were made by directed sampling, which are not part of this survey. In 2006, we began using a YSI 6600 Sonde that collects temperature, depth, dissolved oxygen, turbidity, chlorophyll *a*, conductivity, salinity, pH, date, and time. The Sonde was very slow to respond and time consuming to keep calibrated, so in 2008 it was replaced with a YSI 30 meter that measured temperature and conductivity. Also in 2006, the heads and bodies of delta smelt and striped bass were preserved separately for otolith and histopathological analyses (Bennett, 2005, section 3a ;Teh, 2005, section 3b) or were preserved for diet and condition analyses (Gartz and Slater, 2005, section 3c).

Beginning in 2007 and continuing through 2008, longfin smelt, inland silverside and threadfin shad were collected for diet and condition; the heads and bodies of delta smelt and striped bass were preserved separately for otolith and histopathology analyses; a visual survey technique ranked the density of *Microcystis* observed at all sampling stations and numeric estimation of jellyfish abundance began.

Starting in 2009, turbidity measurements (Hach Model # 2100P Turbidimeter) were taken at each station. In 2010, only a subset of delta smelt and longfin smelt (5/station) will have heads and bodies preserved separately, and up to 5 additional fish of each species along with up to 20 striped bass, threadfin shad, and Mississippi silversides will be preserved whole in formalin for diet and condition.

Time period: Every other week from June through August.

Resources required:

Cost: The 2010 TNS budget is \$24,000 from POD sources.

PI(s): Randy Baxter, Steven Slater and Virginia Afentoulis (DFG)

Contract needed / in place: Reimbursable contracts with DWR and USBR in place or in progress.

Contract manager(s): Rich Breuer (DWR), Erwin Van Nieuwenhuyse (USBR) and Kelly Souza (DFG)

Term of contract: July 2010 – June 2011 (DWR) and through December 31, 2011 (USBR).

Personnel: The field component of this project requires 1 boat operator, 1 biologist, and 1 scientific aide. The laboratory component requires numerous personnel for preseason

preparation, larval fish identification, zooplankton identification, data validation, length weight procedures, and stomach content analysis.

Equipment: A boat with an A-frame and hydraulics appropriate to pull a sled-mounted townet (such as the *R/V Scrutiny or Munson*), laboratory facilities, warehouse space, formalin, ethanol, and suitable containers for sample collection and preservation. A Hach Model # 2100P Turbidimeter is used to measure the turbidity of the water at each station.

Deliverables and dates:

- Survey indices for striped bass (38 mm index) and delta smelt (September 2010).
- Status and trends article for the IEP Newsletter (spring 2011).
- Trends in distribution and abundance of jellyfish will be examined for data collected since 2001.
- 2006 and 2007 zooplankton identification and CPUE calculations (December 2010).
- Fish samples collected in 2010 will be provided to researchers along with associated environmental data (November 2010).

Comments: The TNS collected delta smelt and striped bass for otolith and histopathology investigations, but did not collect sufficient numbers of either to completely support those projects in 2005, so additional field collections were necessary and were added in 2006 and 2007. A similar circumstance is expected for out years. TNS catch data are used to calculate the striped bass 38.1 mm index (Turner and Chadwick, 1972) and an annual abundance index for juvenile delta smelt (Moyle et al., 1992).

20 mm Survey

IEP 2010-033

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: What is the abundance and distribution of POD fish larvae and early juveniles, particularly delta smelt and striped bass? Can catches of larval fishes and zooplankton be used to estimate overlap and possibly recruitment success? What do POD fish larvae eat, and are diet related to availability in the environment? Does turbidity effect larval and juvenile delta smelt feeding success?

Description: This survey targets late-stage larval and early juvenile fishes, particularly delta smelt and has historically focused on providing near real-time information on the distribution of young delta smelt to inform water management decisions. More recently, fish and associated zooplankton data collected have become very important in the investigation of food limitation hypotheses. Starting in March or April and continuing every other week through July, 3 oblique tows are made at each of 47 locations in the upper estuary using a conical, 5.1 m long plankton net composed of 1600 micron mesh with a 1.5 m² mouth, mounted on a weighted tow-frame with skids (Dege and Brown, 2004). In addition, zooplankton is collected during the first tow at each location with a CB net composed of 150 micron mesh attached to the top of the 20 mm net frame. General Oceanics flowmeters mounted in the mouth of each net provide estimates of volumes (m³) filtered. Fish and zooplankton samples are preserved in 10% formalin and identified in the lab.

This survey has sampled annually since 1995 and provides near real-time larval/juvenile delta smelt distribution information via the web (<http://www.delta.dfg.ca.gov/data/20mm/>) and Data Assessment Team (DAT). Samples from 2005 and 2006 provided young delta smelt and longfin smelt for diet analyses and zooplankton data for comparison to diet data (Slater in prep.).

Length data from 1995-2006 contributed to longfin and delta smelt apparent growth analyses. In March 2008, sampling commenced at 6 new north Delta locations to improve spatial coverage. These new locations are permanent additions to this survey.

Time period: Every other week from early March through early July.

Resources required:

Cost: The budget for the 2010 20 mm Survey is \$488,000 from non-POD sources.

PI(s): Bob Fujimura, Julio Adib-Samii (DFG)

Contract needed / in place: In place.

Contract manager(s): Rich Breuer (DWR), Erwin Van Nieuwenhuysse (USBR), Kelly Souza (DFG)

Term of contract: July 2010 – June 2011 (DWR) and through December 31, 2011 (USBR).

Personnel: When the survey starts in late March, it will run 2 boats for the first 3 days and a single boat for the remaining day of survey. Each boat requires 1 boat operator, 1 lead person (biologist, lab assistant or well trained scientific aide) and 1 scientific aide. Numerous lab personnel are needed for preseason preparation, lab processing of samples, and zooplankton and larval fish identification. In 2010, the 20 mm Survey employed 8 fulltime scientific aides to cover laboratory and field personnel needs.

Equipment: This project requires the use of the *R/V Munson* and *R/V Scrutiny* or *RV Beowulf*. Wet lab space is required to process approximately 1,300 larval fish samples and 432 zooplankton samples that are collected throughout the field season. A Hach Model # 2100P Turbidimeter is used to measure the turbidity of water at each station.

Deliverables and dates:

- Larval and juvenile fish distribution data will be posted to the web within 72 hours of field sampling.
- Weekly tabular summaries of smelt catch per 10,000 m³ will be distributed to the Smelt Working Group (SWG) and DAT.
- Weekly data summaries will be posted in graphic and tabular form to the 20 mm Survey webpage (<http://www.delta.dfg.ca.gov/data/20mm/>).
- Highlights article for the IEP Newsletter describing the survey outcome, abundance and relevance of the sampling (spring 2011).
- Relational ACCESS database (December 2010).

Comments: Beginning in 2008, the 20 mm Survey went from a 6 day to a 7 day sampling schedule because of the addition of 6 new north Delta stations. This addition, along with the implementation of the state furlough program, and the need to reduce weekend and overtime work, forced several sampling regime changes. Prior to 2008, each bi-weekly survey was conducted by 1 boat over 5 – 6 days. In 2008, each bi-weekly survey was conducted by 2 boats for the first 2 days and 1 boat for the remaining 3 days of survey. In 2009 and 2010, sampling was conducted by 2 boats for the first 3 days and 1 boat for the remaining day of survey. The 2010 sampling schedule will remain in place for the 2011 field season. In 2010, turbidity readings (in Nephelometric Turbidity Units (NTU)) were taken at each sampling station. Turbidity measurements will be permanently added to the 20- mm protocol.

Spring Kodiak Trawl

IEP 2010-088

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: This survey was developed to identify and monitor the distribution and relative abundance of adult delta smelt throughout the upper estuary during winter and spring. Two other main goals of the survey are (1) to determine gender ratios and (2) to determine stages of sexual maturation of the adult delta smelt population and (3) how male and female maturation varies in time and space. In 2010, we began addressing the question, “Does turbidity effect adult delta smelt migration?”

Description: The Spring Kodiak Trawl Survey (SKT) samples monthly beginning in January and continuing into May. Each monthly Delta-wide survey takes 4-5 days and samples 40 stations from the Napa River to Stockton on the San Joaquin River and Walnut Grove on the Sacramento River. Delta-wide surveys are conducted to locate the areas of adult delta smelt concentration. Historically, Delta-wide surveys were followed by a supplemental survey 2 weeks later that sampled areas intensively for high delta smelt concentration to estimate the proportion of male and female delta smelt and their maturity stages. To minimize take of spawning adults, supplemental surveys are now only conducted under the recommendation of the SWG and the approval of managers. A standard Kodiak trawl, with a mouth opening of 25 ft. by 6 ft., overall length of 65 ft., and 1/4 in. cod-end mesh is used to make 10-minute surface tows at each station. The catch is speciated, enumerated, and measured. Each adult delta smelt is examined onboard for sexual maturity, given a unique alpha-numeric identifier, and dissected for preservation. Quality assurance and control (QA/QC) measures are performed on preserved delta smelt in the laboratory for maturity stage.

Resources required:

Cost: The 2010 SKT budget is \$282,000 from non-POD sources.

PI(s): Bob Fujimura and Julio Adib-Samii (DFG)

Contract needed / in place: In place.

Contract manager (s): Kelly Souza (DFG), Erwin Van Nieuwenhuysse (USBR) and Rich Breuer (DWR)

Term of contract: July 2010 to June 2011 (DWR) and through December 31, 2011 (USBR).

Personnel: Each 4 or 5 day survey requires 2 boats, 2 boat operators, 1 biologist and 1 scientific aide. The biologist, assisted by permanent and seasonal staff, is responsible for pre-season preparation, lab processing of samples, and fish identification and archiving. In 2010, the Smelt Larva Survey shared 5 fulltime scientific aides with the SKT to cover laboratory and field personnel needs.

Equipment: This project requires the use of the *R/V Munson*, *R/V Scrutiny* or *R/V Beowulf*, as well as a suitable “chase boat” similar to 1 of the Kvichaks. Wet lab space is required to process no fewer than 200 fish samples that are collected throughout the field season. A Hach model # 2100P Turbidimeter is used to collect turbidity measurements from each station.

Deliverables and dates:

- Project staff provides partial tabular summaries of delta smelt catch by station for weekly use by the SWG and the DAT.
- Monthly data summaries will be posted in graphic and tabular form to the SKT web page (<http://www.delta.dfg.ca.gov/data/skt/>).
- A Highlights article for the IEP Newsletter will describe the survey outcome and relevance of the sampling (spring 2011).

- Relational ACCESS database, complete with fish and environmental measurements (December 2010).

Comments: There are no planned departures from 2010 protocol scheduled for the 2011 field season.

Directed Field Collections

IEP 2010-089

Point person: Randy Baxter (DFG)

Lead Agency: DFG and DWR

Questions: There are no questions related to the Directed Fish Collections effort. Questions are listed under the project description that the fish are being collected for elements 2010-040, 2010-060, 2010-061, 2010-062, and 2010-131.

Description: In 2005 and 2006, directed, short-term field collections were used to increase the number of delta smelt and striped bass available for otolith analyses and histopathological studies, collect threadfin shad and longfin smelt for disease and histopathological studies (Foot et al., 2006), provide samples of POD fishes for diet and condition, and collect water samples from 15 locations within the upper estuary for invertebrate and fish toxicity tests. Sampling efforts were made once or twice a month to increase the numbers of target fish collected and to allow time for field examination of larval and young juvenile fishes.

In 2007, delta smelt were removed as a target organism to reduce IEP's overall take of these fish. However, longfin smelt, striped bass, threadfin shad and water for both invertebrate and fish toxicity tests were still collected. In 2009, directed field collection effort was used to provide supplemental sampling of the SDWSC for delta smelt from spring through early winter and support field work for the "Use of Acoustics to Estimate Trawl Openings" (see 2010-131). In 2010, directed collections effort will be used for field work associated with the "Use of Acoustics to Estimate Trawl Openings."

Time period: As needed and when staff and boats are available. For water collections, sampling will take place 4 days per month, targeted fish sampling can be up to 2 days per month and gear evaluation (2010-131) can be 4 – 10 days per year.

Resources required:

Cost: The 2010 IEP budget for Directed Field Collections is \$68,000 from POD sources.

PI(s): Dave Contreras, Jennifer Messineo, Steven Slater and Randy Baxter (DFG)

Contract needed / in place: Near execution as of August 16, 2010.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Kelly Souza (DFG)

Term of contract: April 1, 2010 – December 31, 2013 (when executed).

Personnel: The water sampling requires 1 boat operator and 1 crew member from the participating research group, while fish sampling requires 1 boat operator, 1 scientific aide or biologist and 1-3 researchers from the participating research group. In addition, the point person contributes substantial time coordinating logistics for PIs and field crews.

Equipment: a 25-42 ft. vessel (or 2 vessels for Kodiak Trawling) capable of deploying trawl gear for net mouth dimension measurement ("Use of Acoustics to Estimate Trawl Openings").

Deliverables and dates: See specific project descriptions listed above for this information.

Comments: None.

Smelt Larva Survey

IEP 2010-096

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: In 2005 and 2006, we addressed the question, “Can the distribution of larval delta smelt be effectively determined using surface-oriented plankton nets when compared to catches from traditional ichthyoplankton gear and methods?” In 2007, we addressed the question, “What is the vertical distribution of newly hatched larval delta smelt over a 24 hour time period?” The data collected to answer this question has not yet been analyzed. In 2009 and 2010, we addressed the question, “What is the distribution and relative abundance of larval longfin smelt in the upper San Francisco Bay and Delta?” An upper estuary-wide survey that provides near real-time distribution data for longfin smelt larvae in the Delta, Suisun Bay and Suisun Marsh was initiated. These data are used by agency managers to assess vulnerability of longfin smelt larvae to entrainment in south Delta export pumps. We also began collecting turbidity measurements to better understand the role that turbidity or water clarity might have on larval smelt feeding success.

Description: In 2005 and 2006, this survey, under the name Delta Smelt Larva Survey (DSLS), investigated whether surface-oriented, fine-mesh nets could be towed along with traditional 20 mm Survey gear (IEP,1987; Rockriver, 2004; Dege and Brown, 2004) and improve detection of small larval delta smelt since they can pass through the mesh of the 20 mm Survey net. Two field seasons of data collection (2005-2006) were conducted as the basis for evaluating the surface oriented nets with those towed obliquely. However, the January through mid-March sampling period did not overlap with the delta smelt hatching period (only 4 were collected in both years).

A third season in 2007 focused briefly (2 days) on 2 locations where delta smelt were believed to have spawned and when larvae were believed to be present. Based on 2007 sampling, surface-oriented larva tows seemed reasonably efficient, but the relatively small volume sampled only improved detection of larvae for 1 of 7 surveys when compared to standard 20 mm Survey methods. The alternatives, conducting two oblique tows in succession or 2 separate surveys for larva and 20 mm fish, were not feasible with current staff and boats. In 2005 and 2006, sampling began in January and early February to facilitate the capture of larval longfin smelt, and included concurrent collection of zooplankton samples at the 41 20 mm Survey stations plus 3 additional locations in the main channel of central and eastern San Pablo Bay. In 2007, sampling was limited to 2 days (April 16-17) when 10 tows during daylight and 10 tows after dark were made at each of the 2 locations. Sampling was not conducted in 2008.

In 2009, sampling was reinitiated under the name Smelt Larva Survey (SLS) and began in the first 2 weeks in January and repeated every other week through the second week in March targeting longfin smelt larvae. Each 4-day survey consists of a single 10-minute oblique tow conducted at each of the 35 locations using an egg and larva net. The 505 micron mesh net is hung on a rigid, U-shaped frame attached to skis to prevent it from digging into the bottom when deployed. The net mouth area measures 0.37 m². Immediately after each tow, juvenile fishes are removed, identified, measured and returned to the water and the remaining larvae are preserved in 10% formalin for identification. No changes in the 2009 protocol occurred in 2010 except sampling was extended into April. This survey will be ongoing in its current form.

Time period: Every other week from early January through early July (2005-2006), 2 days in April (2007), every other week from early January through mid-March (2009), and every other week from early January through April (2010).

Resources required:

Cost: The 2010 SLS budget is \$269,000 from POD sources and \$201,000 from non-POD sources.

PI(s): Bob Fujimura and Julio Adib-Samii (DFG)

Contract needed / in place: In place.

Contract manager(s): Kelly Souza (DFG), Erwin Van Nieuwenhuysse (USBR) and Rich Breuer (DWR)

Term of contract: July 2010 –June 2011(DWR) and through December 31, 2011 (USBR).

Personnel: Each 2-day survey requires 2 boats, 2 boat operators, 1 lead biologist and 1 scientific aide per day. Numerous lab personnel are needed for preseason preparation, lab processing of samples, and larval fish identification. In 2010, the SLS employed 5 fulltime scientific aides to cover laboratory and field personnel needs.

Equipment: This project requires the use of the *R/V Munson* and *R/V Scrutiny* or *R/V Beowulf*. Wet lab space is required to process approximately 250 larval fish samples that are collected throughout the field season. A Hach model # 2100P Turbidimeter is used to collect turbidity measurements from each station.

Deliverables and dates:

- Distribution data, particularly larval longfin smelt, will be posted to the web within 72 hours of field sampling.
- Weekly tabular summaries of smelt catch and catch per 1,000 m³ will be distributed to the SWG and the DAT..
- Weekly data summaries will be posted in graphic and tabular form to the SLS webpage (<http://www.delta.dfg.ca.gov/data/sls/>).
- A Highlights article for the IEP Newsletter describing the survey outcome and relevance of the sampling (spring 2011).
- A relational ACCESS database complete with fish and environmental measurements (December 2010).

Comments: The 2005 and 2006 Delta Smelt Larva Survey replaced the North Bay Aqueduct monitoring on a pilot basis as required by the USFWS 2005 Operations Criteria and Plan (OCAP) Biological Opinion for delta smelt. The SWG designed this survey as a 2-year trial. Protocol and methods developed in 2005 were used in 2006. Surface tows proved ineffective for larval delta smelt (unpublished DFG data), so the USFWS asked the SWG to modify the sampling design for 2007, wherein oblique tows using standard larval sampling gear (plankton net composed of a 500 micron mesh possessing a mouth 0.37 m² attached to a skid mounted frame) were compared with traditional 20 mm sampling. The sampling for delta smelt larva in 2007 was guided by catches of ripe and spent adult delta smelt that were caught in the SKT. SLS sampling was not conducted in 2008. Sampling was reinitiated in 2009 specifically to provide information on the distribution of larval longfin smelt and their potential vulnerability to entrainment in south Delta exports as part of the minimization measures during the state candidacy period as a threatened species. The 2010 SLS was mandated by the CA DFG State Water Project (SWP) longfin smelt incidental take permit (ITP) (2081). There is no planned departure from 2010 protocol for the 2011 season.

II. On-going Studies

Development and Implementation of Life-cycle Models of Striped Bass and Longfin Smelt in the Bay-Delta Watershed

IEP 2010-038

Point person: Ted Sommer (DWR)

Lead Agency: UCD

Description: Recent declines in the abundances of several pelagic fish species in the Bay-Delta have increased the need for data-supported quantification of the relationships between dynamics of the striped bass and longfin smelt populations and the ecosystem components that affect striped bass and longfin smelt. To better explore these dynamics, a striped bass individual-based life-cycle model (IBM) was developed. More recently, the striped bass population life-cycle model has been extended and modified to be applicable to longfin smelt in the Bay-Delta. Both models are at points where different scenarios could be run through them. An example of potential scenarios include effects of temperature, food web shifts, diversions/operations, modified habitat preferences, contaminant accumulation, mortality rates, peripheral canal, and changes in slot limits of the adult fishery. The original IBM contract ended in February 2009 and a 2 year amendment began in March 2009. The remaining duration of the amendment includes the following tasks:

- Task 1: Longfin smelt abundance measures and rate expressions for specific life-stage processes (e.g., growth, mortality, and fecundity) have been obtained from all permissible and existing data sources. Bioenergetics rate parameters are currently based upon Lantry and Stewart's (1993) work with rainbow smelt in the Great lakes, however a separately funded study will be providing bioenergetics rate parameters specific to longfin smelt in the Bay-Delta. This longfin smelt IBM will be applied to explore the relative significance of specific factors influencing population numbers of longfin smelt and in the prioritization of future data collection.
- Task 2: In February 2010, a series of scenarios was developed to be run in the striped bass IBM. The cumulative goal of these scenarios is to explore the cause of the apparent disconnect between the FMWT index and the adult striped bass abundance index. To fully explore this disconnect, the IBM is being applied under a variety of scenarios, which include evaluating the effect of changes in sex ratios, changes in various life-stage specific mortality rates and changes in spawning patterns.

Time period: The completion date of this project is June 2011.

Resources required:

Cost: \$125,000 (DWR → UCD)

PI(s): Frank Loge (UCD)

Contract needed / in place: In process.

Contract manager(s): Erwin Van Nieuwenhuyse (USBR) and Rich Breuer (DWR)

Term of contract: TBD

Personnel: Tim Ginn, Kenneth Rose, Arash Massoudieh, and Kai Eder

Equipment: This is a data mining exercise.

Deliverables and dates:

- Progress report, Delta Science conference presentation (September 2010).
- Progress report, IEP newsletter article and presentation (March 2011).
- Submission of 3 manuscripts, computer codes for each model, supporting documentation explaining use, inputs and outputs (June 2011).

Modeling the Delta Smelt Population in the San Francisco Estuary

IEP-related 2010-041

Point person: Anke Mueller-Solger (DSC)

Lead Agency: SFSU-RTC, LSU, Stanford and UCD

Questions: What are the best management strategies for conserving this species?

Description: This element was designed to develop and test 3 different modeling approaches for looking at delta smelt population dynamics. The modeling approaches can be generally characterized as particle tracking, matrix projection and individual-based. The purpose of these models is to evaluate how environmental conditions influence population vital rates, which then determine how the modeled population responds.

Time period: 2006-2010

Resources required:

Cost: This contract was not reinstated from the 2008 bond freeze. No funding is allocated to it in 2010.

PI(s): Karen Edwards and Wim Kimmerer (SFSU), Bill Bennett (UCD), Kenny Rose (LSU), Stephen Monismith (Stanford)

Contract needed / in place: Cancelled (see comment below).

Contract manager(s): Shem Ayalew (DSP)

Term of contract: 3 years, beginning in 2006 and expiring on March 31, 2009.

Personnel: Karen Edwards (SFSU)

Equipment: None, this is a modeling exercise.

Deliverables and dates: Originally listed: papers on the individual-based model, matrix modeling and particle tracking. Due dates no longer relevant (see below).

Comments: The bond freeze in late 2008 ran through the end of the contract and DWR elected not to reinstate the contract. We are currently closing out the project even though it is incomplete. We still intend to produce the papers originally promised, but there is no funding to do so. Hence, we must do so on our own time and it will have lower priority than funded projects.

Estimation of Pelagic Fish Population Sizes

IEP 2010-043

Point person: Pete Hrodey (USFWS)

Lead Agency: DFG, DWR, USBR, consultants and contractors

Questions: What are the most efficient regions (strata) for each target species sampled by the TNS, FMWT and Kodiak surveys? Do fixed sampling stations in a highly tidal system approximate random distributions? What are the population sizes for each of the target pelagic species? Should strata variance be calculated based upon a normal distribution? How can long-term monitoring data be used to provide better information to managers regarding trends in biological resources? Additional questions will be developed over time.

Description: Except for adult striped bass, the status of pelagic fish populations has primarily been assessed using relative abundance indices. IEP has been reluctant to translate these data into population estimates since gear efficiencies are unknown for each of the sampling programs, fish distribution tends to be patchy, and surveys do not have random site selection, likely adding substantial variability. Other approaches for pelagic fish population estimation are unreasonable (e.g., direct counts, mark-recapture, and change in ratio). However, the POD effort would benefit

greatly from at least crude population estimates, allowing calculation of mortality rates and population modeling. The development of mean-density expansion estimators based upon stratified random trawl sampling represents the most practicable alternative. As initial steps to estimate population size, Bennett (2005) has used the TNS and FMWT data and Miller (2005) has analyzed the Kodiak trawl data. This element will:

- Build upon earlier efforts to develop population estimates for as many of the target pelagic species as possible, beginning with delta smelt. Refinements of these efforts may include the use of known salinity and temperature effects on target species distributions, updated bathymetry and the particle tracking models to: (a) post-stratify survey data (i.e., set more efficient region boundaries); (b) improve habitat volume estimates represented by fixed stations and regions for each of the surveys; and (c) test the assumption of randomness in the data.
- Study designs for gear evaluation including development of mesh retention probabilities, gear avoidance measures and information on vertical and lateral distributions.
- IEP/POD statistical and analytical support – time to assist agency researchers with study design and analysis questions.
- NCEAS participation and modeling that incorporate delta smelt life history information, information from past and future special studies examining horizontal and vertical distribution behavior and multiple long-term monitoring data sets into a single population model.

Time period: 2008-2010

Resources required:

Cost: \$188,000 in 2010 from POD sources.

PI(s): Ken Newman (USFWS)

Contract needed / in place: Reimbursable contracts are in place for 2010.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Rich Breuer (DWR) and John Netto (USFWS)

Term of contract: 5-year reimbursable with annual modifications.

Personnel: Wim Kimmerer (SFSU), Bill Bennett (UCD), Fred Feyrer (USBR), Jim Thompson and Ralph MacNally (Monash University), Wendy Meiring (UCSB)

Equipment: None

Deliverables and dates: Life history model for some of the fish populations, beginning with delta smelt.

Comments: Although there are substantial obstacles to measurement of population sizes, the recent efforts of Bennett (2005) and Miller (2005) provide a reasonable foundation for future work. Newman's involvement with IEP/POD is anticipated to be long-term, during which time tasks will evolve. For 2010, substantial new tasks will be reviewed by the POD management team (MT) and when approved forwarded to Newman through Kim Webb with an indication of relative priority.

Zooplankton Fecundity and Population Structure

IEP 2010-044

Point person: Anke Mueller-Solger (DSC)

Lead Agency: SFSU-RTC

Questions: Has there been a downward shift in egg production and/or nauplius survival that resulted in lowered ratios of copepodites to adults? Has there been a change in copepodite

survival? What is the biomass of individual zooplankton species, and how has the zooplankton community biomass changed over time?

Description: This work plan element consists of 2 parts:

- The continuation of an analysis of *Pseudodiaptomus forbesi* and *Eurytemora affinis* life stage structure and fecundity from archived zooplankton samples (1996-2005) and associated water temperature data. The goals are to determine whether the recent increase in *Corbula* abundance was associated with an increase in mortality of sub-adult (i.e., copepodite stage) *P. forbesi* and *E. affinis*, and/or a reduction in adult *P. forbesi* and *E. affinis* fecundity.
- The determination of the biomass of zooplankton species and their life stages using laboratory analysis of dry weight and carbon content of freshly collected specimens from the San Francisco Estuary. The goal is to provide species-specific biomass conversion factors that will then allow for more appropriate and realistic analysis of biomass, production, and consumption trends and comparisons with other estuaries.

Time period: 2006-2010

Resources required:

Cost: \$41,000 in 2010 from POD sources.

PI(s): Wim Kimmerer (SFSU)

Contract needed / in place: In place.

Contract manager(s): Ted Sommer

Term of contract: Through December 30, 2010.

Personnel: SFSU

Equipment: None

Deliverables and dates:

- Year 1 progress report will be provided as 1 or more IEP Newsletter articles or manuscripts, depending on the results to date (October 2010).
- Present preliminary results at Delta Science Conference (October 2010).
- Final reports, which will comprise draft manuscripts for submission to journals or the IEP Newsletter (October 2010).

Comments: We have completed analysis of a selected set of zooplankton samples; once QC measures on the data are complete, we will begin analysis of mortality patterns. Biomass analyses are about half complete, but have been delayed several times by instrument problems and staffing issues.

Phytoplankton Primary Production and Biomass in the Delta

IEP 2010-045

Point person: Anke Mueller-Solger (DSC)

Lead Agency: UCD, DWR-DES

Questions: Is there a signal of ammonium discharge from the Sacramento Regional Wastewater Treatment Plant (SRWTP) that can be discerned in the annualized fish indices? Is there a signal of ammonium concentration in estuary waters that can be discerned in these same indices?

Description: It has been suggested that total ammonium and/or free ammonia affect biota in the Delta through inhibition of physiological processes of plankton and higher organisms. This hypothesis has been forwarded on the basis of experimental evidence and general knowledge of ammonia and ammonium toxicity. As there are so many processes at work in the Delta, it is also necessary to show that this specific process

manifests in the long-term variability of Delta population abundances. Otherwise, the hypothesis may be true, but the ecological implications will be insignificant. This study will (1) review existing studies underlying the ammonium hypothesis, particularly those based on long-term data from the Delta; (2) search for an ammonium wastewater discharge and river concentration signal in the annualized fish indices and related data such as fish length; and (3) add an algorithm to Jassby's wq software package to calculate the free ammonia (toxic) fraction based on pH (and pH scale), salinity, depth, and temperature, which will be more appropriate for estuarine analyses than ones in common use.

Time period: Ongoing through 2011.

Resources required:

Cost: \$35,000 in 2010.

PI(s): Alan Jassby (UCD)

Contract needed / in place: In place.

Contract manager(s): Ted Sommer (DWR)

Term of contract: Through June 30, 2011.

Personnel: Alan Jassby (UCD)

Equipment: None

Deliverables and dates:

- Technical report describing results of the above research (July 2011).
- IEP newsletter article describing results of the above research (July 2011).
- Additional journal publication (letter, comment, note or article) is optional, depending on results.
- A journal article (Winder and Jassby, 2010, in revision) for *Estuaries and Coasts* and a conference presentation (American Society of Limnology and Oceanography 2010, Santa Fe) have already been completed based on research conducted during July 1, 2009 - June 30, 2010.

Comments: This work is a natural accompaniment to Jassby's current participation in a summary analysis of the long-term water quality data set collected by the USGS for over 30 years in the San Francisco Bay.

Synthetic Analysis of POD data (Workgroups convened by UC Santa Barbara, formerly NCEAS)

IEP 2010-046

Point person: Larry Brown (USGS)

Lead Agency: USFWS, USGS, DFG, DWR

Questions: Erica Fleishman at the Bren School of Environmental Science & Management at UC Santa Barbara (formerly with NCEAS, which also is part of UCSB) assembled working groups to address 3 general questions. What is the role of contaminants in the POD? What are the direct and indirect drivers of system dynamics in the Delta ecosystem? Does the nearshore ocean and San Francisco Bay affect abundance of POD species, and do atmospheric factors affect abundance of invertebrates and fishes in the Bay? More specific questions are being developed by subgroups within each working group. A new group will consider if the changes in the Bay-Delta constitute a regime shift, what alternative future system states might be possible, and what inputs might be required to achieve those states. The existing groups will finish ongoing work after which they will disband or address new questions.

Description: The overall goal for the UCSB-convened working groups is to conduct or guide the integration, analysis, and synthesis of POD and other relevant data and information in a more efficient, sophisticated, unbiased, and synergistic manner than would be possible with local resources alone. The UCSB-convened working groups on system dynamics, contaminants, and ocean-estuary interactions were formed after consultation with the POD MT and the steering committee for the project indicated that these were areas where the working group process could be most helpful to the IEP because of the lack of IEP expertise. Similarly, the topic of regime shift was identified as a new area deserving attention and will likely provide a useful framework for synthesizing the entire POD effort. The effort will tie together and analyze field data, environmental data, operations information and contaminants/bioassay from POD research components and other sources. These efforts will feed directly into POD synthesis/summary reports through Larry Brown and other POD MT members participating in the UCSB-convened workgroups and steering committee.

Time period: Continuation agreement will extend work into 2012.

Resources required:

Cost: \$751,000 for 2010 from POD sources.

PI(s): Larry Brown (USGS), Ken Newman and Gonzalo Castillo (USFWS), Fred Feyrer (USBR) and other agency personnel.

Contract needed / in place: Continuation agreement required.

Contract manager(s): Paul Cadrett (USFWS) will manage the UCSB-CESU contract. Other investigators are funded through annual reimbursable contracts with USBR and DWR.

Term of contract: Scientifically sophisticated approaches and defensible conclusions require substantial time. It is possible the IEP/POD-UCSB interaction could continue beyond the term of this contract (2012) if additional issues are identified.

Personnel: Other key staff members include Bruce Herbold (EPA), Anke Mueller-Solger (DSC) and Ted Sommer (DWR)

Equipment: None, these are data-mining efforts.

Deliverables and dates:

- Peer-reviewed journal articles and presentations geared at scientific and lay audiences that are authored by individual POD members of working groups convened by UCSB.
- Final Synthesis Report, lead by Larry Brown, will include material from POD sponsored studies and working groups convened by UCSB.

Comments: Existing staff members from the IEP agencies are redirected, as needed, to work closely with UCSB and the working groups to participate in the synthesis of IEP data as it relates to the POD. All such assignments are cleared through supervisory channels.

Evaluation of Delta Smelt Otolith Microstructure and Microchemistry

IEP 2010-060 and 2010-061

Point person: Randy Baxter (DFG)

Lead Agency: UCD

Questions: Do growth rates of delta smelt vary seasonally or geographically? When and where in the estuary is delta smelt produced?

Description: Analysis of delta smelt otoliths can determine daily growth rate and area of origin. In addition, analysis of otoliths that includes microchemical work can provide detailed

information on fish origin and growth that can be related to histopathology analyses and potentially to ambient water toxicity for 2005 and 2006. This work has been done successfully on delta smelt (Bennett submitted).

Otolith age and incremental growth measures for young fish will be derived through the use of imaging software. Furthermore, otolith weighing and morphometric methodologies will be developed and evaluated for reliability in determining age of adult fish (>300 days post hatch). Chemical composition of otoliths at their core will be measured to provide a micro-chemical “signature” of natal habitat and compared via trace elements and isotopic ratios to water samples collected at various locations in the lower rivers and Delta to determine likely region of natal origin. Fish samples for this element will be collected by the SKT, TNS, and FMWT, with supplemental sampling based on availability of boats and crews. In 2010, delta smelt will again be preserved, so otolith microstructure can be examined, but these samples are expected to contribute to a subsequent study.

Time period: Extended through June 30, 2011.

Resources required:

Cost: Estimate is \$292,000 from the Ecological Restoration Program (ERP) to process approximately 500-600 samples per year for aging.

PI(s): Bill Bennett, Swee Teh and James Hobbs (UCD)

Contract needed / in place: In place.

Contract manager(s): Mitsuko Grube (DFG)

Term of contract: Through June 30, 2011.

Personnel: Above named investigators and other UCD staff.

Equipment: Equipment for this project has already been purchased.

Deliverables and dates:

Semi-annual reports for 2009 are overdue. The last semi-annual report covered the time period from July 1, 2008 – December 31, 2008.

Comments: This work will be an extension of the delta smelt work (Bennett submitted) and striped bass work carried out by Bennett et al. (1995). This work is part of a larger contract including a histopathological element (2010-061) and a food availability and feeding element (not listed), but similar to the diet and condition work (2010-062), which together provide a comprehensive view and timeline of the relative condition of the fish that we could compare to timing of potential stressors.

Quantitative Analysis of Stomach Contents and Body Weight for Pelagic Fishes

IEP 2010-062

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: What is the feeding ecology of pelagic fishes in the estuary? Is there evidence of reduced feeding success during specific times of the year or in certain parts of the estuary? If so, are these changes associated with changes in growth rate, relative weight or condition?

Description: This study was initiated in 2005 to investigate temporal and spatial differences in the diet composition and feeding success of age-0 delta smelt, striped bass, and threadfin shad that might help explain the pelagic species decline. This list of target fishes has since been expanded to include longfin smelt, Mississippi silversides and American shad. Food habit studies have been done on many of the fish and zooplankton found in the estuary (IEP, 1987; Orsi, 1995; Lott, 1998; Nobriga, 2002; Feyrer et al., 2003). However, many of these studies were done more

than 10 years ago and the feeding habits of the local Mississippi silverside and threadfin shad populations have only been studied in a limited geographical range (Grimaldo, 2004).

As evidence that feeding success may be an important issue for survival, initial studies by BJ Miller suggest that delta smelt survival in different parts of the estuary was linked to whether there was co-occurrence of prey. In 2003, IEP started a study of fish length-weight relationships needed to estimate species biomass and to develop a program to monitor trends in relative weight. Collection of length and weight data was expanded beginning in 2005 to include calculation and analysis of fish condition for those fish processed for gut contents. When observations indicated that fish lengths and weight changed over time in preservative, a study was conducted in 2006 and 2007 to evaluate changes and their effect on several measures of fish condition. Work on diet and condition continued in 2006 and 2007, and was expanded to include larval fish collected by the 20 mm Survey and longfin smelt from salvage. In 2010, specimens will be collected through December and will be archived for future investigations or increased sample size when staff time permits. However, work in 2010 will focus on completing lab work, analysis and initial publication. Fish collected from 2010 will be archived for future processing and analysis.

Time period: 2006 - 2010

Resources required:

Cost: \$40,000 from POD sources.

PI(s): Randy Baxter and Steve Slater (DFG)

Contract needed / in place: Close to execution as of August 16, 2010.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Kelly Souza (DFG)

Term of contract: April 1, 2010 – December 31, 2012

Personnel: Field collection is conducted by all long-term fish monitoring surveys. Fish are retained after reaching quotas for otolith and histopathology samples; all POD fishes are retained by the San Francisco Bay Study (SFBS) for condition and diet analyses. Laboratory personnel (3 scientific aides and a part time senior lab assistant) are directed in sample processing by a biologist.

Equipment: Current long-term monitoring vessels and gear will be employed; some gear modification may occur for directed sampling. Laboratory equipment is currently available at DFG Stockton.

Deliverables and dates:

- Annual IEP Workshop poster presentation (March 2010).
- Manuscript describing the effects of formalin preservation on the determination of fish condition. Target journal is the North American Journal of Fisheries and Management (NAJFM) (August 2010).
- Draft delta smelt diet and condition manuscript. Target journal is the San Francisco Estuary and Watershed Science (SFEWS) (September 2010).
- IEP Newsletter article about longfin smelt, striped bass and threadfin shad diet composition (December 2010).

Comments: In 2006, examination of parasite load was transferred to researchers conducting histopathological investigations (2006-061, 2006-042). DFG staff will collect samples and process diet information. The IEP has extensive experience with these techniques, but lost leading staff for this work. Lab work was delayed substantially by the loss of a biologist and long delays in hiring scientific aides and senior lab assistants.

Trends in Benthic Macrofauna Biomass

IEP 2010-065

Point person: Karen Gehrts (DWR)

Lead Agency: DWR

Questions: What are the long-term trends in biomass, production and grazing rates of benthic species? How are these changes related to physical-chemical gradients? How do changes in benthic functions, such as production and grazing, affect the pelagic food web?

Description: Over the past 3 decades, the IEP EMP has collected benthos community composition and abundance information at 22 sites, including 4 long-term monitoring stations. Unfortunately, the EMP monitoring did not include measurements of benthic macrofauna biomass. Biomass data are crucial in determining the role of benthic organisms in the ecosystem, especially the feeding potential of various functional groups, potential availability and transmission of contaminants bioaccumulated in benthos, and trends in production as well as the ecological significance of changes in benthic community composition and abundance. The EMP has developed a comprehensive plan to analyze archived benthos samples dating back to 1975, which can be used for biomass estimation using a simple wet-weight method.

The objective of this project is to measure and examine the biomass of benthic organisms collected from the EMP benthic monitoring stations. Currently, staff collects an additional benthic sample and obtains ash-free dry mass for *Corbula amurensis* and *Corbicula fluminea*. This will allow the project team to create site specific biomass conversion factors for historic length data. To accomplish the goal of obtaining and examining complete biomass time series at 2 long-term stations, the Benthic Biomass project has received a grant from DSP to obtain length measurements for stations D4 and D28A from 1976 to the present. Data analysis is conducted in part as part of work plan element IEP 2010-078 and other ongoing EMP data analyses. The USGS will help to characterize long-term trends in biomass at the 2 EMP sites.

Time period: Data for site-specific biomass conversions will be collected on an ongoing basis. Data for the long-term analysis of stations D4 and D28 will continue through 2010.

Resources required:

Cost: Redirected staff from the EMP will be used to accomplish this work. Funding for the analysis of D4 and D28A is from a DSP grant, \$40,000.

PI(s): Karen Gehrts (DWR), Wayne Fields (Hydrobiology) and Janet Thompson (USGS)

Contract needed / in place: In place early 2010 (DSP →DWR).

Contract manager(s): Karen Gehrts (DWR)

Term of contract: Calendar year (CY) 2010

Personnel: Dan Riordan and Tiffany Brown (DWR)

Equipment: Staff will use existing oven, furnace, scale, microscope and computer to determine ash-free dry mass, length to weight conversions and historical shell length data.

Deliverables and dates:

- Final report to DSP (December 2010).
- Delta Science Conference presentation (October 2010).

Comments: Analysis of D4 and D28A invasive bivalve biomass does not depend on the availability of new field data. However, for the determination of site-specific biomass estimates, an additional grab sample needs to be taken during the monthly EMP benthic survey.

Corbula salinity tolerance, distribution and grazing rates

IEP 2010-076

Point person: Ted Sommer (DWR)

Lead Agency: DWR, USGS and SFSU

Questions: What is the salinity tolerance of *Corbula amurensis*? How well do salinity tolerances explain the distribution of *Corbula*? What are regional trends in benthos and grazing rates?

Description: Initial analyses of benthic abundance trends suggest that *Corbula* distribution shifted upstream around 2001, perhaps in response to recent salinity increases during autumn. To better evaluate this hypothesis, we need to develop salinity tolerance information for the clam. This will be performed in a controlled laboratory setting, likely at SFSU Romberg Tiburon Center.

Time period: Mid-2006 - 2010

Resources required:

Cost: No funds required for 2010.

PI(s): Jonathon Stillman and Wim Kimmerer (SFSU)

Contract needed / in place: Expired

Contract manager(s): Ted Sommer (DWR)

Term of contract: October 1, 2006 – September 30, 2008

Personnel: Jonathon Stillman, Adam Paganini, Nathan Miller, and additional student researchers (SFSU).

Equipment: Several items of equipment used for physiological assays, including a microplate spectrophotometer used in enzyme assays and feeding rate determination, have been purchased.

Deliverables and dates:

- Manuscript representing final report of this work was submitted to Marine Ecology Progress Series (April 2010).
- Revised manuscript will be sent to Aquatic Biology (September 2010). This manuscript, once published, can be summarized as an IEP Newsletter article, if desired.
- Research presented at numerous conferences, including the annual IEP Workshop, Estuarine Ecology Team (EET) meetings, California Estuarine Research Society Conference (CAERS), American Physiological Society Conference (APS), and Bay-Delta Science Conference, from spring 2009 to fall 2010.

Comments: This work explores a range of physiological responses of *Corbula* to environmental salinity and temperature, including thermal tolerance, salinity tolerance, cardio-respiratory and metabolic responses to salinity and temperature, energetics and feeding rates at different salinities, enzyme activities at different salinities, and osmoregulatory physiology. Our research led to a number of interesting findings, including the fact that *Corbula* always hyporegulate their osmotic content, but do so to a greater extent at high salinity, suggesting greater metabolic demands and food intake at high salinity may be related to osmoregulatory costs.

Field Survey of Microcystis Aeruginosa Bloom Biomass and Toxicity

IEP-related 2010-079

Point person: Rich Breuer (DWR)

Lead Agency: DWR-DES, DFG

Questions: Is *Microcystis* biomass or toxicity increasing over time in the Delta? Does *Microcystis* bloom biomass or microcystins toxicity occur in areas important to pelagic fish species in the Delta? What are the long term trends in *Microcystis* and potential controlling

factors? What are the origins of the *Microcystis* blooms? What environmental factors affect bloom development?

Description: In 2010, phytoplankton sample analysis and data analyses will be conducted to evaluate density, biomass and toxicity data collected during *Microcystis aeruginosa* blooms in the Delta between 2007 and 2008.

Time period: Summer and fall 2007-2011.

Resources required:

Cost: \$144,000 in 2010 from Delta Science grant.

PI(s): Peggy Lehman (DWR)

Contract needed / in place: In place.

Contract manager(s): Ted Sommer (DWR)

Term of contract: Through December 31, 2011.

Personnel: This work would be conducted by DWR personnel.

Equipment: None

Deliverables and dates:

- Semi-annual reports submitted by June and December 2009 and 2010.
- Oral progress reports to IEP project work teams by September 2009 and September 2010.
- Oral or poster presentations at the 2009 and 2010 IEP Workshop.
- Post study report to peer reviewed journal and/or published in the summer 2010 IEP Newsletter.
- Protocol for laboratory culture of threadfin shad (August 2010).

Comments: Toxicity analysis will be done by G. Boyer of the State University of New York, an expert on cyanobacterial toxicity. His group has extensive experience in the determination of cyanobacterial toxins and routinely analyzes samples for NOAA, Center for Disease Control (CDC, and other departments of health and conservation for several states. They also participated in the previous surveys. Future analyses may be possible at DFG's Water Pollution Control Laboratory.

Food-web Support for Delta Smelt and Estuarine Fishes in Suisun Bay and Upper Estuary IEP 2010-082

Point person: Anke Mueller-Solger (DSC)

Lead Agency: SFSU-RTC

Questions: Within the low salinity zone (LSZ) of the northern estuary:

How do benthic grazing, available solar irradiance, and the concentrations and composition of nitrogenous nutrients interact to influence the species composition and production of phytoplankton? How does bacterial production respond to changes in particulate and dissolved organic carbon (POC & DOC) delivered primarily through river flow? What is the role of the microbial food-web in supporting higher trophic levels?

To what extent is copepod production dependent on these alternative energetic pathways (phytoplankton and bacterial production)?

Description: This is a DSP-funded study focused on 2 related topics:

- Topic 1: The threatened delta smelt (*Hypomesus transpacificus*) is now the principal species of concern for management of freshwater flow and diversions in the Delta, and the principal target for restoration in the upper San Francisco Estuary. The abundance of this federally-listed threatened species has been low since the early 1980s and it has not recovered to the point where it can be considered for delisting; indeed, the 2004

abundance index was the lowest on record. Potential reasons for its low abundance are many, but evidence points to the direct and indirect effects of export pumping of freshwater in the south Delta, toxic substances, and low food supply as likely contributing factors. We believe that the feeding environment of delta smelt may be implicated in the continued low abundance of this species. Delta smelt feed for their entire lives on zooplankton, principally copepods, mainly in the brackish waters of the western Delta and Suisun Bay. As outlined in the submitted proposal, copepod abundance is depressed in this region.

- Topic 2: Previous work on the responses of the estuarine ecosystem to inter-annual variation in freshwater flow has demonstrated a decoupling between the abundance of lower trophic levels and that of fish and shrimp (Kimmerer 2002; Kimmerer 2004). This decoupling may imply that variability in food-web support is unimportant to variability of higher trophic levels, but there are some important pieces missing from the puzzle. Chief among these is the fact that the supply of labile organic matter from freshwater to the LSZ varies with freshwater flow and this flux has not been accounted for in analyses of the estuarine food-web.

The funded proposal includes efforts aimed at understanding and possibly improving the food-web supporting delta smelt and other estuarine species.

Time period: 2006-2010

Resources required:

Cost: \$162,000 per year from a DSP PSP-funded grant.

PI(s): Wim Kimmerer (SFSU)

Contract needed / in place: In place.

Contract manager(s): Shem Ayalew (DSP)

Term of contract: 3 years

Personnel: Richard Dugdale, Frances Wilkerson, Edward Carpenter, Alex Parker (SFSU); Risa Cohen (Georgia Southern University); Janet Thompson, Francis Parchaso (USGS); George McManus (University of Connecticut)

Equipment: The bulk of the laboratory work will be conducted at Romberg Tiburon Lab.

Deliverables and dates: There are 5 tasks within this element: phytoplankton (task 1), benthic grazing (task 2), bacteria (task 3), microbial foodweb (task 4), and copepods (task 5).

- Tasks 1, 2, 3, and 5 will all produce at least 1 peer-reviewed journal article, at least 1 presentation at the Delta Science Conference, and presentations at the Estuarine Ecology Team and other venues.
- Tasks 3 and 4 will produce a joint paper together.
- All tasks will produce a synthesis article. The target milestone for paper submission is December 2010.

Comments: This project was delayed by the bond freeze. The delay did more than push the project back and caused a serious disruption that has made it difficult to advance to the synthesis stage. Many presentations have been made as a result of this project. At least 5 manuscripts are now in preparation and data are ready for 2 more.

Power Plant Operations

IEP 2010-087

Point person: Ted Sommer (DWR)

Lead Agency: DWR

Questions: What are the characteristics of the cooling water diversions associated with the Contra Costa and Pittsburg power plants, and what effects might they have on pelagic fishes? Have there been recent increases in pelagic fish entrainment?

Description: This study was previously a component of the 2005 work element, “Analysis/summary of recent changes in Delta water operations”. Based on the initial data review, we believed that the issue warranted a focused study. The purpose of this element is to closely examine power plant operations to identify whether there were effects strong enough to contribute to the long-term and recent apparent step-change in pelagic fish abundances.

Time period: Ongoing through April 2011.

Resources required:

Cost: \$25,000 from DWR POD sources.

PI(s): Carol Raifsnider (Tenera Environmental)

Contract needed / in place: In place.

Contract manager(s): Steve Bauman (Mirant) manages the contract with Tenera.

Funding for Brian Schreier (DWR) is from DWR POD sources.

Term of contract: Current through 2011.

Personnel: Brian Schreier (DWR)

Equipment: None

Deliverables and dates:

- Monthly progress reports detailing entrainment and impingement results, submitted to IEP agency biologists.
- Annual summary report submitted 4 months after laboratory sampling has been completed.

Comments: Entrainment and/or impingement surveys were coordinated to occur during the same time as the FMWT, SKT, TNS and the 20-mm Survey, when the surveys were conducted in the vicinity of the power plants. Monitoring was conducted at a minimum of once a month from November 2007 through October 2009. Two annual reports summarizing results of entrainment and impingement monitoring were submitted. The study plan was amended (Amendment 2) and approved in March 2010. Amendment 2 Entrainment and Impingement Monitoring Program is conducted only during times when the power plants generate electricity and is no longer conducted parallel to IEP sampling.

SAV Abundance and Distribution

IEP 2010-102

Point person: Ted Sommer (DWR)

Lead Agency: USGS and UCD

Questions: Has submerged aquatic vegetation (SAV) increased in the Delta? Has SAV altered the habitat to effect fish populations? Has SAV increased retention of suspended solids to create a less turbid environment, which is less hospitable to delta smelt?

Description: Using hyperspectral imagery, this project will provide annual acreage calculations of SAV and quantify regional distribution trends in the Delta for 2003-2006. Three tasks have been identified: (1) classify SAV in the Delta using Hy Map hyperspectral imagery, (2) create SAV distribution maps in concert with SAV acreage calculations to quantify distribution trends in the Delta, and (3) investigate methodology that will detect and monitor turbidity trends in the Delta using surface and aircraft hyperspectral remote sensing instruments and satellite remote sensing data.

Time period: 2006-2010

Resources required:

Cost: This program has billed remaining funds.

PI(s): Susan Ustin (UCD)

Contract needed / in place: In place.

Contract manager(s): Ted Sommer (DWR)

Term of contract: Through June 30, 2010.

Personnel: Erin Hestir (UCD)

Equipment: Equipment provided by UCD.

Deliverables and dates:

- Oral presentation at CALFED Science Conference, October 22, 2008 (tasks 1 and 2).
- Oral presentation at IEP Workshop, February 28, 2009 (tasks 1, 2 and 3).
- Progress report to DWR, February 28, 2009 (tasks 1, 2 and 3).
- IEP Newsletter article and at least 1 written manuscript for peer-reviewed journal, June 30, 2009 (tasks 1, 2 and 3).
- Poster presentation at State of the Estuary, October 2009 (task 3).
- Oral presentation at IEP Workshop, February 28, 2009 (task 3).
- IEP Newsletter article summarizing results of each task, June 30, 2008.
- Manuscript detailing the method used to create the spatially explicit geographic information system (GIS) layer of SAV, June 30, 2010.

Comments: None

Fish Facility History

IEP 2010-107

Point person: Marty Gingras (DFG)

Lead Agencies: DFG, USBR

Questions: What changes have occurred at the state and federal fish facilities that would change the reported number of salvaged fish?

Description:

This project will identify changes that have occurred at the state and federal fish facilities from 1956 to 2010 that may have impacted the reported number of salvaged fish. However, this investigation will not report on the survival of fish once counted (debris loads in holding tanks and impacts on released fish).

Time period: Through December 31, 2010.

Resources required:

Cost: \$50,000 from USBR Operations (non-POD).

PI(s): Jerry Morinaka (DFG) and Brent Bridges (USBR)

Contract needed / in place: N/A

Contract manager(s): Erwin Van Nieuwenhuysse and Ron Silva (USBR)

Term of contract: N/A

Personnel: Above named investigators.

Equipment: No equipment is required for this analysis.

Deliverables and dates:

- IEP technical report describing changes that have occurred at the state facilities (December 2010).

- The federal portion of this element will begin in August 2010 and the target deadline is winter 2011.

Comments: None

Delta Smelt Culture Facility

IEP 2010-108(a)

Point person: Rich Breuer (DWR)

Lead Agency: UCD

Questions: Reliable supplies of all life stages of delta smelt are valuable to management and scientific communities for a number of reasons. Cultured delta smelt provide specimens with known rearing history (required for toxicological experiments), aids research and design of fish screen efficiency and pre-screen losses, allows investigations into basic biology with application to wild populations, and enables the development of a delta smelt refugial population.

Description: This program will collect sub-adult broodfish via purse seine from the wild and spawn, and rear all life stages of delta smelt in the following year in accordance to the Delta Smelt Culture Manual (Baskerville-Bridges et al., 2005). Delta smelt will be housed and reared at the newly expanded Fish Conservation and Culture Laboratory (FCCL).

Time period: Ongoing

Resources required:

Cost: \$48,000 available from DWR contract and \$881,000 from USBR contract.

PI(s): Drs. Raul Piedrahita, Joan Lindberg and Brad Baskerville-Bridges (UCD)

Contract needed / in place: In place.

Contract manager(s): Rich Breuer (DWR), Erwin Van Nieuwenhuysse (USBR), Joan Lindberg (UCD)

Term of contract: July 1, 2007 to June 30, 2010 (DWR); January 1, 2010 to December 31, 2014 (USBR).

Personnel: UCD personnel.

Equipment: Equipment to rewire the collection, handling, transport and release (CHTR) building, and the purchase of additional chillers will be needed.

Deliverables and dates:

- 5,000 adult delta smelt (>50 mm).
- 10,000 juvenile delta smelt (20-50 mm).
- Annual Production Report (December 2010).

Comments: In 2010, SKT personnel will experiment with artificial gamete stripping of wild-caught delta smelt and cold storage of gametes. If cold-stored gametes prove viable, this will provide a means to introduce additional genetic diversity into cultured delta smelt. These same fish will be sampled for delta smelt genetic testing (2010-135) and be preserved for otolith (2010-060) and histopathology (2010-061).

Quantifying Effects of Naturally Occurring Physical Stimuli on Delta Smelt Behavior

IEP 2010-108(b)

Point Person: Pete Hrodey (USFWS)

Lead Agency: UCD

Questions: Do juvenile and adult delta smelt discriminate levels (and types) of turbid and saline environments, make behavioral choices based on this information? Does delta smelt show differences in feeding, survival and swimming behavior at several levels and/or types of

turbidity? Does salinity or temperature affect feeding, survival and swimming behavior under optimal turbidity conditions?

Descriptions: The study includes the following tasks:

- **Task 1: Effect of salinity and turbidity on juvenile and adult delta smelt behaviors:** In the first task, we aim to test whether juvenile and adult delta smelt discriminate levels (and types) of turbid and saline environments and make behavioral choices based on this information. We will develop test equipment (similar to our successful raceway design used in testing larval delta smelt) to try to document residence time (or volitional movements) under conditions where the animals can choose between turbid and clear water, and between saline and fresh water. Identifying volitional responses of the late juvenile and sub-adult life stages to salinity and turbidity and to various types of turbidity will yield practical information regarding the smelt's responses to changes in environmental elements and inform management decisions. We will test delta smelt juveniles and adults in "Y-maze" raceways, with physical variables randomized between arms of the mazes. An additional test apparatus may be explored where fish are tested in a large oval or circular tank in which water velocity could be manipulated to simulate tidal movements (flood, slack, ebb).
- **Task 2: Effects of turbidity, salinity, and temperature on feeding, survival, and behavior of delta smelt:** Building on experiments performed at the FCCL (Baskerville-Bridges et al., 2004), delta smelt at juvenile and early adult stages will be exposed to different levels (NTU) and types of turbidity (phytoplankton, humic acids, suspended sediments/silt) for 1-7 days in a flow-through system. Swimming behavior, feeding and survival will be monitored to assess optimal conditions. Experiments will be performed at different water temperatures or salinities to determine potential interactive effects. Current test protocols will follow those developed at UCD-Aquatic Toxicology Laboratory (ATL) and described in detail by Werner et al. (Final Report to IEP-POD, 2009) for testing with delta smelt larvae at different stages of development. A flow-through system will be used for testing. Upon arrival at UCD-ATL, the transport containers with fish will be placed into a temperature-regulated water bath maintained at 16° C. Glass beakers are used to gently transfer fish from transport containers to replicate exposure tanks. Six fish will be placed into each of the test tanks containing 7 L of filtered hatchery water adjusted to (1) a range of turbidity (0-25 NTU) using phytoplankton, humic acids or silt, and (2) optimal turbidity determined in experiment 1 at different salinities/conductivities (150-200 uS/cm, 1.5-2 ppt, 5-8 ppt). Fish will be acclimated to different salinities/conductivities for 2 weeks before the test, and hatchery water and electric conductivity (EC) adjusted hatchery water will be used as control water. During acclimation and testing, fish will be fed 3 times a day with *Artemia*. Water quality parameters (EC, pH, temperature, dissolve oxygen (DO), turbidity and ammonia concentration) will be measured, and dead fish will be recorded and removed daily. At test termination, mortality will be recorded. Half of the fish will be then be fed and gut contents examined subsequently. The other half will be subject to swimming trials.

Time Period: June 2010 – June 2011

Resources required:

For Task 1:

1. Facilities at the FCCL of UCD, located near Byron, CA.
2. Cultured delta smelt: 2,000-3,000 fish of each life stage, juveniles and adults as follows:

20 delta smelt per behavior test x 2 tests x 30 replicate experimental runs = 1,200 juveniles and 1,200 adults. Estimate double the number of fish for developing techniques, mortalities, and unforeseen complications.

For Task 2:

1. Facilities at the UCD - ATL.
2. Cultured delta smelt: 1,000 juveniles-subadults (140-150 dph); 2 tests, 12 treatments with 4 replicates per test, 6 fish/replicate; 400 fish for developing techniques and unforeseen complications which necessitate repeat testing.

Cost: \$100,000

PI(s): Joan Lindberg and Inge Werner (UCD)

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuyse (USBR) and Joan Lindberg (UCD)

Term of contract: January 2010 – December 2014

Personnel: Above named investigators and other UCD staff.

Equipment: No new equipment at this time.

Deliverables and dates:

Task 1

- Results will be presented at the Delta Science Conference or annual IEP Workshop.
- Report at termination of project.

Task 2

- Results will be presented at the Delta Science Conference or annual IEP Workshop.
- Report at termination of project.

Comments: This study will complement earlier studies conducted at the FCCL on larval movement and feeding behavior in response to turbidity and incident light. In addition, it follows up on observations made during larval testing at UCD-ATL, indicating that different types of turbidity may influence feeding and swimming activity of delta smelt. If the juvenile and adult smelt show increased residence times and better feeding behavior in turbid and slightly-saline environments, then this would have management implications for habitat protection and hydraulic manipulations of the low-salinity and higher turbidity zone in the Delta.

Striped Bass Bioenergetics Evaluation

IEP 2010-115

Point person: Ted Sommer (DWR)

Lead Agencies: DWR and DFG

Questions: What are the trends in estimated population consumption demand of age-1 and older striped bass? Has age-1 and older striped bass consumption demand decreased more slowly than prey relative abundance/relative biomass?

Description: This element will couple bioenergetics analyses to data provided by element 2010-116 (adult striped bass population dynamics) to estimate the long and short-term (i.e., POD years) trends in consumption demand of striped bass.

Time period: 2010; assuming the population demographic data are available in early 2010.

Resources required:

Cost: \$30,000 in 2010 from POD sources.

PI(s): Marty Gingras (DFG), Gina Bengino (DWR)

Contract needed / in place: Not needed, work will be covered by agency staff.

Contract manager(s): N/A

Term of contract: N/A

Personnel: Jason DuBois (DFG)

Equipment: None required – this is a data mining/data analyses effort.

Deliverables and dates: Draft manuscript for publication (December 2010).

Comments: None

Long-term Sources and Early Warning Signals in Turbidity Monitoring Data

IEP 2010-126

Point person: Ted Sommer (DWR)

Lead Agency: USGS

Questions: (1) Do smelt move towards the water diversions because they are following specific turbidity pulses, or because their habitat has shifted towards the pumps? (2) Where are the sources of turbidity for water exported by the SWP and Central Valley Project (CVP)?

Description: This study will determine the origin, movement, and extent of turbidity pulses that affect delta smelt behavior and salvage at the pumps. Recent analyses by DWR and Metropolitan Water District show that delta smelt salvage counts increase at the CVP and SWP during periods (i.e., days) immediately following the first significant storm event in the basin. Though the mechanism is unclear, water turbidity is a good measure of the storm event and the data indicate that salvage typically begins when turbidities increase over 12 NTU. As a result of a recent court ruling, this finding was incorporated into water project operations as a tool to help reduce fish losses. However, the sources of turbidity are not well understood. A more comprehensive source of long-term turbidity data could help elucidate the question as to whether fish are following a turbidity gradient to the salvage facilities or whether they are migrating in response to a flow pulse. Data sources will focus on USGS continuous suspended-sediment concentration time series at Rio Vista, Jersey Point, Threemile Slough, and Stockton from 1998-2005.

Time period: 12 months.

Resources required:

Cost: Funding for this element was obligated with previous year funds.

PI(s): David Schoelhamer and Scott Wright (USGS)

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Darcy Austin (USGS).

Term of contract: 12 months, expiring 9/30/2011.

Personnel: Tara Morgan (USGS) in collaboration with Lenny Grimaldo (USBR).

Equipment: None

Deliverables and dates:

- Process 2003-2005 suspended sediment (SSC) data.
- Develop relations between turbidity and SSC from the Bay Delta and Tributaries (BDAT) database for specific sites and seasons relevant for delta smelt.
- Develop statistical models of turbidity at USGS sites for 1993-1997.
- Define origin, movement, and extent of key turbidity pulses identified by Lenny Grimaldo (USBR) from 1993-2005.
- Collaborate with Lenny Grimaldo (USBR) on peer reviewed manuscript.

Comments: Processing of 2003-2005 SSC data and development of turbidity and SSC relations is proceeding.

Contaminants and Biomarkers Work

IEP 2010-127

Point person: Rich Breuer (DWR)

Lead Agency: UCD

Questions: Is water in ecologically sensitive areas of the Delta toxic to delta smelt and other pelagic fish and their prey? What are the causes and sources of water column toxicity in areas of the Delta and how do they affect fish species of concern? How sensitive are Delta species to contaminants in comparison to surrogate species commonly used in toxicity testing? Is it meaningful to use surrogate species for toxicity monitoring in the Delta? Are contaminants associated with wastewater treatment effluents affecting fish species of concern? Is there a relationship between toxicity results and other POD study components, such as histopathological examination of fish and *Microcystis* blooms?

Description: The overall goal of this study is to assess the potential for contaminated water to contribute to the observed declines of pelagic species in the Delta. The study is designed to build from the results of the 2006 and 2007 Delta-wide monitoring project, which investigated toxicity of Delta water samples to invertebrates and early life stages of fish species of concern.

The 2009 – 2010 study will intensify toxicity testing in the Cache Slough, lower Sacramento, Suisun Marsh and Suisun Bay areas. Like the 2006 and 2007 study years, if toxicity is detected, toxicity identification evaluations (TIEs) and chemical analysis will be used to identify toxicant(s). *In situ* tests with delta smelt, fathead minnow or inland silversides, and *Hyaella azteca* will be conducted at Hood on the Sacramento River, and Rough and Ready Island on the San Joaquin River. This study will generate sensitivity data (in the form of 96-h LC50, EC50, no observed effect level (NOEC), and lowest observed effect level (LOEC)) and compare sensitivity of Delta species with that of standard toxicity test species for *Pseudodiaptomus forbesi*, *Eurytemora affinis*, *Ceriodaphnia dubia*, *Hyaella azteca*, delta smelt, and fathead minnow or inland silverside for select chemicals. Lastly, molecular biomarkers developed for striped bass in 2006-2007 will be used to detect and quantify stress responses in field-collected specimens from 2005-2009 to detect sublethal toxic effects and help identify the causative chemical(s) or other stressors. Biomarker development for delta smelt will continue with the immediate aim of selecting appropriate biomarkers for use in field and *in situ* studies, as well as in laboratory studies to determine cause and effect.

Time period: This study will be completed in 2010.

Resources required:

Cost: \$453,000 from POD sources in 2010.

PI(s): Drs. Inge Werner and Swee Teh (UCD)

Contract needed / in place: In place.

Contract manager(s): Ted Sommer (DWR)

Term of contract: Through June 30, 2010.

Personnel: UD Davis staff, Kevin Reece (DWR)

Equipment: Equipment will be provided by UCD.

Deliverables and dates:

- Semi-annual progress reports will be submitted to the DWR contract manager and will include the number of samples processed, number of samples analyzed, results and a timeline for the completion of the analyses.
- Oral progress reports will be given to the IEP project work teams (September 2009 and 2010).
- Annual IEP workshop presentation (February 2009 and May 2010).
- A post-field progress report that describes the study and outcome to a peer-reviewed journal and/or published in the IEP Newsletter (summer 2009).

Comments: This project is dependent on the ability to obtain 5,000 – 6,000 delta smelt larvae aged 30 – 90 days each year for the toxicity testing and sensitivity studies, and 1,000 40 – 45 day old larvae for tests with ammonia and wastewater treatment effluent from Sacramento River water.

Feasibility of Using Towed Imaging Systems

IEP 2010-130

Point person: Ted Sommer (DWR)

Lead Agency: USBR

Questions: Are towed video imaging systems a feasible technique for measuring the abundance and distribution of pelagic fishes in the Delta and San Francisco Estuary? If so, what species, life stages, and regions would be most suitable for this technique?

Description: Trawls presently form the foundation of IEP fish monitoring. Although these gear types have proved exceptionally useful, they are much less effective with patchy or rare species such as delta smelt. Moreover, the recent decline in delta smelt population has led to concern over lethal “take” by trawling methods. Recent progress in towed video imaging systems may provide a supplemental method that could be used to examine pelagic fish distribution and abundance. The potential use of underwater video cameras retrofitted to trawl surveys as a new non-lethal tool to measure the abundance and distribution of pelagic fishes has gained recent interest. An underwater housing has been designed, fabricated, and tested to attach to the cod-end of a towed-net and perform under strict buoyancy, pitch, yaw, and roll specifications. The device has been tested in flows up to 0.9 m/sec (3 ft/sec) in turbid water. The aluminum hull contains a video imaging system, using a black and white, right-angled Gig-E high-speed camera to capture image data and send to boat-side computer system over Ethernet. The electrical system includes lighting, light driving, sensor sampling, cameras, camera power, lightning protection, and smaller subsystems. Sensors have been included that interface with the computer and provide information pertaining to the global positioning system (GPS) location, pressure for depth (inner and outer), humidity, temperature (air inside hull), water flow rate (from pressure differential), and tilt angle (and rate) from accelerometers.

The computer system processes images in real-time using a combination of NVIDIA compute unified device architecture (CUDA) parallel computing platform, LabVIEW, and the image processing toolkit from National Instruments. Vision system improvements include the programming of algorithms software to determine fish profile outline and estimate approximate pose, develop and train software to recognize objects as fish from a library database and statistical models, and classify objects automatically, as fish or non-fish and give probability of each object of being a delta smelt. An adaptive software reduction of turbidity through known fish position and gamma correction is an ongoing process and will be further designed.

The software system is composed of 2 parts: (1) a user interface that includes data visualization, sensor acquisition, alarming conditions, camera inputs, and data recording, and (2) a vision algorithm that is actively detecting fish-sized objects using CUDA (CUDA is the computing engine in NVIDIA graphics processing units). Other feature trackers and algorithms have been implemented to detect and describe local features that will help in tracking objects and deciding whether or not it is a fish, debris, or bubbles. A more rigorous testing phase will occur in the fall, winter, and spring 2010–2011, including more laboratory and field analysis of the underwater camera fish recognition system.

Time period: Through 2011 and possibly beyond, depending on how long it takes to build an object image library.

Resources required:

Cost: \$201,000 from POD sources.

PI(s): Don Portz (USBR) and Darren Odom (SureWorks, LLC.)

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Ted Sommer (DWR)

Term of contract: Annual

Personnel: USBR, DFG and DWR staff

Equipment: Aluminum underwater submersible, Dalsa high-speed camera, laptop computer, LED lighting, camera cabling, flumes, and use of trawls.

Deliverables and dates:

- Progress report (September 2010).
- IEP Newsletter article draft (December 2010).

Comments: The software system that the Bureau of Reclamation and SureWorks, LLC. is designing requires more work on the vision algorithm, some flume testing, and field trials for the fall, winter, and spring 2010–2011.

Use of Acoustics to Estimate Trawl Openings

IEP 2010-131

Point person: Randy Baxter (DFG)

Lead Agency: DFG

Questions: What are the mouth dimensions of a standard midwater trawl? Do these dimensions vary based on differences in trawling depth or current direction? Are trawl mouth dimensions affected by variation in rigging among the vessel being used (i.e., does the block height and width among vessels affect mouth dimensions)? The information should improve the accuracy of abundance indices and will be directly applicable for the development of mean-density expansion estimators (see 2010-043 above) used to estimate the population size of delta smelt and other fishes susceptible to the trawling. Results may suggest appropriate alternative rigging configurations to improve consistent deployment of trawls.

Description: This study will employ a commercially-available transmitter/transponder/computer system to calculate dimensions of a midwater trawl mouth while the net is being towed during special deployments to emulate a variety of different depth, current, vessel speed, and vessel configuration circumstances. The midwater trawl will be towed by all currently operational DFG vessels used for the FMWT and SFBS. If time permits, the gear will be used to measure mouth dimensions of the SKT, SFBS otter trawl, and the USFWS Chipps Island midwater and Kodiak trawls.

Time period: January – December 2010

Resources required:

Cost: \$13,000 from POD sources.

PI(s): Randy Baxter, Steve Slater, Dave Contreras and Jennifer Messineo (DFG)

Contract needed / in place: Not needed.

Contract manager(s): N/A

Term of contract: N/A

Personnel: 2 -3 biologists, 1 mate, and 1 scientific aide would be temporarily directed.

Equipment: Software and hardware have been purchased and successfully used to complete part of the list of net measurements. The equipment malfunctioned and has been returned to the vendor for repair, which must be successful to complete this study.

Deliverables and dates:

- Not less than 5 memo-reports and companion Excel summary files, each submitted within a month following any deployment of the equipment.
- IEP newsletter article describing initial field work (fall 2010).
- Draft a manuscript describing inter-vessel effects on trawl mouth dimensions including a discussion on how rigging affected trawl dimensions (December 2011).

Comments: Equipment malfunctions may limit work possible in 2010.

The Effects of the Cache Slough Complex on North Delta Pelagic Habitat: Regional Transport of Turbidity, Phytoplankton, and Zooplankton

IEP 2010-132

Point person: Ted Sommer (DWR)

Lead Agencies: DWR and USGS

Questions: This study will examine the hydrodynamic “footprint” of Liberty Island in the north Delta, one of the key habitats of delta smelt. Two of the alternative hypotheses to be tested include:

Alternative 1: Liberty Island has a dominant effect on the hydrodynamics of Cache Slough Complex channels and perhaps other parts of the north Delta. Hence, we expect to see pulses of turbidity, chlorophyll *a*, and zooplankton coincident with tidal export from Liberty Island. Specifically, we should see these pulses on an ebb tide in lower Cache Slough (and perhaps Rio Vista) and on a flood tide in the Deep Water Ship Channel.

Alternative 2: Transport of particulates in the Cache Slough complex is determined largely by localized production (phytoplankton, zooplankton) or resuspension (sediment). For turbidity, the hypothesis is that sediments are deposited in Cache Slough Complex channels during flood events. Hence, high turbidity during other times of the year would occur in the Deep Water Ship Channel and lower Cache Slough during (1) peak ebb and flood tides; (2) high wind periods; or (3) both conditions. Chlorophyll *a* in lower Cache Slough and the Deep Water Ship Channel would show diel and seasonal variation, but no tidal effects. Zooplankton would show seasonal effects and perhaps diel effects.

Description: The POD has created substantial interest in characterizing the habitat of pelagic fishes. This includes analyses of short term (e.g., tidal) and long-term (e.g., seasons, years) changes. Recent studies have revealed several important observations: (1) the Cache Slough Complex represents key habitat for delta smelt; (2) turbidity is an especially important component of habitat for the Delta pelagic fishes; (3) much of the sediment and organic carbon flux in the estuary passes through the Cache Slough Complex; and (4) the Cache Slough

Complex appears to be an important “food bank” for the Delta. Hence, understanding the patterns of hydrodynamics, turbidity, chlorophyll *a*, and zooplankton is important to describe the habitat of pelagic fishes including delta smelt. The proposed study will examine the hydrodynamic “footprint” of Liberty Island, the major body of water in the Cache Slough Complex. Flux of phytoplankton out of Liberty Island will be studied by Peggy Lehman as part of the Breach III study, providing a good opportunity to examine the fate of the exported material. We suspect that transport of biological and physical constituents from Liberty Island has a dominant effect on the channels of the Cache Slough Complex and perhaps a large area of the north Delta. The study approach includes both continuous monitoring and 24-hour flux studies. Ongoing work in 2010 will include 2 additional 24-hour sampling events to capture “wet” winter conditions, and more focused continuous monitoring in the upper Cache Slough and Liberty Island areas.

Continuous Monitoring:

Wind: Rio Vista, Hastings (existing stations)

Flow: Cache Slough at Ryer Island, Upper Cache Slough, Deep Water Ship Channel, Miner Slough, and Lindsay Slough.

Turbidity: Cache Slough at Ryer Island, Upper Cache Slough, Ulatis, Deep Water Ship Channel, Miner Slough, and Lindsay Slough.

24-Hour Intensive Studies (once every 3 months):

Phytoplankton and zooplankton: Liberty Island, Cache Slough, Deep Water Ship Channel.

Chlorophyll *a*: Liberty Island, Cache Slough at Ryer Island, Upper Cache Slough, Deep Water Ship Channel, Shag Slough, Lindsay Slough.

Time period: 2008 to 2010

Resources required:

Cost: \$334,170

DWR biological sampling \$89,270

DWR hydrodynamic sampling \$27,000

USGS turbidity sampling \$217,900

PI(s): Gina Benigno, Shawn Mayr and Ted Sommer (DWR); Tara Morgan and Dave Schoellhamer (USGS).

Contract needed / in place: In process.

Contract manager(s): Rich Breuer (DWR), Erwin Van Nieuwenhuysse (USBR) and Roger Fujii (USGS)

Term of contract: To be determined.

Personnel: Identified PIs plus additional IEP staff during the 24 hour flux studies.

Equipment: An Acoustic Doppler Current Profiler (ADCP) for some of the locations, turbidity sensors for all the sites, and the use of 3 small boats during the four 24-hour flux studies.

Deliverables and dates:

- Delta Science Conference presentations (October 2010).
- Peer reviewed manuscript for publication (December 2010).

Comments: None

Impacts of Largemouth Bass on the Delta Ecosystem

IEP 2010-133

Point person: Ted Sommer (DWR)

Lead Agency: UCD

Questions: How much time do centrarchids spend foraging in pelagic habitat? What is the relationship between increasing biomass of *Egeria* and centrarchids density, as well as other species in the littoral zone? What are “ballpark” estimates of centrarchid population sizes and prey requirements?

Description: Although “top-down” effects are a key part of the POD conceptual model, predation from inshore piscivores represents a relatively poorly understood source of mortality. There is good evidence that centrarchid populations have thrived as a result of the expansion of *Egeria* beds; however, it is unclear whether this may have contributed to the POD. Specifically, we need estimates of inshore predator abundance and information about their effects on pelagic habitat. Four tasks comprise this element: (1) field surveys of largemouth bass and other fish, (2) feasibility study of acoustic tracking methods for largemouth bass in the Delta, (3) mesocosm studies on largemouth bass diets, and (4) statistical models of largemouth bass abundance and impacts.

Time period: 2009-2010

Resources required:

Cost: \$239,000 in 2010 from POD sources.

PI(s): Drs. Andy Sih, Peter Moyle and Peter Klimley (UCD)

Contract needed / in place: This work is included in the DWR-UCD umbrella contract that expired June 30, 2010 and will be continued through a USBR-UCD contract.

Contract manager(s): Ted Sommer (DWR)

Term of contract: Through June 30, 2010 (DWR).

Personnel: Louise Conrad (DWR) and Anna Stephensen (UCD)

Equipment: This work requires the use of an electrofishing boat, likely borrowed from DFG.

Deliverables and dates:

- Progress report for task 1 to IEP MT (fall 2010).
- IEP Newsletter submissions for task 1 (fall 2010).
- Oral presentation at Delta Science Conference for tasks 1, 2 and 4 (October 2010).
- Poster presentation for task 1 at State of the Estuary Conference (October 2009).
- Oral presentation for task 3 at the IEP Annual Workshop (May 2010).
- 1 or more peer-reviewed journal articles for task 1, 2, 3 and 4 (fall 2010).

Comments: None

Delta Smelt Genetics

IEP 2010-135

Point person: Randy Baxter (DFG)

Lead Agency: USFWS

Questions: What is the current and historic population structure of the delta smelt population? What is the mating strategy of delta smelt? What is the extent of hybridization between delta smelt and wakasagi smelt or longfin smelt? What type of genetic breeding plan would be effective to maintain genetic variability in a captive refugial population? Is genetic variability being maintained in the captive refugial population?

Description: The delta smelt abundance has declined recently to record low levels (Sommer et al., 2007), prompting petitions to “uplist” it from the current threatened status to endangered under both the California and Federal Endangered Species Acts (The Bay Institute et al., 2006 and 2007). Low abundance levels also lead to unprecedented restrictions in biological sampling in 2007, promulgated under IEP delta smelt “take” and DFG collection permit authority. Other efforts to manage delta smelt include investigations into reproductive biology and culture conducted at the UCD FCCL, located near the Skinner Fish Facility in Byron. The facility has been acquiring wild brood stock annually to produce young for experimental purposes, but these collections were restricted in 2007. As a result of the declining abundance and reduced access to wild brood stock, a refugial population has been developed. However, to effectively establish and maintain a refugial population, the genetic structure and dynamics of the population must first be determined. This study proposes to increase understanding of wild delta smelt population structure, hybridization, population dynamics, and spawning strategies through several years of study. Based on these findings, a breeding plan will be developed that will maintain “natural” genetic variation and population structure in the captive population. In addition, wild population genetic studies will help guide conservation management of delta smelt and permit genetic monitoring of the wild population.

Four tasks comprise this element:

- **Task 1:** Microsatellite marker development (begun in July 2007 and completed in April 2008).
 - Develop and optimize delta smelt specific primers to characterize microsatellite loci.
 - Test cross-species microsatellites (Saint-Laurent et al., 2003; McLean and Taylor, 2001; Kaukinen et al., 2004).
- **Task 2:** Determine the population structure of delta smelt throughout the Delta using microsatellite markers. Genotype all individuals by polymerase chain reaction (PCR) amplification using highly polymorphic loci lacking null alleles from already collected samples in collections held by R. Baxter and W. Bennett in addition to any wild individuals caught during project. Estimate allele frequencies, observed (H_o) and expected (H_e) heterozygosities, and inbreeding coefficients (F_{IS}) for all populations.
 - Determine the existence of genetically distinct populations of delta smelt in the Delta.
- **Task 3:** Assess the extent of hybridization between delta smelt and longfin (*Spirinchus thaleichthys*) or wakasagi smelt (*H. nipponensis*) to evaluate the role of hybridization in population abundance of delta smelt. Genotype longfin and wakasagi smelt by PCR amplification using polymorphic loci used for delta smelt.
 - Compare delta smelt genotypes to detect levels of hybridization and any introgression.
 - Evaluate level and percentage of hybridization with delta smelt.
 - Examine mitochondrial DNA of hybrids to determine if hybridization is species-directional.
 - Assess degree of downstream movement of wakasagi smelt from their original introduction in reservoirs.
- **Task 4:** Determine spawning strategies using breeding experiments and microsatellite markers to understand delta smelt population dynamics.
 - Conduct breeding experiments of delta smelt in a controlled environment.

- Examine progeny from tanks containing multiple individuals with different ratios of males to females.
- Use highly polymorphic microsatellite markers and computer algorithms to assign parentage to progeny.
- Assess timing and frequency of female egg release and male contribution to progeny to determine mating system (polygamous, polyandrous, or both) and spawning strategies.
- Determine effective population size from clarification of spawning strategies.
- **Task 5:** Develop a breeding plan to maintain natural genetic variation and population structure in captive populations using information obtained on population structure, dynamics, and spawning strategies of delta smelt.
 - Characterize the wild founding fish population structure and implications for management of potential breeding population.
 - Design breeding plan using information obtained from Tasks 2 and 4 to maintain natural variation in refugial population.
 - Determine number of refugial populations according to wild delta smelt population structure.
 - Assess natural breeding sex ratios and timing to facilitate desired crosses.
 - Develop a monitoring program to assess maintenance of genetic variation in wild and captive populations.

Time period: 2007 - 2012, with possible extensions and additional work funded through additional sources (see comments for contract extension).

Resources required:

Cost: \$53,000 from non-POD sources for 2010. This element is underfunded and projects needing an additional \$250,000 to complete all tasks by 2012.

PI(s): Bernie May and Katie Fisch (UCD)

Contract needed / in place: In place.

Contract manager(s): Paul Cadrett (USFWS) and Bernie May (UCD)

Term of contract: September 2007– September 2012

Personnel:

Equipment: Work for this element will take place at the Genomic Variation Laboratory (GVL) at UCD. Additional genetic work may be considered between the GVL, W. Bennett and S. Cohen, or between GVL and the FCCL.

Deliverables and dates:

General:

- Progress reports to agency funders as required in contracts.
- Year end final contract reports.
- Refereed publications.
- IEP Annual Workshop presentation (May 2010).
- Delta Science Conference presentation (October 2010).
- Scientific meeting presentations (e.g., Evolution and/or American Fisheries Society (AFS) annual and regional meetings).

Task 1

- Microsatellite markers for use in subsequent genetic analyses.
- Peer reviewed journal article (Published in Molecular Ecology Resources Issue 9, 2009).

Task 2

- Population structure.
- Peer reviewed journal article (2010).

Task 3

- Hybridization
- IEP Newsletter or peer reviewed journal article (2010).

Task 4

- Spawning strategy of delta smelt.
- Peer reviewed journal article (2011).

Task 5

- Genetic management plan and monitoring.
- Peer reviewed journal article (2012).

Work on this element will be completed by September 2012 if fully funded.

Comments: A new contract is being developed with USBR to fund genetic component of yearly maintenance of captive delta smelt refugial population through 2015.

Feeding and Growth of Delta Smelt

IEP 2010-136

Point person: Ted Sommer (DWR)

Lead Agency: SFSU

Questions: Year 1 questions: Do larval and juvenile striped bass and delta smelt exhibit selection for different zooplankton species? How does prey selection differ among larvae and juveniles? How do light and turbidity influence prey selection? What are the underlying mechanisms that determine prey selection?

Year 2 questions: What are the growth rates of larval delta smelt that are fed different prey species? What are the assimilation efficiencies of larval delta smelt feeding on different prey species? How do growth rates and assimilation efficiencies of larval delta smelt differ among diets comprised of different prey species?

Description: Video-graphic techniques will be used to record observations of predator-prey interactions and specific patterns of prey selection will be used to develop quantitative models of prey selection. Growth rates of larval delta smelt fed field-collected zooplankton will be measured in laboratory experiments. Length and weight measurements of larval fish, combined with biomass estimates of copepods will allow for calculations of assimilation and growth efficiency of larval fish feeding on different copepods. Data on respiration, ingestion, growth and excretion will be used to create an energy budget for larval delta smelt, allowing for the possibility of more accurate models of population dynamic.

Time period: 2009 - 2010

Resources required:

Cost: \$17,000 in 2010 from DWR POD.

PI(s): Lindsay Sullivan (SFSU)

Contract needed / in place: Requires modification of existing SFSU contract.

Contract manager(s): Ted Sommer (DWR) and Wim Kimmerer (SFSU).

Term of contract: Requires modification of existing contract with SFSU.

Personnel: Wim Kimmerer (SFSU)

Equipment: No new equipment is needed at this time.

Deliverables and dates:

- Presentation of results to date at Delta Science Conference (October 2010).

- Progress report to IEP MT (winter 2010).
- Scientific paper prepared for peer-reviewed journal (winter 2009 and 2010).
- Presentation of results at the State of the Estuary Conference (fall 2009).

Comments: This study is dependent on obtaining 5,000 newly hatched larvae for use in laboratory experiments. Prey selection and growth experiments have been completed. The data are currently being analyzed and publications are in preparation on both. Preliminary results were reported at the 2009 IEP Food Webs and Invasive Species Workshop, 2009 State of the Estuary Conference and 2009 Coastal and Estuarine Research Federation Conference. An abstract has been submitted to the 2010 Delta Science Conference.

Population Genetics and Otolith Geochemistry of Longfin smelt

IEP 2010-137

Point person: Ted Sommer (DWR)

Lead Agency: UCD

Questions: What is the genetic population structure of longfin smelt? Is there evidence of a recent change in life-history variability based on otolith chemistry and growth? Have there been changes in the demographics of longfin smelt?

Description: Longfin smelt recently has been proposed for listing as an endangered species. Although there is substantial concern about the status of this pelagic fish, there has been relatively little research on this species compared to delta smelt or striped bass. Hence, the basic population structure, demographics, and life history variability are unknown. The present study seeks to address some of these data gaps using 3 major tasks:

- **Task 1-Compare life-history variability from pre-POD to POD era (PI: Hobbs):** Using fish collected in the 1999-2001 FMWT Survey and the 2003 Bay Study, as well as the more recent collections from these sampling programs, the life-history of these longfin smelt samples will be reconstructed through examination of the strontium isotope $^{87}\text{Sr}:$ ^{86}Sr ratios to reflect salinity history. In addition, we plan to expand on this tool by including the carbon isotope history $^{12}\text{C}:$ ^{13}C since this tool has recently discovered high resolution relationships with salinity. Strontium isotopes samples are collected with the ablation multi-collector, inductively coupled plasma mass spectroscopy (ICP-MS) and the carbon isotopes are collected by micro-milling samples from the otolith and are analyzed with gas chromatography combustion isotope ratio mass spectrometry (GC-CIRMS).
- **Task 2-Genetic population structure of longfin smelt (PI: Israel, May):** This task will identify population structuring among tissue collections from the Bay-Delta, Klamath River, and west coast. Additional collection sites are currently being considered from museum collections at the California Academy of Science and Burke Museum at the University of Washington. In 2009, 17 microsatellite markers were optimized from the Bay-Delta and Washington State longfin smelt collections, and subsequently published for studying genetic variation in longfin smelt (Israel and May, 2010). These microsatellite markers appear to characterize 2 distinct populations of longfin smelt between the Bay-Delta and Washington State collections. They are now being used to evaluate genetic structuring based on clustering of genotypes among annual groups of collections at local, state, and regional scales. Traditional population differentiation analyses will be utilized for evaluating genetic structuring among tissue sample collections.

- **Task 3-Contemporary demographics of longfin smelt (PI: Israel, May):** This task will evaluate multiple annual collections of longfin smelt collected pre-POD and post-POD decline to compare genetic variation in the population during these periods. This information can be insightful into demographics, effective population size, and assessing bottlenecks within a population.

Time period: 2009 - 2010

Resources required:

Cost: \$113,000

PI(s): Bernie May and James Hobbs (UCD), Josh Israel (USBR).

Contract needed / in place: This work is being included in the DWR-UCD umbrella contract, currently in process.

Contract manager(s): Ted Sommer (DWR) and Bernie May (UCD).

Term of contract: Through June 30, 2010.

Personnel: Emily Ringelman (UCD)

Equipment:

Deliverables and dates:

- Manuscript accepted to Conservation Genetics (spring 2010).
- Oral presentation at the annual IEP Workshop (May 2010).
- Oral presentation at the Delta Science Conference (October 2010).
- Final report, which will comprise draft manuscripts for submission to journals or the IEP Newsletter (spring 2011).
- Databases, as appropriate (July 2011).

Comments: The project is intended to complement "data-mining" studies (e.g., Feyrer, Baxter) that will examine changes in longfin smelt habitat, abundance, and distribution.

Effects of Wastewater Management on Primary Productivity in the Delta

IEP 2010-138

Point person: Anke Mueller-Solger (DSC)

Lead Agency: SFSU, RTC & Water Boards

Questions: Do ambient ammonia levels in the Sacramento River affect phytoplankton primary production and community composition in the Delta?

Description: Primary production rates and standing chlorophyll levels in the Delta are among the lowest of all major estuaries in the world and may be declining further. The reason(s) for this are unclear, but decreasing primary production rates are cited as a possible cause of the pelagic organism decline (IEP, 2007). Recent work by Dugdale et al. (2007) and Wilkerson et al. (2006) has shown that elevated ammonia concentrations reduce phytoplankton production rates in San Francisco and Suisun bays by inhibiting nitrate uptake. Should phytoplankton production in the freshwater Delta be suppressed in the same way, this may contribute to the long-term declines in pelagic productivity and thus constitute an important "bottom-up" factor in the POD conceptual model. The primary sources of ammonia to the Delta are sewage treatment plants, principally the SRWTP.

Description: This work plan element focuses on the SRWTP discharge to the lower Sacramento River. SRWTP employs secondary treatment and the main form of nitrogen in its effluent is ammonium. Field studies will include transect surveys of nutrients and phytoplankton as well as phytoplankton "grow-out" enclosures experiments (Dugdale et al, 2007) upstream and

downstream of the SRWTP discharge location. Controlled laboratory experiments with added effluent, ammonium, and nitrate will complement the field study.

Time period: 2008 – 2010

Resources required:

Cost: \$119,000

PI(s): Richard Dugdale, Alex Parker and Francis Wilkerson (SFSU)

Contract needed / in place: SWRCB contract in place.

Contract manager(s): Mark Gowdy (SWRCB) and Chris Foe (CVRWQCB)

Term of contract: January 2008 – March 2010

Personnel: Above named investigators and 2 technicians.

Equipment: None

Deliverables and dates:

- Annual reports to the SWRCB, Delta Science Program and POD MT.
- IEP Newsletter articles, presentations as appropriate and a manuscript.

Comments: This element was completed as of March 31, 2010.

Effects of Microcystis aeruginosa on Threadfin Shad (Dorosoma petenense)

IEP 2010-139

Point person: Ted Sommer (DWR)

Lead Agency: UCD

Questions: What are the lethal and sublethal effects of microcystins (MCs) on growth and survival of threadfin shad (TFS)? How does the accumulation and fate of MCs affect TFS? The hypotheses to be tested are:

- H1: Microcystis can cause lethal and sublethal toxicity to embryo, larval, and juvenile TFS.
- H2: Single-celled form of MC is more lethal than colonial form of MC in filter feeder such as TFS.
- H3: Microcystis affect the quality and quantity of *E. affinis*, and *P. forbesi* and thus lead to poor survival and growth of TFS.
- H4: Microcystis affects the growth and reproduction of TFS.

Description: TFS are small fresh water plankton feeders which inhabit open waters of reservoirs, lakes, and shallow water habitat in the upper San Francisco Estuary (SFE). The purpose of this study is to examine the potential effects of the toxic alga *Microcystis*, which creates blooms that overlap substantially with the range of TFS. The working hypothesis is that TFS can be exposed to these toxins either during feeding (especially for filter feeder such as TFS) or passively when the toxins pass through gills during breathing. Work to date has successfully established culture of: (1) copepods (*Eurytemora affinis* and *Pseudodiaptomus forbesi*), (2) positive (microcystin-LR+) and negative (microcystin-LR-) strains of single-celled forms of *Microcystis aeruginosa*, and (3) a re-circulating temperature of controlled fish culture system in our laboratory.

The study is divided into 4 tasks. Task 1 will establish laboratory culture of TFS. Task 2 will perform lethal *Microcystis* studies on TFS. The work will provide information on the sensitivity of TFS to *Microcystis*. Data will be compared to microcystin studies in other fish species and copepods. The elements of this task will evaluate: (1) acute toxicity (lethal concentration that kills 50% of the TFS) of MCs on larval and juvenile TFS; (2) water exposure of larval and juvenile TFS to environmentally-relevant concentrations of MCs; and (3) dietary exposure of larval and juvenile TFS to single-celled and colonial form of *Microcystis*. Task 3

will focus on sublethal *Microcystis* studies on TFS including growth, histopathological, and reproductive effects of MCs. Finally, task 4 will determine bioaccumulation and fate of MCs in TFS.

Time period: Ongoing through 2010.

Resources required:

Cost: \$178,000 for 2010 from POD sources.

PI(s): Swee Teh (UCD) and Peggy Lehman (DWR)

Contract needed / in place: Requires modification of existing contract.

Contract manager(s): Rich Breuer (DWR) and Swee Teh (UCD)

Term of contract: Through June 30, 2010.

Personnel: UCD staff; Kevin Reece and Ted Sommer (DWR)

Equipment: UCD laboratory facilities, as well as some IEP boat time to help collect threadfin shad for laboratory cultures.

Deliverables and dates:

- Progress report, presentation and protocol for laboratory culture of threadfin shad (February 2010).
- Progress reports, presentations and publications for tasks 1, 2, 3 and 4 (December 2010).

Comments: None

Mark-Recapture Study to Estimate Delta Smelt Pre-screen Loss and Salvage Efficiency

IEP-related 2010-140 (2006 CALFED PSP)

Point person: Anke Mueller-Solger (DSC)

Lead Agency: USFWS

Questions: What is the relation between salvage and total entrainment losses for juvenile and adult delta smelt at the SWP?

Description: The purpose of this project is to evaluate the efficacy of mark-recapture tests to ensure a feasible approach to quantifying the extent of entrainment losses of juvenile and adult delta smelt in the south Delta. Delta smelt was historically one of the most common open-water species of fish in the Delta. Delta smelt declined significantly between the late 1970s and early 1980s and is now listed as a threatened species by the California and Federal Endangered Species Acts. Record low abundance indices for delta smelt and other pelagic fishes in the Delta have been observed since the early-mid 2000's. Leading factors potentially implicated in this pelagic organism decline are water project operations, introduced species and contaminants. Despite the lack of information to quantify absolute entrainment losses of delta smelt to water exports and diversions, such losses have long been assumed to be a factor contributing to the decline of delta smelt and other species, particularly in the south Delta where the SWP and CVP water export facilities are located. The tasks for the project are listed as follows:

- Task 1 - Project management.
- Task 2 - Culture delta smelt for mark-recapture experiments.
- Task 3 - Mark delta smelt for mark-recapture experiments.
- Task 4 - Mark-recapture experiments for juvenile delta smelt.
- Task 5 - Mark-recapture experiments for adult delta smelt.
- Task 6 - Analyze and interpret results of mark-recapture experiments.

Time period: August 2007 –July 2010

Resources required:

Cost: \$15,000 in 2010. This work is co-funded with USBR bridge funds and DSP.

PI(s): Gonzalo Castillo (USFWS), Robert Fujimura (DFG), Joan Lindberg (UCD), Jerry Morinaka (DFG), Victoria Poage (USFWS)

Contract needed / in place: CESU agreement with UCD is completed. The cooperative agreement with DFG is in place until December 11, 2010 (delayed due to unforeseen logistical issues).

Contract manager(s): Kim Webb (USFWS) and Gina Ford (DSP)

Term of contract: August 2007 – July 2010

Personnel: Above named investigators and Jason Dubois (DFG), Luke Ellison and Galen Tigan (UCD), and scientific aides.

Equipment:

Deliverables and dates:

- Progress report 1 and IEP Workshop presentation (March 2009).
- Progress report 2 (September 2009).
- Progress report 3 and IEP Workshop presentation (March 2010).
- Final report (September 2010).
- Submission of 1 manuscript to a scientific journal (October 2010).
- Delta Science Conference presentation (October 2010).

Comments: Field work was completed in 2009. Data analyses and manuscript preparation are in progress. A second methods manuscript about marking is anticipated in 2011.

3-D Modeling of the Delta

IEP 2010-141 (Expansion of existing element)

Point person: Fred Feyrer (USBR)

Lead Agency: DWR

Questions: What are the predicted pathways of particle movement through the Delta and what is the likelihood of entrainment in the pumps for different release locations and Delta conditions? How do wind-driven velocities influence the vertical migration behavior of fish in Clifton Court Forebay (CCF) and other regions?

Description: The focus of the present scope is to further refine the particle tracking simulation approach and simulate additional periods and scenarios. The scope is divided into 5 major tasks. Under the first task, the particle tracking approach will be refined to take into account additional biological variability (e.g., hatching rates and periods). Under the second task, the refined approach will be applied to periods of low abundance of delta smelt, such as 2007, 2008 or 2009. These particle tracking applications will focus largely on comparisons of model predictions with observed salvage to evaluate the ability of the modeling approach to predict entrainment during recent years, which are characterized by low delta smelt abundance. Under the third task, the model will be applied to stimulate adult delta smelt distribution and entrainment for a hypothesized “tidal surfing” behavior. Under the fourth task, the model will be applied to simulate a period of field observations in CCF. The fifth task includes meetings, presentations, and documentation of the work completed under this scope of work.

Time period: June 2009 – June 2010

Resources required:

Cost: Obligated with previous year funding.

PI(s): Edward Gross (Bay Modeling)

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuyse (USBR)

Term of contract: June 2009 – September 2010

Personnel: Edward Gross (Bay Modeling) and Michael MacWilliams (River Modeling)

Equipment: None

Deliverables and dates:

- IEP Presentation (February 2010).
- Final Technical Report (May 2010).
- Manuscript submitted to journal (September 2010).

Review and Synthesis of Existing Discharger and Ambient Surface Water Toxicity and Chemical Contaminant Monitoring in the Delta

IEP 2009-146 (Expansion of existing element)

Point person: Stephanie Fong (CVRWQCB)

Lead Agency: CVRWQCB

Questions: Is there evidence that the magnitude, duration and frequency of toxicity, and/or chemical concentrations presently being measured in the Delta might impair its aquatic beneficial uses? Is there evidence that water column toxicity is contributing, at least in part, to the POD? How might ongoing surface water monitoring be improved to more definitively answer the above 2 questions?

Description: The purpose of this work plan element is to develop and synthesize contaminant and toxicity information needed to support development of a comprehensive regional monitoring program for the Delta and review chemical, toxicity, and histological data from monitoring programs to investigate impacts of contaminants while assessing the impact of discharges from irrigated lands on beneficial uses in the Delta, specifically on emerging concerns related to the POD. Specific tasks are as follows:

- Prepare a synthesis report of the monitoring programs conducting ambient monitoring in the Delta and 30 miles upstream on the Sacramento and San Joaquin rivers, and include monitoring conducted under all Regional and State Water programs (e.g., storm water and wastewater NPDES permits, Irrigated Lands Regulatory Program, Surface Water Ambient Monitoring Program (SWAMP), etc.), the IEP and USGS.
 - The report will summarize the programmatic objectives for the monitoring, parameters and locations being monitored, the frequency of monitoring, and the anticipated term of future monitoring.
 - The report will include an estimate of the cost of the ambient monitoring summarized in the synthesis report. The cost estimate shall be broken down by funding entity, sampling costs, and analysis costs.
- Prepare recommendations for a framework to coordinate monitoring summarized in the synthesis report under above named task. The framework should build upon the ongoing efforts to coordinate monitoring in the San Joaquin and Sacramento River basins as well as the Delta Science Program's proposal to develop a monitoring strategy for the Delta.

Time period: 2008 - 2010

Resources required:

Cost: \$141,000 in 2010.

PI(s): Michael Johnson (UCD)

Contract needed / in place: In place.

Contract manager(s): Mark Gowdy (SWRCB)

Term of contract: 2008 - 2010

Personnel: Above named investigator and additional technicians.

Equipment: None

Deliverables and dates:

- Synthesis Report written jointly by the contractor and expert panel (spring 2010).
- Final oral report to the Contaminants Project Work Team.

Comments: None

BREACH III: Evaluating and Predicting ‘Restoration Thresholds’ in Evolving Freshwater-Tidal Marshes

IEP-related, 2010-147 (Ecosystem Restoration Program)

Point person: Pete Hrodey (USFWS)

Lead Agency: USFWS

Description: The purpose of this project is to provide a predictive level of understanding about (1) how abiotic and biotic factors influence a restoring (levee breach) wetland, Liberty Island and Little Holland Tract, control vegetation colonization and expansion and subsequent responses by native fish and wildlife, and (2) how restoration processes influence local flooding and levee erosion over the course of the restoration. A quantitative approach to predicting the ecological responses to change in habitat structure will be developed as a restoring system passes through the vegetation recolonization threshold and continues to expand into a predominantly vegetated wetland landscape. The models used will also be valuable for interpreting or, with modification, even predicting alternative flood conveyance scenarios as the Island evolves. Specific tasks are as follows:

- **Informing Flood and Erosion Hazard Management Decisions**
This task will address the knowledge gap concerning the risks of flooding and levee erosion associated with the projected geomorphic evolution of Liberty Island from flooded island to tule marsh and mudflat. The modeling results will inform resource management and planning processes related to the Lower Yolo Bypass.
- **Channel/Tidal Flat Morphology and Wave Climate**
This task will measure channel and mudflat dimensions to provide time series data on landscape features and processes, and investigate the geomorphic and ecological processes that define channel evolution across Liberty Island.
- **Landscape Structure and Change**
This task will describe vegetation and geomorphology changes of Liberty Island from 1997 to 2008. Geomorphology, topography and data from other tasks in this agreement will be used to develop rules that govern the expansion and development of emergent vegetation after initial vegetation colonization. Qualitative and quantitative predictions of restoration landscapes and tidal channel development, both at Liberty Island and other Bay-Delta tidal freshwater restoration sites, will be produced.
- **Elevation Change**
This task will quantify above-ground vegetative processes, below-ground biomass accumulation, and sediment deposition responses to plant colonization on open mudflats as factors in elevation change of marsh substrate. Field studies will be structured to identify differences in soil building processes (i.e., rates, and the relative role of mineral sediment vs. organic accumulation) in newly colonizing vegetated areas.
- **Plant Colonization Dynamics**

This task will study physiological tolerances and biotic interactions of key emergent wetland plant species under 2 sets of dispersal conditions in order to discover the opportunities and constraints for the establishment of key plant species. The experimental results will be used to produce an assessment of the range of suitable habitats for desirable species.

- **Macroinvertebrate Response**

This task will identify macroinvertebrate assemblages associated with spatial and temporal development of the restoring tidal freshwater wetland. These assemblages will be monitored, in the vicinity of restoring wetlands, for changes during the course of vegetation colonization. Emphasis will be placed on monitoring the assemblages known to be the prominent prey of key fish species.

- **Nekton Response**

This task will evaluate fish use of incipient vegetation colonization habitat structure and early expanding habitats by studying the use of restoring wetlands by specific life stages of delta smelt, longfin smelt, Chinook salmon, striped bass, threadfin shad and Sacramento splittail.

- **Food Web Sources and Pathways**

This task will conduct monitoring for food web sources and pathways at Liberty Island to determine whether Liberty Island is a source or sink of organic carbon, phytoplankton and zooplankton biomass, nutrients and suspended solids. Local, exported and imported productivity will be characterized at various scales.

- **Hydrodynamic and Morphological Modeling**

This task will develop and apply a detailed numerical model to improve understanding of the physical processes (e.g., sediment supply, delivery, resuspension and redistribution) that control morphologic evolution of Liberty Island and to assess the validity of using models as predictive tools. Recommendations on how management and restoration activities can be applied to leveed sites within Liberty Island for a variety of desired endpoints will be prepared.

- **Development of Ecological Modeling Tools**

This task will develop the Liberty Island Basin Model from process-based algorithms using existing models, modules, and other rules and inputs. Model parameters will be collected and developed sufficiently to enable resource managers and restoration planners to assess restoration and management activity for long-term trajectories associated with levee breach restoration and other landscape alterations.

- **Development and Use of Predictive Modeling Tools Using a Synthesis Process**

The models, modules, data, and other rules and inputs developed in previous tasks will be integrated to produce a system of linked hydrodynamic-ecological numerical models (model system). The model system will convert resource management and ecosystem restoration parameter inputs into ecosystem outcomes for a specified location and range of variables.

Time period: March 2008 – March 2011

Resources required:

Cost: \$2,440,000 over 3 years (CALFED ERP → DFG → USFWS → subcontractors) with approximately \$1,100,000 allocated for 2010.

PI(s): Pete Hrodey (USFWS), Charles Simenstad (UW), Philip Williams (PWA), Nadav Nur (PRBO), Denise Reed (UNO), Mark Hester (ULL), Enrique Reyes (ECU), Stephen Bollens (WSUV), and Peggy Lehman (DWR)

Contract needed / in place: Sub-contracts not in place and affected by the December 2008 'stop-work' order.

Contract manager(s): Pete Hrodey (USFWS), Steven Rodriguez (DFG) and Leann Androvich (GCAP Services)

Term of contract: March 2008 – March 2011

Personnel: Above named investigators and additional technicians.

Equipment: A variety of equipment will be purchased by the various subcontractors and agencies associated with this work.

Deliverables and dates:

Year 1:

- Quarterly progress reports.
- Breach III internet website.
- Flood and Erosion Hazard Analysis Report.
- Restoration Practitioners and Resource Managers Modeling workshop.

Year 2:

- Quarterly progress reports.
- Restoration Practitioners and Resource Managers Modeling workshop.

Year 3:

- Quarterly progress reports.
- Restoration Practitioners and Resource Managers Modeling workshop.
- Final task reports.
- Final Hydrodynamic and Morphological Model Development.
- Liberty Island Basin Model Development.
- Final report.
- Several peer-reviewed journal articles to follow.

Comments: None

Spatial and Temporal Variability of Delta Water Temperatures: Long-term Trends and the Dynamics of Refugia

IEP 2010-148

Point person: Larry Brown (USGS)

Lead Agency: UCB

Questions: How do Delta water temperatures vary temporally and spatially? What are the historical trends in water temperature? Are there trends at the decadal and longer timescales that may be explanatory of the observed population declines? Spatially, how representative are point measurements of temperature in determining the thermal habitat of fish and other pelagic organisms? As an extension to this question, what is the spatial extent and persistence of thermal refugia? How do water temperature dynamics differ between Delta sloughs and the primary river channels? What is the spatial scale of interaction between these habitats?

Description: As global climate trends increase, Delta water temperatures are expected to increase, making delta smelt and other temperature-sensitive species more vulnerable. Previous work through the Computational Assessments of Scenarios of Change for the Delta Ecosystem (CASCaDE) project has developed statistical models of water temperature that can effectively

forecast over a 100-year time horizon for a number of different climate change scenarios. The statistical model employed uses air temperature, solar radiation and water temperature from the previous day. The resulting model is more than just a correlation of water temperature with atmospheric drivers, but also has an auto-regressive component. This approach successfully projects what water temperature would be measured at the instrumentation site but it is unclear how representative those temperature measurements are of local or regional water temperatures. Because all of the long-term stations are located along the channels of the Sacramento or San Joaquin rivers, the question of how representative the observed water temperatures are of conditions in other sloughs and channels in the Delta still needs to be investigated. Specific tasks are as follows:

- **Task 1- Historical analysis of Delta water temperatures:** The statistical model will be applied to the last century to examine how water temperatures have evolved in the last 50 – 100 years. The first step will be to collect and analyze historical atmospheric data (air temperature) which provides the dominate forcing of the statistical model. Once the forcing data is identified and processed, the model will be explored in back-casting mode, where calculations are run backwards in time. Analyses of backcast temperatures will be pursued as motivated by discussion with the POD managers during presentations of results and other meetings.
- **Task 2- Evaluating spatial variability of Delta water temperatures:** This task will be accomplished by performing shorter timescale observations of spatial variability focused on the local variation of temperature in the Cache Slough complex, which serves as both a representative Delta habitat and has its own distinct ecological interest. Within Cache Slough, using direct observations of flows, temperature and conductivity, we will analyze the spatial and temporal variability of transport and refugia at a range of scales. Spatially, we will examine dispersion from the scale of the Cache Slough-Liberty Island complex to the mixing of differentially heated waters within a channel to the detailed exchange between shallow habitats and the deeper sloughs. Temporally, we will examine tidal and seasonal variations. Information from this task will help in understanding the dynamics of water temperature and thermal refugia in dendritic channel systems that include shallow water habitats.
- **Task 3-Analysis and modeling:** The primary modeling goal will be to develop descriptions of the spatial and temporal structure of water temperature locally and regionally surrounding historical measurements sites. Additionally, the investigators aim to develop a fundamental understanding of how temperature refugia may be developed and maintained.

Time period: June 2009 – June 2011

Resources required:

Cost: Funding was obligated with previous year funds.

PI(s): Mark Stacey (UCB), Erwin Van Nieuwenhuyse (USBR)

Contract needed / in place: In place via CESU.

Contract manager(s): Mark Stacey (UCB) and Erwin Van Nieuwenhuyse (USBR).

Term of contract: June 2009 – May 2011, considering an extension.

Personnel: Graduate student at UCB.

Equipment: None.

Deliverables and dates:

- Annual report including a summary of data collected and analysis pursued (December 2010).
- Two peer-reviewed publications and a PhD thesis will result from this research.

Plankton Dynamics in the Sacramento-San Joaquin Delta: Long-term Trends and Trophic Interactions

IEP-related 2010-150 (2009 CALFED Science Fellow)

Point person: Anke Mueller-Solger (DSC)

Lead Agency: UCD

Questions: What are the long-term trends of the Delta's zooplankton community and can distinct sub-regions be identified that show similar patterns? What are the long-term patterns in zooplankton species and functional groups? How does phytoplankton and environmental variability affect zooplankton production on a Delta-wide scale and appropriate sub-regions? Are seasonal patterns between primary producers and zooplankton consistent throughout the sampling record? How do changes in plankton community composition relate to biotic and environmental variation?

Description: This project seeks to identify: (1) long-term spatial and temporal patterns in zooplankton; (2) long-term interactions between primary producers and zooplankton; and (3) biotic interactions in the plankton community. Analyses of historical data, trends, seasonal variability, and foodwebs will be conducted using a variety of techniques appropriate to each analysis area.

Time period: September 2008—September 2010

Resources required:

Cost: \$83,000 in 2010.

PI(s): Monika Winder and Geoffrey Schladow (UCD)

Contract needed / in place: In place.

Contract manager(s): Rebecca Fris (DSP)

Term of contract: September 2008 – September 2010

Personnel: Above named investigators.

Equipment: None

Deliverables and dates:

Year 1:

- Annual progress report.
- Presentations at local (Delta Science Conference) and national or international professional meetings.
- Draft of first manuscript.

Year 2:

- Annual progress report and final research report summarizing results and accomplishments.
- Presentations at local (Delta Science Conference) and national or international professional meetings.
- Peer-reviewed scientific publications (at least 2 are anticipated).

Comments: None

Environmental Controls on the Distribution of Harmful Algae and Their Toxins in San Francisco Bay

IEP-related 2010-152 (2009 CALFED Science Fellow)

Point person: Anke Mueller-Solger (DSC)

Lead Agency: DSP

Questions: What predicted environmental changes favor dinoflagellates and cyanobacterial growth in the South Bay and Delta, respectively? What predicted environmental changes will result in increased frequency in blooms of these species? Does alleviating light limitations increase the influence of nutrient availability and relative concentrations, and does that result in enhanced toxicity of the harmful algae? Can spatial and temporal mapping of harmful algae and their toxins help explain the triggers for the algal blooms, and provide a baseline for measuring future trends and forecasting?

Description: This project will determine the distribution of harmful algae and their toxins in San Francisco Bay and characterize the environmental parameters that control toxin production by harmful algae in the Bay. It will combine monitoring and mapping of biological, chemical, and physical components throughout the Bay and Delta along with controlled manipulations to examine specific parameters likely to control growth and toxicity in the natural population of these species.

Time period: September 2008—August 2010

Resources required:

Cost: \$82,000 in 2010.

PI(s): Cecile Mioni and Adina Paytan (UCSC)

Contract needed / in place: In place.

Contract manager(s): Rebecca Fris (DSP)

Term of contract: September 2008 – September 2010

Personnel: Above named investigators.

Equipment:

Deliverables and dates:

- Monthly monitoring and data collection of harmful algae and toxins in the San Francisco Bay and Delta.
- *In-situ* incubation experiments conducted seasonally.
- Publications describing the results of the research.
- Presentations describing the results of the research.
- Monthly report cards describing the health of the Bay, published on the website.
- Monthly reports of elevated levels of toxins or toxic algal in the “Harmful Algal Blooms” section of the website.

Comments: None

Comparison of Nutrient Sources and Phytoplankton Growth and Species Composition in the Sacramento and San Joaquin Rivers: Their Roles in Determining Productivity and Food Web Conditions in Suisun Bay and the Delta

IEP-related 2009-153 (2008 CALFED Science supplemental grant)

Point person: Anke Mueller-Solger (DSC)

Lead Agency: SFSU-RTC & DSP

Questions: How do differences in nutrient and phytoplankton community composition between the San Joaquin (SJ) and Sacramento (Sac) rivers influence conditions downstream in the Delta and Suisun Bay? How do phytoplankton growth rates and community structure respond to the

differences in nutrient concentrations in the SJ River vs. the Sac River, resulting from differences in wastewater treatment?

Description: Low primary production rates and changes in phytoplankton community composition may play an important role in the pelagic organism decline. Recent work by Dugdale et al. (2007) and Wilkerson et al. (2006) has shown that elevated ammonium concentrations reduce phytoplankton production rates in San Francisco and Suisun bays by inhibiting nitrate uptake. The primary sources of ammonium to the Delta are wastewater treatment plants (WWTPs). It is not yet known whether ammonium at concentrations measured in the Delta, inhibits freshwater diatom production rates (similar to the inhibition observed in San Francisco and Suisun bays) and thus could be a contributing cause to the low primary production rates in the Delta.

This work plan element is closely linked to an ongoing DSP project (PIs: Dugdale and Wilkerson; April 2007- February 2010, Agreement #1039) focusing on phytoplankton production in Suisun Bay and extends these investigations into the Delta. It focuses on the effects of the 2 main WWTPs in the Delta: the SRWTP and the Stockton WWTP ("Stockton"). The SRWTP is the largest sewage treatment plant in the Delta. Its effluent is discharged into the Sacramento River at Freeport. It employs secondary treatment and the main form of nitrogen in its effluent is ammonium. The Stockton WWTP is the largest WWTP in the southern part of the Delta. It completed an upgrade to tertiary treatment in 2008 and now discharges primarily nitrate into the lower San Joaquin River in Stockton. These treatment differences, along with differences in river nutrient loadings, offer a great opportunity for a comparative investigation of the effects of regional differences in river and WWTP nutrient loadings on Delta phytoplankton. Field studies will include transect surveys of nutrients and phytoplankton along the Sacramento and San Joaquin rivers and as well as phytoplankton "grow-out" enclosures experiments (Dugdale et al., 2007).

Field sampling will consist of under-way measurements of temperature, salinity, turbidity, fluorescence, as well as continuous sampling using flow cytometry and the FluoroProbe. Discrete sampling at river and Delta stations will include vertical profiles of temperature and salinity, and light penetration using a conductivity, temperature and depth (CTD) instrument fitted with a [photosynthetically active radiation](#) (PAR) sensor as well as a Secchi disk. Water will be sampled at the surface and near the bottom for determination of inorganic nutrients (NO_3^- , NO_2^- , NH_4^+ , PO_4 , Si(OH)_4), dissolved inorganic carbon (DIC), and size fractionated chlorophyll. Primary production and phytoplankton NO_3^- and NH_4^+ uptake will be assessed under saturating light (50% of PAR) and light limited (10% of PAR) conditions using $^{15}\text{N}/^{13}\text{C}$ tracer methods in light attenuated, flow-through incubators. Comparison of NH_4^+ gradients along the river with $^{15}\text{NH}_4^+$ uptake rates will provide an indication of nitrification. Samples will be collected for phytoplankton identification by microscopy, flow cytometry high-performance liquid chromatography (HPLC) technique, and the FluoroProbe at each of the 24 discrete stations. Grow-out experiments will be conducted at locations downstream of the Sacramento and Stockton WWTPs. If NH_4^+ depletion rates in grow-outs are greater than the measured $^{15}\text{NH}_4^+$ uptake rates, then this will suggest the presence of nitrification.

Time period: 2008 – 2010

Resources required:

Cost: \$338,377 through a DSP (previously CALFED) supplemental grant.

PI(s): Richard Dugdale, Alex Parker and Frances Wilkerson (SFSU) and Anke Mueller-Solger (DSC)

Contract needed / in place: In place.

Contract manager(s): Shem Ayalew (DSP)

Term of contract: January 2008 – April 15, 2010

Personnel: Above named investigators and 2 technicians.

Equipment: None

Deliverables and dates:

- Data and technical reports, presentations at IEP meetings, EET workshops and the Delta Science Conference.
- Publications in peer-reviewed journals.

Spatial and Temporal Quantification of Pesticide Loadings to the Sacramento River, San Joaquin River, and Bay-Delta to Guide Risk Assessment for Sensitive Species

IEP-related 2009-154 (DSP-funded)

Point person: Rich Breuer (DWR)

Lead Agency: DWR

Questions: What is the fate and transport of agricultural chemicals in the Sacramento and San Joaquin rivers? When and where should we be looking for potential toxicity based on modeling results?

Description: A weight-of-evidence analysis is being conducted to identify major sources of pesticide loadings to the Sacramento River, San Joaquin River, and Bay-Delta estuary. The objective of this study is to improve decision making and optimize resource spending across a number of federal, state, and regional water quality programs. Objectives are being addressed through a combination of tools, including GIS analysis, simulation modeling, and an evaluation of existing in-stream monitoring. Results are being used to: (1) provide further knowledge of the fate and transport of agricultural chemicals (e.g., copper, organophosphates) and emerging pesticides (e.g., pyrethroids), (2) match results to the location of sensitive species critical habitats, (3) identify and rank pesticide source areas, (4) evaluate implications of future pesticide use trends and changes in climatic conditions, (5) aid in developing plans to improve ecosystem quality and water quality by strategic placement of best management practices and hydrologic operations, (6) support future monitoring programs (strategic locations, sampling frequency), (7) link results to life cycle models currently under development for striped bass and delta smelt, as well as existing models for salmonids, and (8) provide a data-link to support other water quality models and population models.

Time period: 2008 – 2010

Resources required:

Cost: \$395,000

PI(s): Minghua Zhang(UCD) and W. Martin Williams (Waterborne Environmental, Inc.)

Contract needed / in place: DSP receivable and subcontracts to UCD and Waterborne Inc. in place.

Contract manager(s): Brianne Noble (DWR)

Term of contract: July 2008 – June 2010

Personnel: Above named investigators and Technical Advisory Group (TAG) that provides technical direction on the project to ensure that cross-agency goals are met. The

TAG consists of 15 members representing federal and state agencies, universities and private industries. Their involvement will begin at the initiation of the project to coordinate data collection and avoid duplication of efforts with other activities that either have occurred or are in progress. Interaction with the TAG will be continuous and iterative.

Equipment: None

Deliverables and dates: The TAG will receive progress reports and other interim communication and redirect efforts as necessary to maximize the success of this study.

2010 New IEP POD Elements

Acute and Chronic Toxicity of Contaminant Mixtures and Multiple Stressors

IEP 2010-157

Point person: Stephanie Fong (CVRWQCB)

Lead Agency: CVRWQCB

Questions: How do various contaminant mixtures affect their toxicity to *Hyaella azteca*?

Description: This study will address the mixture toxicity of several contaminants of concern in the Delta in the presence and absence of other environmental stressors (temperature, food deprivation). A sensitive resident invertebrate species, the amphipod *Hyaella azteca*, will be used as test organism, and toxicity will be quantified by means of acute (10 day survival) and chronic (growth, swimming ability) endpoints.

Time period: March 2010-June 2011

Resources required:

Cost: \$40,000 contract and CVRWQCB staff time.

PI(s): Inge Werner

Contract needed / in place: In place May 2010.

Contract manager(s): Stephanie Fong (CVRWQCB) and Inge Werner (UCD ATL)

Term of contract: N/A

Personnel: Inge Werner and graduate student (UCD)

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Peer-reviewed study plan (July 2010).
- Progress reports and/or participation in the POD Contaminants Work Team (CWT), quarterly throughout the study.
- Final report (June 2011).

Comments: None

Advancing Procedures for Extracting and Recovering Chemicals of Concern from Sediment Interstitial Water

IEP 2010-158

Point person: Stephanie Fong (CVRWQCB)

Lead Agency: CVRWQCB

Questions: How do sediment-bound toxicants react to TIE manipulations when extracted from the interstitial water?

Description: This study will address a critical need to develop better techniques to extract and recover organic chemicals in sediment interstitial water, as part of the TIE Phase II toxicant identification process. The study will implement a set of tests to evaluate better extraction and elution techniques, and is designed to complement other ongoing efforts.

Time period: February 2010 – June 2011

Resources required:

Cost: \$40,000 contract and CVRWQCB staff time.

PI(s): Ron Tjeerdema and John Hunt (UCD)

Contract needed / in place: In place February 2, 2010.

Contract manager(s): Stephanie Fong (CVRWQCB)

Term of contract: N/A

Personnel: John Hunt, Bryn Phillips, and laboratory staff (UCD) and Stephanie Fong (CVRWQCB)

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Peer-reviewed study plan (July 2010).
- Progress reports and/or participation in the POD CWT, quarterly throughout the study.
- Final report (June 2011).

Comments: This will augment ongoing efforts by San Francisco Estuary Institute (SFEI) .

Investigation of Pyrethroid Pesticides in the American River

IEP 2010-159

Point person: Stephanie Fong (CVRWQCB)

Lead Agency: CVRWQCB

Questions: Is storm water runoff to the American River toxic to *H. azteca*?

Description: Samples will be collected during 3 separate rain events at a minimum of 4 sites on the American River below Folsom Dam. Each river site will be sampled between 2 and 4 times during each rain event. Samples will be collected at up to 5 of the largest discharges of runoff to the American River below Folsom Dam. Each discharge site shall be sampled 2 times during 3 separate rain events, yielding up to 30 samples. All samples will undergo chemical analysis for pyrethroids. River samples (not discharge samples) will be tested with *H. azteca*, and toxic samples will undergo toxicity identification evaluations and follow-up sampling.

Time period: December 1, 2009 – May 30, 2011

Resources required:

Cost: \$100,000

PI(s): Don Weston (UCB)

Contract needed / in place: In place May 26, 2010.

Contract manager(s): Stephanie Fong (CVRWQCB) and Don Weston (UCB)

Term of contract:

Personnel: Don Weston and laboratory staff (UCD)

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Quality Assurance Project Plan (QAPP) and monitoring plan (July 2010).
- Sample collection (March 2010).
- Submission of electronic data (May 2010).
- Progress reports and/or participation in the POD CWT, quarterly throughout the study.

- Final report (May 2011).
- Journal article, if accepted (June 2011).

Comments: This study is following up on toxic results from monitoring performed in early 2009.

Full Life-Cycle Bioassay Approach to Assess Chronic Exposure of Pseudodiaptomus forbesi to Ammonia

IEP 2010-160

Point person: Mark Gowdy (SWRCB)

Lead Agency: SWRCB

Questions: What are effects of chronic exposure to ammonia for the copepod *Pseudodiaptomus forbesi*?

Description: The study will use full life-cycle tests to assess the effects of chronic exposure to ammonia for the estuarine copepod *Pseudodiaptomus forbesi*. This copepod is the dominant zooplankton during summer and fall months and is important food for pelagic fishes in the upper San Francisco Estuary. *P. forbesi* has 3 distinguishable life stages (i.e., naupliar, copepodite (juvenile), and sexually mature adult stages). By performing a 30 day bioassay at 25°C, eggs will hatch and grow up to all life stages, therefore facilitating counting at test termination. Copepods have an easily distinguishable sexual morphology and their egg sacs are external, making clutch sizes easy to count. Such characteristics will allow for the assessment of life table parameters and sublethal ammonia effects on the reproduction of this copepod in a period of 30 days.

Time period: January 2010 – April 2010

Resources required:

Cost: \$77,000

PI(s): Swee the (UCD)

Contract needed / in place: In place December 2009

Contract manager(s): Mark Gowdy (SWRCB)

Term of contract: January 2010 – April 2010

Personnel: Swee Teh and ATL staff

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Draft final report (March 2010).
- Presentation of results at the POD-CWT Conference (date TBD).
- Final summary report (April 2010).

Comments: None

Acute Toxicity of Ammonia/SRWTP Effluent on Delta Smelt and Surrogate Species

IEP 2010-161

Point person: Mark Gowdy (SWRCB)

Lead Agency: SWRCB

Questions: What is the range of NOEC and low LOEC effect ranges of SRWTP effluent mixed into Sacramento River water from Garcia Bend for delta smelt? Can larval rainbow trout be used as a surrogate species for toxicity testing and toxicity identification evaluations?

Description: The goal of toxicity testing (including reference toxicants) with delta smelt is to determine the range of SRWTP effluent toxicity. In previous toxicity tests with delta smelt, ammonia/um in effluent was shown to be more toxic (7-day LC50: 5.4 mg/L) than ammonia/um

alone (7-day LC50: 7.5 mg/L). In addition, larval rainbow trout will be used in concurrent toxicity tests (including reference toxicants) to evaluate if they could be used as a surrogate species for delta smelt in toxicity testing and enable the application of toxicity identification evaluations when needed.

Time period: January 2010 – December 2010

Resources required:

Cost: \$65,000

PI(s): Inge Werner (UCD)

Contract needed / in place: In place.

Contract manager(s): Mark Gowdy (SWRCB) and Inge Werner (UCD ATL)

Term of contract: January 2010 – December 2010

Personnel: Inge Werner and ATL staff

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Final draft report (October 2010).
- Peer review of draft report and submission of final report (December 2010).

Comments: None.

Potential Loss of Life History Variation and the Decline of Delta Smelt

IEP 2010-162

Point person: Mark Gowdy (SWRCB)

Lead Agency: SWRCB

Questions: Has selective entrainment of early-spawned larvae been of sufficient magnitude and duration to cause undesirable evolutionary change in delta smelt? If such changes have occurred, how can management reverse the process and contribute to restoration of the species?

Description: This project will investigate the so-called, Big Mama hypothesis (term coined by Kenny Rose, LSU), for the decline of delta smelt. Larger, more robust, delta smelt (i.e., big mamas) tend to reproduce earlier in spring, spawning larger numbers of better provisioned larvae that are more likely to survive and reproduce than those hatching from smaller parents later in the spawning season. Early-spawned larvae, however, comprise the majority of annual entrainment losses in the SWP and CVP export facilities. This imposes a form of artificial selection on the population that may oppose natural selection. Moreover, the very short generation time (only 1 year) of delta smelt implies that any potentially undesirable evolutionary changes would manifest rather quickly (e.g., < 10 years). Given the seriousness of the delta smelt problem, verifying the accuracy and scope of the Big Mama hypothesis is critical for understanding restoration options for this imperiled species. The primary product of this work will be a completed manuscript for submission to a peer-reviewed scientific journal.

Time period: October 2009 – July 2010

Resources required:

Cost: \$32,000

PI(s): Bill Bennett (UCD)

Contract needed / in place: In place

Contract manager(s): Mark Gowdy (SWRCB)

Term of contract: October 2009 – July 2010

Personnel: Bill Bennett (UCD)

Equipment: Software upgrade.

Deliverables and dates:

- Manuscript evaluating the Big Mama hypothesis (July 2010).

Comments: None

Comparison of Flow and Transport Models for the Sacramento-San Joaquin Delta

IEP 2010-163

Point person: Mark Gowdy (SWRCB)

Lead Agency: SWRCB

Questions: How do the 1- and 2-D model simulations of flow and contaminant transport in the delta compare relative to one another and against measured conditions?

Description: The purpose of this study is to compare the performance of diverse modeling approaches in 1- and 2-D to simulate the flow and transport in the Sacramento-San Joaquin Delta, to identify model improvement needs, and Delta locations with poorer flow representation. The project will define 7 flow scenarios (4 corresponding to low and high flows for which field data are available) to compare model performance, for a common set of boundary conditions. Two more scenarios will correspond to events of low and high pumping, to be defined. All models will be run with the same boundary conditions, the same domain and with the same flow forcing mechanisms. An additional scenario of sea level rise will be included in the analysis. Models to be compared will include: RMA 1- and 2-D models, RMA-TAM (tidally averaged model), and DSM2.

Time period: March 2009 – June 2010

Resources required:

Cost: \$59,000

PI(s): Fabian Bombardelli (UCD)

Contract needed / in place: January 2009

Contract manager(s): Mark Gowdy (SWRCB)

Term of contract: March 2009 – June 2010

Personnel: Fabian Bombardelli (UCD) and PhD student

Equipment: Required computer equipment and software is already available.

Deliverables and dates:

- Final report of results (June 2010).
- Presentation of results to the California Water Environmental Modeling Forum (CWEMF) and DSM2 users' group meetings (dates to be determined).

Comments: None

Spatial and Temporal Variability in Nutrients in Suisun Bay in Relation to Spring Phytoplankton Blooms

IEP 2010-164

Point person: Karen Taberski (SFBRWQCB)

Lead Agency: SFBRWQCB

Questions: How do nutrients vary in Suisun Bay temporally and spatially and how does this relate to spring phytoplankton blooms? What are the major sources of ammonium in Suisun Bay?

Description: This study would be an extension of earlier work on the effect of ammonia on phytoplankton blooms in the estuary. The purpose of this project is to better understand the

variability of nutrients in Suisun Bay, their relation to spring phytoplankton blooms, and sources of ammonium.

Time period: Sampling will be conducted weekly from March to June 2010. Analysis should be completed by August 31, 2010.

Resources required:

Cost: \$25,000 contract from SWAMP to Dick Dugdale SFSU-RTC for nutrient analysis.

PI(s): Dick Dugdale (SFSU)

Contract needed / in place: In place February 1, 2010.

Contract manager(s): Russell Fairey/ Karen Taberski (SFBRWQCB)

Term of contract: February 1, 2010 – December 30, 2010

Personnel: Water Board SWAMP staff and SFSU Romberg Tiburon staff

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Data delivered to SWAMP Data MT at Moss Landing Marine Labs in SWAMP format on August 31, 2010.
- We intend to contract for a report next fall.

How is this activity related to HSG? The results of this analysis will contribute to the understanding of toxicant effects in the Delta.

Comments: None

Ammonia Sampling Program for the Sacramento-San Joaquin Delta Estuary

IEP 2010-165

Point person: Chris Foe (CVRWQCB)

Lead Agency: CVRWQCB

Questions: What are the concentrations and distribution of chlorophyll, ammonia, and various nutrients, primarily in the lower Sacramento River and northern Delta? Are total and unionized ammonia concentrations in these areas potentially toxic to sensitive resident aquatic organisms? Provide data to support development of an ammonia fate and transport model.

Description: The purpose of this study is to collect water quality data, including total and unionized ammonia, primarily in the lower Sacramento River and northern Delta to determine whether ambient concentrations are potentially toxic to sensitive resident aquatic organisms and to support development of an ammonia fate and transport model. A spatial emphasis is placed on the lower Sacramento River and northern Delta as the biological risk from elevated ammonia is likely to be greatest here. However, other areas of the estuary are also proposed for monitoring as it is likely that the SRWTP is not the only source of ammonia. Temporally, the sampling emphasizes the months of March through June as delta smelt spawn in the northern Delta around Lindsey and Cache Sloughs and juvenile salmon are migrating down the Sacramento River and out into the estuary during these months. Samples are being collected by CVRWQCB staff and analyzed by Randy Dahlgren's lab at UCD.

Time period: March 2009 – July 2010

Resources required:

Cost: \$68,000 (analytical services only)

PI(s): Chris Foe (CVRWQCB) and Randy Dahlgren (UCD)

Contract needed / in place: (analysis contract) March 2009

Contract manager(s): Mark Gowdy (SWRCB)

Term of contract: March 2009 – March 2010 (analytical contract)

Personnel: Chris Foe and CVRWQCB staff and Randy Dahlgren's lab staff

Equipment: Nothing new; typical sampling and laboratory equipment.

Deliverables and dates:

- Regional Board staff draft interpretive report (May 2010).
- Final report 30-days after receiving comments from POD-CWT.

Comments: None

Using Genetic Techniques to Detect Mississippi Silverside (*Menidia audens*) Predation on Larval Delta Smelt (*Hypomesus transpacificus*)

IEP 2010-166

Point person: Ted Sommer (DWR)

Lead Agency: DWR

Questions: Can PCR be used to detect delta smelt DNA in silverside guts? How sensitive is the PCR method to detecting delta smelt DNA in gut contents? Do wild silverside predate larval delta smelt and at what frequency? Are there any patterns between larval delta smelt predation and environmental variables (flow, turbidity, salinity, temperature) or other constituents in the gut?

Description: This project addresses top-down trophic effects on delta smelt populations by developing a new set of genetic tools for detecting predation of delta smelt larvae (though the technique will be equally applicable to juvenile and adult delta smelt). These genetic tools (PCR assays) are being developed for use on Mississippi silversides, which are theorized to predate larval delta smelt. Testing and characterization of the assays will use captive silversides and delta smelt, and will include experiments to model the degradation of smelt DNA in the guts of silversides and determine the sensitivity of the assay using dilution experiments. Cross-reactivity will also be tested using genetic samples from multiple fish species from around the Delta. The refined assay will then be used on wild silversides sampled from areas where larval delta smelt are known to occur for the purposes of detecting predation events.

Time period: June 2010 – June 2011

Resources required:

Cost: \$68,389

PI(s): Brian Schreier (DWR), Bernie May and Melinda Baerwald (UCD)

Contract needed / in place: In process.

Contract manager(s): Rich Breuer (DWR) and Bernie May (UCD)

Term of contract: TBD

Personnel: Brian Schreier and Nick Van Ark (DWR)

Equipment: Molecular reagents, lab supplies, fish aquaculture supplies, dissection and preservation supplies.

Deliverables and dates:

- Poster presentation at IEP Annual Workshop (2011).
- Results and assay details presented at Delta Science Conference and published in an appropriate peer-reviewed journal (October 2010).
- Results and assay details published in an appropriate peer-reviewed journal.
- Development and characterization of a successful assay will be reported in the IEP Newsletter.

Comments: This pilot study aims to develop tools that will be applicable for detecting predation on delta smelt by multiple different predators. These predators will be expanded in future studies

to include multiple species of non-native centrarchids and striped bass. Additionally, utilization of the assay to detect predation on longfin smelt will be relatively straight forward and may also be a future direction this research could take.

Investigating the Presence, Migration Patterns and Site Fidelity of Sub-Adult Striped Bass
2010-167

Point person: Ted Sommer (DWR)

Lead IEP Agency: DWR

Questions: What geographical areas are sub-adult bass using and when? How do the patterns of presence and movements vary seasonally, annually and between age classes?

Description: Striped bass (*Morone saxatilis*) are the major pelagic predator in the San Francisco Estuary. “Top down” effects from striped bass and other species are being evaluated as part of ongoing POD studies (Sommer et al., 2007). Of particular interest is whether striped bass have substantial effects on the threatened delta smelt. Since striped bass are a relatively well-monitored species, there has been good progress in understanding some of the major factors that affect striped bass populations (Kimmerer et al., 2001). However, there are important data gaps in our knowledge of this species. For example, current monitoring programs do not effectively measure the population and distribution of sub-adult (1-3 year old) striped bass, however this group is likely to be the most abundant group of pelagic predators in the estuary. Hence, there is a need to understand basic information about the distribution and movements of sub-adult fish in the estuary.

Using the existing 456 km telemetry array located between Colusa on the Sacramento River out to the Golden Gate on the San Francisco Bay, 100 sub-adult striped bass (290-350 mm) total will be tagged with V9-2L coded ultrasonic tag with an inter pulse burst interval of about 90 seconds and a life of 417 days. Equal numbers of fish will be caught and tagged between 3 general geographic areas: Sacramento River, Delta and San Francisco Bay.

Time period: June 2010 – July 2011

Resources required:

Cost: \$75,000

PI(s): Cynthia LeDoux-Bloom and Ted Sommer (DWR), Bernie May and Melinda Baerwald (UCD)

Contract needed / in place: Not needed.

Contract manager(s): N/A

Term of contract: N/A

Personnel: Brian Schreier and Cynthia LeDoux-Bloom (DWR)

Equipment: 100 Vemco tags, existing receiver array, small boat during tagging.

Deliverables and dates:

- Poster or oral presentation at the IEP Annual Workshop (2011).
- Presentation at 2011 Delta Council.
- IEP Newsletter article.
- Journal articles, if appropriate.

Comments: Top-down effects from striped bass and other species are being evaluated as part of ongoing POD studies (Sommer et al., 2007). The impact of striped bass on the POD is thought to be significant, although little data exists showing their migration and movement patterns due to lack of data. Knowledge of striped movement patterns will provide valuable information

regarding timing of habitat usage and suitability, and may have the potential ability for future integrated analyses of multiple data sets.

Monitoring inter-annual variability of delta smelt population contingents and growth

IEP 2010-168

Point person: Randy Baxter (DFG)

Lead Agency: UCD

Questions: Can life-history and growth of fish salvaged at CVP and SWP be compared to fish that survive the TNS to determine the effects of entrainment and salvage? What are the habitat effects on delta smelt population dynamics? Do life-history contingents vary inter-annually, in association with growth, freshwater outflow, water temperature, abundance? Does growth rate increase with increased fall outflow?

Description: The primary goal of this research is to gain a better understanding of the mechanisms (e.g. climate variability, hydrology) responsible for different life history contingents and how salvage at CVP and SWP could alter life history diversity. Archived samples from 1999 – 2008, already prepared for otolith microstructure and microchemistry studies, will be assayed with a laser line from the core to the edge to reconstruct the entire life history. Sub-adult and adult samples collected by the IEP in 2010 will be examined for microchemistry and growth rates will be quantified by otolith microstructure analysis.

Time period: June 2010 – June 2011

Resources required:

Cost: \$98,000

PI(s): Jim Hobbs (UCD)

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR)

Term of contract: 1 year.

Personnel: Above named investigator

Equipment: Needed equipment is already housed at the Interdisciplinary Center for Inductively Coupled Plasma Mass Spectrometry at UCD.

Deliverables and dates:

- Oral presentations will be provided to the IEP Management Team as requested.
- IEP Annual Workshop presentation, as requested.
- A manuscript will be submitted to a high-impact journal.

Delta Smelt Feeding and Food Web Interactions

IEP 2010-169

Point person: Larry Brown (USGS)

Lead Agency: SFSU

Questions: To what extent is individual delta smelt limited by the food supply in the LSZ, and how is food limitation affected by flow variability? What are the food availability and quality for delta smelt in the LSZ, and how are they affected by flow variability? What are the effects of gelatinous plankton in the LSZ on delta smelt and the food web, and how are they affected by flow variability?

Description: The proposed project investigates the food supply for delta smelt, how it is affected by predators and competitors, and how these interactions are affected by interannual variability in freshwater flow. This work comprises 3 tasks: (1) food limitation and functional response of

larval to early juvenile delta smelt; (2) population dynamics and production of the zooplankton prey of delta smelt; and (3) the role of jellyfish in the delta smelt food web.

In Task I, we will examine how the feeding of the early life history stages of delta smelt depends on concentration of copepod prey (=functional response). We will also measure the metabolism of early stage delta smelt, which will be used to estimate the energetic impact of the functional response, and investigate behavioral changes in feeding related to predator avoidance and turbidity. In Task II, we will employ a combination of field, laboratory, and modeling approaches to examine the population dynamics of the copepod *Pseudodiaptomus forbesi*, the principal prey of delta smelt in the LSZ and the freshwater Delta during fall. In Task III, we will measure the abundance and distribution of gelatinous predators throughout the LSZ during fall and quantify feeding rates of the gelatinous competitors of delta smelt using gut content analysis.

Time period: June 2010 – June 2011

Resources required:

Cost: \$399,840

PI(s): Wim Kimmerer and Lindsay Sullivan (RTC-SFSU); Jan Thompson (USGS)

Contract needed / in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Wim Kimmerer (SFSU)

Term of contract: TBD

Personnel: Toni Ignoffo and Anne Slaughter (RTC-SFSU)

Equipment: Stereomicroscope with image analysis capabilities.

Deliverables and dates:

- Poster or oral presentation at the IEP Annual Workshop or EET (February 2011).
- IEP Newsletter article (early 2011).
- Final 2010-11 report (July 2011).

How is this activity related to HSG? This proposed project addresses the "bottom-up" or food supply topic of the Pelagic POD conceptual model (Sommer et al., 2007; Baxter et al., 2008). As applied to delta smelt, this topic has 2 components or subsidiary questions: (1) To what extent is growth or survival of delta smelt food limited; and (2) What limits the availability of food for delta smelt? In particular, this project focuses on the food supply of delta smelt in the LSZ during late summer to fall. The principal reason for this focus is that salinity during this period has been persistently high since around 1999, roughly coincident with the POD, and several possible causal links between fall flow/salinity and smelt abundance have been identified (USFWS, 2008). This project addresses numerous potential mechanisms by which flow variability may affect the abundance of food for delta smelt. Broadly these include hydrodynamic effects, food effects, and predatory losses: (1) Changes in the physical shape or size of the low-salinity habitat cause a reduction in abundance of delta smelt or their food when X2 is high (landward); (2) Low flow results in reduced transport of phytoplankton and zooplankton from the freshwater Delta into the LSZ, reducing biomass in the LSZ; (3) Landward X2 exposes foodweb organisms to pumping losses, reducing abundance in freshwater and therefore transport to the LSZ; (4) Low flow results in a higher concentration of ammonium entering the LSZ, suppressing phytoplankton growth; (5) A landward X2 value (LSZ position) results in landward settlement of the clam *Corbula amurensis* and, in turn, reduction in biomass of phytoplankton, bacteria, microzooplankton, and mesozooplankton in the LSZ due to clam grazing; (6) A stable X2 value (particularly during clam recruitment periods) allows for a high abundance of clams to overlap with the LSZ over a period of months, maximizing consumption of copepods by clams, whereas

movement of the LSZ in either direction reduces this overlap; (7) Overlap between the copepods *Pseudodiaptomus forbesi* and *Limnoithona tetraspina* increases with a landward X2, intensifying competition for food between these copepods; (8) Overlap between *P. forbesi* and the predatory copepod *Acartiella sinensis* increases with a landward X2, intensifying predation on early stages of *P. forbesi*; and (9) Recruitment of gelatinous plankton to the LSZ is higher when X2 is landward, increasing predation on zooplankton and possibly also delta smelt.

Comments: This project will build on our experience and knowledge gained in the CALFED-funded Food web Study (CALFED Project SCI-05-C107, 2006-2010) and delta smelt feeding studies (CA Bay-Delta Authority, Sea Grant Authorization, U-04-SC-005, 2007-2009).

Additionally, we will use IEP zooplankton and water-quality monitoring data to provide a long-term context for the proposed study. It will further be linked to the following funded and proposed projects: (1) Copepod feeding study (funded, National Science Foundation), (2) Hydrodynamic modeling (Kimmerer, Gross, MacWilliams, IEP), (3) Clam grazing study (Thompson, Gehrts, IEP), (4) Clam physiology study (Stillman, IEP) and (5) Ammonium investigation (Dugdale, Wilkerson, Parker, IEP).

How Will Longfin Smelt Respond to Fall X2 Manipulations? Experimentally Determining Early Life Stage Sensitivity to Salinity

IEP 2010-170

Point Person(s): Randy Baxter (DFG)

Lead IEP Agency: UCD

Questions: What is the optimal salinity range for longfin smelt embryo hatching, and larval rearing success? Does otolith core chemistry reflect maternal or environmental influence? How does the fall X2 position affect spawning and larval rearing habitat for longfin smelt recruitment?

Description:

- **Develop culture techniques and examine salinity tolerance for longfin smelt eggs and larvae:** We will develop fish culture techniques to provide specimens for research and information on egg-incubation-time necessary for assessing the duration of embryo vulnerability to disturbance (e.g., dredging or changing salinity). The fish-culture protocols developed for delta smelt (Baskerville-Bridges et al., 2005) will be modified to accommodate the more euryhaline longfin smelt. We will capture longfin smelt directly or receive live adults from on-going field monitoring studies. Adults will be held close to the salinity and temperature of field-capture locations. Fish are monitored for ripe ova and sperm at transfer to culture tanks and either spawned immediately (through manual expression of gametes) or monitored closely at field-caught salinities until ready to spawn. Fertilization of ova is conducted in fresh water or 2 ppt salt water. Fertilized eggs are totaled (volumetrically estimated) and transferred to egg-incubators (ca. 2-3000/incubator, 10-12 °C). Incubation duration will be documented. At hatch, larvae will be transferred to a larval tank system (70l L black tanks with a re-circulating system at 12 °C) for rearing at 2 or 4 ppt saltwater. Water will be greened (*Nanochloropsis* spp; 10 NTU) daily. Two cultures of live prey (rotifers and *Artemia nauplii*) will be reared and fed to longfin smelt larvae at 2-hour intervals daily. Subsets of larvae will be monitored for growth and development, and preserved to create a developmental series and for subsequent otolith analysis. Salinity tolerance of eggs and larvae will be examined in order to shed light on, whether fall X2 location influences longfin smelt spawning locations, egg incubation and larval rearing locations, and potential vulnerability of adults

and larvae to south delta entrainment. The total number of treatments and replications depends on the number of broodfish collected, subject to our permit limits. Test salinities for 2011 may deviate slightly from what is described below based on recent findings (January - February 2010).

- **Maternal and environmental influence on otoliths core chemistry:** Previous POD funded research examining salinity histories of longfin smelt revealed that many individuals exhibit otolith strontium isotope ratios indicative of (1) high salinities at the core, followed by (2) low salinities shortly outside the core (post-larval to juvenile stages (3) and then a return to high salinity into maturity. This pattern may reflect a maternal transfer of marine-derived strontium to developing offspring (i.e., otolith core) or, could reflect salinities at which eggs were incubated and hatched. We will test the environmental influence mechanism by exposing embryos to different salinity levels and rearing larvae at fresh, 2 ppt and 5 ppt for about 30 days. Experiments will be conducted in conjunction with culture experiment from task 1 (above). Maternal influence will be examined by reconstructing maternal life history of brood fish with otolith strontium isotope ratios prior to first clutch release, and holding adults for a period of time in freshwater prior to a release of a second clutch. Results from these experiments will provide us a powerful tool (validated otolith chemistry) for examining how freshwater management may influence the life-history and recruitment of longfin smelt.

Time Period: October 2010 – November 2011

Resources required:

Cost: \$70,000

PI(s): Joan Lindberg and James Hobbs (UCD)

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Joan Lindberg (UCD)

Term of contract: Through December 2014.

Personnel: Above named investigators and other UCD personnel.

Equipment: None at this time.

Deliverables and dates:

Task 1

- Development of cultured animals for research, preserved developmental series of larvae reared in 2 and 4 ppt salt water, and report on development of longfin smelt culture methodologies.
- Culture of larvae and juveniles reared in saline conditions to evaluate reliability of otolith core tracing developmental salinity exposure of developing longfin.
- Results presented at the at the Delta Science Conference or IEP workshop following termination of this 12-month study.
- Report at termination of project.

Task 2

- 100 otoliths (50 from embryo exposures and 50 for larval salinity tests (April 2011).
- 4 broodstock fish (4) from the salinity trials in Task I (b) (April 2011).
- 100 TNS samples from 2010 (November 2010).
- 100 TNS samples from 2011 (November 2011).
- 100 samples from the FMWT and Bay Study (April 2011).
- Delta Science conference presentation (October 2011).

- Report (November 2011).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? The study will evaluate the salinity tolerance of longfin smelt embryos, larvae, and refine culture techniques for the longfin smelt, a key POD species. The experiments will provide otolith age and microchemistry validation for interpreting data for the current IEP funded POD project (May et al., 2008). The effect of salinity on longfin smelt spawning/rearing habitat has direct implications for the manipulation of fall flow variation and will provide evidence for a mechanistic relationship between fall X2 and longfin smelt recruitment.

Comments: With this study, we hope to develop many of the specific methods for successful longfin culture, thereby elucidating physical parameters that enable or improve holding and rearing success. Comparisons can be drawn between the several life stages of delta smelt and longfin smelt and relative survival and/or growth for fish reared under 1 or more salinity conditions, as follows: holding of adult wild fish, fecundity (or egg-clutch) estimates, spawning and fertilization of eggs, and the rearing of larvae and juveniles. Species segregation, of the 2 smelts in question, and location in the natural habitat appears to depend on salinity, at least for several life stages, and manipulation of the position of X2 in the fall could affect these life stages significantly. This study will provide validation for otolith strontium isotope-salinity relationship developed in May, Israel and Hobbs (2008), as well as provide information for studies regarding the influence of fall X2 and variable delta salinity management strategies.

Remote Sensing Mapping and Monitoring of Microcystis and Turbidity in the Upper San Francisco Estuary

IEP 2010-171

Point Person: Anke Mueller-Solger (DSC)

Lead Agency: UCD

Questions: What are the reflectance properties of the water in the San Francisco Estuary across gradients of suspended solids and toxic algae (*Microcystis*) abundance? How do chlorophyll *a* concentration, total and volatile suspended solids, total and dissolved organic carbon and *Microcystis* abundance vary across the Delta over seasons? What are the distributions of total suspended matter and *Microcystis* blooms and how do these change with time?

Description: We will investigate and develop a procedure to map the spatial distribution of suspended sediments, colored dissolved organic matter and *Microcystis* blooms using NASA Landsat satellite imagery. The Landsat Thematic Mapper (TM) satellite images, which have a 30 x 30 meter pixel resolution, are free and have a weekly return interval over the San Francisco Estuary. Our procedure will have the potential to be used to create spatially contiguous weekly maps that will characterize both the spatial and temporal variation in suspended solids and *Microcystis*.

Our research approach begins with 6 field sampling excursions over July, August and September 2010 concurrent with Landsat TM overpasses to obtain training data for the mapping procedure. We will collect GPS locations and discrete water samples to derive relevant optical properties, along with surface and subsurface water reflectance data. These data will be complemented by monthly discrete monitoring data which will be used to implement and validate the mapping procedure. Using the field data we will investigate 2 approaches to mapping total suspended solids and *Microcystis* (Task 3). The first approach correlates field measurements with reflectance measurements to create a statistical model that can be used to

predict water quality at each pixel of the Landsat image from the reflectance information contained in that pixel. The second approach uses a radiative transfer model (Hydrolight), which models the physical interaction between light and water, also retrieving a map of water quality at each pixel of the Landsat imagery. This model can also estimate in-water light fields which are useful inputs into primary productivity models for the estuary. We will provide a procedural manual for remote sensing mapping of total suspended solids and *Microcystis* and other water quality constituents to the IEP.

Time Period: June 2010 - June 2011

Resources Required:

Cost: \$134,000

PI(s): Susan Ustin and Erin Hestir (UCD), Peggy Lehman (DWR) and Bryan Downing (USGS)

Contract needed/in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Susan Ustin (UCD).

Term of contract: TBD

Personnel: Above named investigators and Jonathan Greenberg, George Scheer, Richard McIlvaine (UCD).

Equipment: None

Deliverables and Dates:

- Excel file containing water quality and *Microcystis* abundance data with GPS location information for model parameterization and calibration (Task 2A).
- Excel file containing water quality and *Microcystis* abundance data with GPS locations for model implementation and calibration (Task 2B).
- Digital maps (raster image or shapefiles) of all water quality variables successfully modeled and mapped with accuracy estimates for each one. Maps will be uploaded to CaSIL hosted on Cal Atlas (<http://www.atlas.ca.gov>).
- Quarterly progress reports (September 2010, January 2011, April 2011).
- 1 draft manuscript for submission to a peer review journal describing the spatial and temporal variation of *Microcystis* in relation to turbidity (July 2011).
- 1 procedural manual written in the form of a how-to training guide for mapping procedures developed during the investigation (July 2011).

How is this activity related to the HSG? Remote sensing maps of turbidity and *Microcystis* can improve existing habitat quality monitoring by providing synoptic (wall-to-wall) measurements of habitat quantity and quality. This information can be used to understand the spatial distribution of delta smelt habitat, and identify discontinuities or changes in habitat availability, as well as identify the spatial distribution of *Microcystis* blooms. We anticipate developing maps of total suspended solids, *Microcystis*, and the in-water light field for Landsat dates in July, August, and September, allowing investigation into the relationship between fall habitat conditions and those occurring earlier in the year as well as the overall trophic conditions of the estuary. Furthermore, our mapping procedures manual will provide the information needed to continue monitoring throughout the entire year. Measuring habitat quality in space and in time is critical to understanding not only how much potential habitat is available to delta smelt in the fall, but how their fall habitat varies spatially and temporally in relation to spring and summer habitat conditions and hydrologic variability.

Comments: None

The Role of Pyrethroid Insecticides in Limiting Prey Availability for Delta Smelt in the North Delta

IEP 2010-172

Point person: Stephanie Fong (CVRWQCB)

Lead IEP Agency: CVRWQCB

Questions: Data emerging over the past few years have shown that pyrethroid insecticides are entering the northwest Delta waters of Cache Slough and nearby areas. These waters are critical habitat for delta smelt, as it is an important spawning area, and particularly so during dry years. This study is designed to determine if pyrethroid pesticides in the Cache Slough region could be reducing populations of copepod prey upon which spawning adult smelt and larval fish depend.

Description: Sampling will be conducted in Cache Slough and surrounding areas during February through June, as the adults gather to spawn and as the larvae/post-larvae remain in the area. Sampling will include water samples for pyrethroid analysis and toxicity testing. Plankton samples will be collected both to quantify availability of copepod prey, and determine if copepod toxicity, as seen in the pyrethroid analytical data and laboratory toxicity tests, is reflected in a concurrent decline in resident populations.

Toxicity testing will be done with the amphipod *Hyaella azteca* for all samples, and with the copepods, *Eurytemora affinis* and *Pseudodiaptomus forbesi*, for selected samples. Should toxicity be found, the pyrethroid chemical analysis will help establish if they are the causative agents, and this linkage further established by TIE tools. Piperonyl butoxide can be used to increase toxicity, if initially caused by pyrethroids, and enzymes engineered to hydrolyze pyrethroids can be used to decrease toxicity due to these compounds.

Time period: December 1, 2010 – November 30, 2011

Resources required:

Cost: \$158,000

PI(s): Donald Weston (UCB), Swee Teh (UCD), Michael Lydy (Southern Illinois University) and Fred Feyrer (USBR)

Contract needed/in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR)

Term of contract: 12 months, once executed.

Personnel: The PIs listed above, assisted by students and staff in their labs.

Equipment: None

Deliverables and dates:

- Quarterly progress reports.
- Draft final report (October 2011).
- Final report (November 2011).
- IEP newsletter article (November 2011).
- Two oral presentations during term of contract.

Comments: IEP-NCEAS has convened a working group, chaired by Dan Schlenk, on pyrethroids in urban runoff and their potential impact on POD species. The Cache Slough area is of specific interest to this group, and a modeling exercise is now in progress to estimate pyrethroid inputs to those waters, and compare this estimate to thresholds of toxicity for aquatic life. The planned study will provide a great deal of data from the Cache Slough area that will be extremely valuable in confirming model predictions and validating the modeling approach as a tool for risk assessment.

Distribution, Concentrations and Fate of Ammonium in the Sacramento River and the Low Salinity Zone: Determination of Phytoplankton Uptake and Bacterial Nitrification Rates
IEP 2010-173

Point person: Larry Brown (USGS)

Lead Agency: SFSU

Questions: Can pelagic nitrification rates be measured (and validated to a degree) in the San Francisco Bay using ^{15}N labeling, the NH_4^+ micro-diffusion technique and mass spectrometry? What are the rates of (a) bacterial/archaeal nitrification and (b) phytoplankton NH_4^+ uptake downstream from Sacramento to Suisun Bay in spring, summer and fall? Does the fate of NH_4^+ (i.e., uptake and nitrification) change with season, salinity and flow?

Description: This research will emphasize quantifying 2 key biological processes influencing river NH_4^+ distribution, bacterial nitrification (= NH_4^+ oxidation) and phytoplankton uptake, and in future years will investigate the degree of river flow-dependence on these processes. This funding will focus on working collaboratively with C. Kendall (USGS, Menlo Park), to develop a protocol for measuring water column nitrification using ^{15}N -labeled NH_4^+ as a tracer. In addition, we will work with estuarine scientists to investigate other tracer-based nitrification methods and determine the most efficient means for determining rates. Towards the end of year 1 it is expected that the protocol will start to be applied to archived river samples that will be incubated and collected in spring and summer 2010 (as part of the CALFED-funded “Two Rivers” project, Dugdale and Mueller-Solger, Lead-PIs) and the Fall 2010 IEP Foodweb (Parker, et al., 2010). C. Kendall will also be involved by collecting samples for natural abundance stable isotope work, for independent estimates of nitrification and phytoplankton N uptake.

Time period: June 2010- May 2011

Resources required:

Cost: \$77,000

PI(s): Dick Dugdale, Alex Parker and Francis Wilkerson (SFSU)

Contract needed / in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR)

Term of contract: N/A

Personnel: Above named investigators, Al Marchi (Research Technician) and a graduate student yet to be determined.

Equipment: Temperature controlled large shaker table / freeze dryer.

Deliverables and dates:

- Presentation at an IEP forum such as EET or the CWT (May 2011).
- If methodology is successful and applied to samples, then preliminary results will be written up as an IEP newsletter article (May 2011).
- With continued support to analyze more samples, it is expected that a peer reviewed journal article will be prepared and the data presented at a national meeting such as Coastal Environmental Rights Foundation (CERF).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? The primary productivity of the river ecosystem is reduced by the change from a NO_3^- based phytoplankton system to an NH_4^+ based phytoplankton and bacterial system, forced by the input of NH_4^+ from the SRWTP. Recovery of the system to NO_3^- based phytoplankton productivity and recovered primary production requires the reduction of ambient NH_4^+ concentrations to $\sim 4 \mu\text{M}$ to initiate blooms, and to $1 \mu\text{M}$ to substantially relieve NH_4^+ inhibition of

NO_3^- uptake. Consequently, the conversion of NH_4^+ to NO_3^- (not available to the phytoplankton) represents a potential loss of primary productivity to the section of the river with elevated NH_4^+ concentrations. The most desirable form of inorganic nitrogen, NO_3^- , is exported downstream and out of the area of concern for the POD. The rate of nitrification determines the point in the river at which NH_4^+ concentration is reduced to the critical point for bloom formation (4 μM) and then to 1 μM for access to the now elevated pool of NO_3^-

Comments: Synergies with past and present research programs:

- Parker/ Dugdale: State Water Contractors (SWC) Project - ^{15}N uptake by phytoplankton and NH_4^+ distributions available.
- Dugdale/A. Mueller-Solger CALFED “Two Rivers” CALFED Project.
- Kendall: SWC Project and IEP Project, parallel measurements of stable isotopic $\delta^{15}\text{N}$ (natural abundance).
- Patin: Masters thesis research, SFSU, “Abundance of Ammonia oxidizing Archaea in SF Bay” (Ashby, Advisor).
- Melloy: Masters thesis research, SFSU, “Use of bacterial inhibitors to evaluate the role of nitrifying Archaea” (de la Torre, Advisor).

The influence of elevated ammonium (NH_4) on phytoplankton physiology in the San Francisco Estuary Delta during fall: exploring differences in nutrients and phytoplankton in the Sacramento and San Joaquin Rivers and how variation in irradiance via changing river flow, modulates NH_4 effects.

IEP 2010-174

Point person: Anke Mueller-Solger (DSC)

Lead Agency: SFSU

Questions: What are the rates of primary production and phytoplankton NO_3 and NH_4 uptake in the Sacramento and San Joaquin rivers during the fall period? What role does DIN composition and concentration play in modulating the above phytoplankton rates and phytoplankton species composition? How does river flow affect nutrient distribution and phytoplankton rates? Does the conceptual model of NH_4 suppression of phytoplankton NO_3 uptake and primary production hold under low-light conditions?

Description: Emerging evidence suggests that shifts in nutrient composition from NO_3 to NH_4 as a result of the Clean Water Act and population increases (Jassby 2008) likely play a role in the long-term phytoplankton decline that has been observed in the northern SFE (Dugdale et al. 2007). NH_4 likely affects SFE phytoplankton species composition and also appears to influence primary production rates by modulating phytoplankton physiology, particularly that of fast growing and nutritious diatoms (Dugdale et al. 2007). Elevated NH_4 concentrations ($>4 \mu\text{mol L}^{-1}$) appear to inhibit phytoplankton NO_3 uptake. Only during brief periods NH_4 is reduced to low concentrations and phytoplankton NO_3 uptake increases rapidly (termed “shift-up”, Dugdale et al. 1990) resulting in increased phytoplankton growth. The major source of dissolved inorganic nitrogen (DIN, i.e. NH_4 and NO_3) in the northern SFE is from agriculture and municipal wastewater treatment plants (Hager and Schemel, 1996; Jassby 2008) with DIN in the Sacramento River coming primarily in the form of NH_4 . In contrast, due to differences in municipal wastewater processing, discharge in the San Joaquin River is largely in the form of NO_3 . The contrast between the Sacramento River (NH_4 -dominated) and the San Joaquin River (NO_3 -dominated) provides a natural experiment to test primary production and phytoplankton community composition responses as a result of NH_4 and NO_3 loading.

One outstanding question is whether the NH₄ inhibition effect or the NO₃ shift-up that follows NH₄ exhaustion occurs at low irradiances characteristic of the natural system. Research in marine settings has demonstrated an irradiance response for phytoplankton DIN uptake, including a differential response for phytoplankton NH₄ and NO₃ uptake (McCarthy et al. 1996; Kudela et al. 1997; Parker et al. in review). Phytoplankton DIN versus irradiance relationships are not clear for the SFE or estuarine environments, generally.

Time period: July 2010 - July 2011

Resources required:

Cost: \$114,000

PI(s): Alex Parker, Francis Wilkerson and Richard Dugdale (SFSU).

Contract needed / in place: In place.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Richard Dugdale (SFSU).

Term of contract: 1 year.

Personnel: Al Marchi and Erica Kress (SFSU) and above named investigators.

Equipment: None

Deliverables and dates:

- IEP final report (July 2011).
- Poster or presentation at the 2011 IEP annual workshop.
- IEP Newsletter article (spring 2011).
- Manuscript draft for submission to a peer-reviewed journal (July 2011).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? This project aims to better constrain the role that NH₄ plays in modulating primary production and ultimately, the bottom-up controls on the food web of POD species. Historically nutrients have not been considered important in regulating primary production as the SFE was identified as a light-limited estuary. Recently, there has been considerable renewed interest in the role of nutrients (particularly NH₄, but also PO₄) in shaping the phytoplankton community composition and primary production rates in the Central Delta. Long-term shifts in phytoplankton taxa (including the proliferation of harmful algal bloom species such as *Microcystis*) in the northern San Francisco Bay as well as the decades long decline in primary production rates are likely a function of multiple ecosystem drivers. Changes in wastewater treatment practices may be a potentially beneficial management strategy in support of improved conditions for POD species.

Comments:

- We will work with Kimmerer et al. (funded) for cruise planning / preparation and in synthesis of results.
- Under this budget we will collect, incubate and archive all ¹⁵N nitrification rate measurements in fall 2010 for analysis under the Dugdale et al nitrification project (funded).
- This dataset will provide greater temporal coverage (by completing transects in fall) for the “Two Rivers” project underway by Dugdale and Anke Mueller-Solger (DSC).
- These projects will extend the dataset currently being collected by C. Foe (CVRWQCB) on delta-wide DIN distribution.
- We will extend the dataset collected during the SWC transects, including data collection during the fall and in the San Joaquin River complimenting the existing Dugdale and Parker State Water Contractors Project.

Evaluation of the Effect of Seasonal Variations in Flow on the Spatial and Temporal Variations of Nutrients, Organic Matter and Phytoplankton in the Sacramento River and Northern San Francisco Bay

IEP 2010-175

Point person: Anke Mueller-Solger (DSC)

Lead Agency: USGS

Questions: How does the location of X2, especially in the fall, affect constituents important to the base of the smelt and other foodwebs in the northern San Francisco Bay, Delta, and lower Sacramento River? What areas in the Delta act as nutrient sources and sinks, and what processes are involved? Does the Yolo/Cache/Liberty area act as a sink or a source for ammonia and other nutrients? Under what conditions do nutrient sources from the Yolo/Cache/Liberty area support the food web in the LSZ? Are there regional, river reach-dependent, and temporal variations in nitrification rate or other rates of nutrient degradation? If so, what is the dominant cause? Does algal growth rate appear to depend on the relative proportions of ammonia and nitrate? What are the dominant processes affecting the downstream location (denoted in river miles or RM) where NH_4^+ concentrations approach levels where NH_4^+ no longer appears to inhibit NO_3^- uptake?

Description: The purpose of this data synthesis project is to investigate how seasonal and spatial changes in freshwater flow and the relative amounts of water from different sources (e.g., the Sacramento River, the San Joaquin River, the Yolo Complex, SRWTP, et – as estimated using DSM2 and RMA) affect the temporal and spatial variations in the sources, transport, and sinks of NH_4^+ , NO_3^- , chlorophyll, and organic matter in the Sacramento River and lower Delta. In specific, we propose to take advantage of several large sets of existing chemical and isotopic data to ask questions about how the location of the 2% salt front (X2), especially in the fall, relates to constituents important to the base of the smelt and other foodwebs in the northern San Francisco Bay, Delta, and lower Sacramento River.

A main focus of this project will be the calculation of nitrification rates for different river reaches and seasons, and identifying how seasonal changes in flow, effluent levels and composition, proportions of water from different sources, temperature, etc. affect nitrification rates. Available chemistry from the recent Foe and Dugdale transects – and our nitrate isotope data – show that nitrification is the dominant N cycling process in the Sacramento River below SRWTP and is considerably more important for reducing NH_4^+ levels to below the inhibition threshold than algal uptake.

Time period: July 2010 - July 2011

Resources required:

Cost: \$42,000

PI(s): Carol Kendall (USGS)

Contract needed / in place: In progress.

Contract manager(s): Erwin Van Nieuwenhuyse (USBR) and Roger Fujii (USGS).

Term of contract: To be determined.

Personnel: Megan Young, Steve Silva, and Tamara Kraus (USGS); Marianne Guerin (RMA); Chris Foe (CVRWQCB); Alex Parker, Dick Dugdale, and Frances Wilkerson (SFSU)

Equipment: None

Deliverables and dates:

- Quarterly progress reports with invoices.
- An update presentation at an IEP forum (such as EET or the POD CWT), and a talk or poster at a CALFED, American Geophysical Union (AGU), or other national meeting during the first year.
- At the completion of the project, we will provide an electronic copy of our entire database, and 1 or more USGS-approved and journal-intended articles.

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? This project addresses 2 of the priority research topics: habitat effects and food web effects on delta smelt population dynamics. By “habitat effects”, we mean the quantification of abiotic variables such as net flow and residence time within the study area or in sub-regions of that area, and the effect of these variables on nutrient sources and concentrations within the LSZ and elsewhere within the study area. The project focus on “food web effects” is at the level of nutrients and primary productivity. This proposed project also supports 3 of the 4 key research priorities from the 12 framework-panel-identified topics listed as part of the “CALFED Science Program Issue Summary” on the role of ammonium (http://science.calwater.ca.gov/pdf/publications/Ammonium_one-pager_070109.pdf):

- Use models to explore the transport of ammonia/ammonium within the Delta and effects on the amount of, type of, and growth of algae, the base of the Delta food web.
- Determine the main sources of ammonia/ammonium (and phosphorus) and trace the fate of these substances within the Delta.
- Explore possible links between specific types of algae and aquatic plants and the amount of ammonium in the water.

Comments:

Determination of Influences of Water Quality and Submerged Aquatic Vegetation on Largemouth Bass Distribution, Abundance, Diet Composition and Predation on Delta smelt in the Sacramento-San Joaquin Delta

IEP 2010-176

Point person: Ted Sommer (DWR)

Lead Agency: UCD

Questions: How do abiotic and biotic factors influence largemouth bass distribution and abundance in the Delta, and their impacts on delta smelt and other pelagic species? What are the abundance and the diet composition of largemouth bass and other potential predators in areas where delta smelt are known to be present? What is the relationship between biomass density and species composition of submerged vegetation beds and the invertebrate community assemblage and biomass? Does variation in the invertebrate community within the submerged vegetation explain variability in the abundance and/or diet composition of juvenile largemouth bass captured at the same locations?

Description: This project builds on an IEP-supported study (currently underway) that is investigating abiotic and biotic influences on the spatial distribution, abundance, size distribution, movements, and diet of largemouth bass (LMB). The proposed project will allow the current sampling program to carry out 3 more months of the current study’s sampling protocol to complete 2 full years of field surveys at 33 sites spread throughout the Delta. This field effort will complete 2 full years of bimonthly data collection, leaving approximately 6 months for data compilation, analyses, and preparation of manuscripts. We will produce a

spatially explicit dataset that integrates aquatic macrophytes, abundance of each fish species, diet composition of LMB, and water quality parameters. Analyses will address how seasonal and inter-annual variation in environmental conditions, including salinity, fall X2, and SAV influence LMB distribution, density, and diet, as well as the general fish assemblage in the littoral zone. In addition to extending the sampling program already underway, this project will also examine the abundance and diet composition of LMB and other predators in specific locations where delta smelt are known to occur. A second new goal is to examine how the increased biomass of SAV—largely due to the proliferation of the invasive Brazilian waterweed—may contribute to the success of the largely non-native littoral fish assemblage, particularly LMB, by boosting secondary (invertebrate) production that in turn provides a prey base for juvenile fish.

Time period: July 2010 – June 2011

Resources required:

Cost: \$173,000

PI(s): Andrew Sih, and Peter Moyle (UCD); Louise Conrad (DWR)

Contract needed / in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR)

Term of contract: To be determined.

Personnel: In addition to PIs, other UCD staff include Patrick Crain, Matthew Young, Kelly Smith, a second junior specialist and an, undergraduate student researcher.

Equipment: No new equipment needed at this time.

Deliverables and dates:

Task 1: Influence of biotic and abiotic factors on LMB abundance, distribution, and diet

- Final report to IEP (June 2011).
- 2 manuscripts for publication in peer-reviewed journals (June 2011).

Task 2: Abundance and diet composition of largemouth bass and other predators co-occurring with delta smelt

- Final report to IEP (June 2011).
- IEP newsletter and/or presentation at relevant IEP conference or workshop, as requested (spring 2011).

Task 3: Influence of the SAV species and biomass on invertebrate community composition and biomass

- 1 manuscript for publication (June 2011).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? LMB are voracious piscivores that have been recognized as keystone species because of their far-reaching impacts on the resident food web in systems where they have been introduced. LMB are introduced in the Delta and have demonstrated a dramatic population increase in recent decades. Thus, they may be important predators of native fishes, including species that are part of the POD. Major goals of our proposed work are thus to quantify potential impacts of LMB on delta smelt, threadfin shad, and juvenile striped bass and to elucidate factors that explain the recent success of LMB in the Delta. Specifically, we will complete a 2-year field effort to examine how biotic factors (e.g., aquatic macrophytes) and abiotic variables (e.g., temperature, salinity, and fall flow variation) influence the distribution and abundance of LMB throughout the Delta. The spatial distribution of LMB with respect to environmental factors such as annual variation in Delta outflow will aid in predicting the degree to which delta smelt are likely to overlap with LMB. However, while

spatial overlap between these 2 species suggests the possibility of LMB predation of delta smelt, LMB foraging behavior may limit this possibility. Thus, we will also examine diet contents of LMB in areas where the presence of delta smelt has been established by recent native fish surveys conducted by DFG or other resource agencies to better assess the impact of LMB on delta smelt.

Comments: This project builds on IEP-funded research that is currently underway, under the POD work team's 2008-2010 contract with UCD, entitled, "Impacts of Largemouth Bass on the Delta Ecosystem."

Metabolic Responses to Variable Salinity Environments in Field-Acclimatized Corbula amurensis

IEP 2010-177

Point person: Larry Brown (USGS)

Lead Agency: SFSU

Questions: How much metabolic variation do we see in *Corbula* acclimatized to different salinities across sites and seasons? How are *Corbula* partitioning energy? How does variation in water chemistry and planktonic assemblage alter the metabolic physiology of *Corbula*?

Description: We propose to continue our studies on *Corbula amurensis* (*Corbula*) metabolic physiology in field-acclimatized specimens collected at monthly intervals from sites representing extremes in salinity variability over the distribution range of *Corbula* in the northern San Francisco Estuary. We will collect clams at 3 sites representing nearly fresh, medium and high salinity fluctuation to address the above questions. For question 1, we will measure metabolic rates and feeding rates of clams within 12 hours of collection. For question 2, we will freeze clams upon collection and characterize tissues and/or biochemical activities associated with growth, reproduction, osmoregulation, metabolism, and energy stores. For question 3, we will measure salinity, temperature, pH and turbidity and will characterize the energetic content of large and small size class plankton assemblages.

In sum, these measurements will tell us at what rate and for what purposes the clams are using energy in the natural habitat across gradients in salinity and temperature exposure, and what the energy reserves of the clams are like depending on when they settle. Clams that have lower energy reserves or that are using greater amounts of energy per unit time will likely have a stronger impact on food-webs as they must graze more to obtain their energy. Thus this information will result in improved food web modeling of clam impacts on pelagic organisms' food supply.

Time period: August 2010 to December 2011, with a hiatus from December 2010 to February 2011, when sampling of *Corbula amurensis* is not feasible.

Resources required:

Cost: \$137,000

PI(s): Jonathon Stillman (SFSU)

Contract needed / in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR)

Term of contract: To be determined.

Personnel: Nathan Miller (SFSU), Postdoctoral Fellow

Equipment: No new equipment needed at this time.

Deliverables and dates:

- Written IEP reports at end of project (January 2012).

- Manuscripts for publication in peer reviewed journals (e.g., Marine Ecology Progress Series (MEPS), Estuaries and Coasts) at end of project (January 2012).
- Poster and/or oral presentations at CERF in Nov 2011, CAERS in spring 2011, IEP/EET in spring 2011 and the Society for Integrative and Comparative Biology (SICB) in January 2012.

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? The invasive clam *Corbula amurensis* is thought to exert a strong bottom-up influence on the pelagic food web by its filtration of phyto- and zooplankton to supply its metabolic demands. This project seeks to characterize the metabolic physiology of *Corbula amurensis* in locations representing the extremes of their salinity distribution ranges in the northern San Francisco Estuary. Understanding the rate at which clams require and use energy is a critical component of their relative interaction strengths in food-web models of the area. Assessment of the variation in metabolic physiology across seasonal variation in freshwater input and at sites that vary across degree of tidally-influenced salinity fluctuations is important in order to parameterize how these clams are likely to respond to natural and anthropogenic variation in salinity.

Comments: This project produce results that should be integrated with results from proposed projects by Thompson & Gehrts, Kimmerer & Sullivan, Gross, MacWilliams & Kimmerer, and Dugdale, Wilkerson & Parker.

Bivalve Effects on the Food Web Supporting Delta Smelt 2010-178

Point person: Larry Brown (USGS)

Lead Agency: USGS

Questions: How do *Corbula amurensis* and *Corbicula fluminea* affect the food web supporting delta smelt, and how are they affected by flow variability? How do the grazing rates of *Corbicula fluminea* and *Corbula amurensis* vary with longitudinal location in the Delta and the LSZ? How do the grazing rates of *Corbicula fluminea* and *Corbula amurensis* vary with water depth in the Delta and the LSZ? How do the population dynamics of *Corbicula fluminea* and *Corbula amurensis* (recruitment, growth, and mortality) vary as a function of X2 position in fall? How do antecedent fall salinity conditions in the LSZ affect bivalve population biomass and grazing rates in the following spring?

Description: Our task is to establish the distribution, population dynamics, and grazing rate of *Corbula* and *Corbicula* within the LSZ and within the tidal excursion of the LSZ. We will augment current field and laboratory procedures to avoid duplication of current programs, and use DWR EMP monitoring data and USGS benthic bivalve data to provide a long-term context to the study. DWR currently does a spatially intensive benthic Generalized Random Tessellation Stratified (GRTS) sampling study in the spring and fall that will be augmented with a few stations in the region of the strongest gradients in bivalve grazing (within a tidal excursion of the LSZ) and in some shallow water areas to insure that we can adequately model benthic grazing in these areas with some accuracy. Previously collected samples from the fall 1999 “experiment” and the spring of 1999 will be processed in addition to samples collected in 2010. The tasks are listed as follows:

- Task 1: Collect 22 additional stations during the GRTS spring and fall 2010 sampling (USGS and DWR).

- Task 2: Sort 22 additional samples for fall 1999 (we took part in the fall 1999 “experiment”), spring and fall 2010 (USGS).
- Task 3: Measure bivalves from 1999 GRTS samples and from 2010 GRTS samples (USGS).
- Task 4: Convert bivalve measurements to biomass and grazing rate (USGS and DWR).
- Task 5: Report biomass and grazing rate numbers as they become available to IEP and fellow PIs in fall study (USGS and DWR). Work with numerical modelers to establish grazing rates within grids (USGS).

Time period: July 2010 – July 2011

Resources required:

Cost: \$89,000

PI(s): Janet Thompson (USGS and Karen Gehrts (DWR)

Contract needed / in place: In process,

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Janet Thompson (USGS)

Term of contract: To be determined.

Personnel: Above named investigators.

Equipment: None at this time.

Deliverables and dates:

- Grazing rates available for 1999 (January 2011).
- Grazing rates available for spring 2010 (March 2011).
- Written IEP reports and fall 2010 grazing rates available at end of project (August 2011).
- Presentation at IEP Annual Workshop (2011 and 2012).
- Manuscripts for publication in peer reviewed journal (July 2012).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? The conceptual model of the delta smelt foodweb shows (1) confirmed links between clams and phytoplankton, larval copepods, bacteria and ciliates (current work by V. Greene with Kimmerer); and (2) potential links with a broader range of zooplankton larvae and ciliates and other forms of POC and DOC. Phytoplankton production is severely light limited in this system and thus positive net production is confined to shallow areas where accelerated vertical mixing rates expose phytoplankton cells to more light than in the channel. However, the phytoplankton cells, while being more rapidly vertically mixed in the shallow water, are also being exposed to more clam grazing. We have 2 species of clams in this system that overlap in the salinity range of 0-10: adult *Corbicula* can live in salinities of 10 and we have observed adult *Corbula* in the tidal river in areas where bottom salinities are at or near zero for large portions of each day. Juveniles of both species can withstand salinities of 2. The species have different pumping rates and our knowledge of their ability to filter zooplankton and ciliates differs. Therefore the distribution and biomass of these clams will determine the loss rate of phytoplankton and other portions of the lower foodweb due to benthic grazing.

Why Fall?: Because *Corbicula* and *Corbula* distributions are a function of salinity, and their larval recruitment periods occur in spring and fall, their effect on the foodweb is a function of the salinity distribution in spring and fall. Once larvae settle, they are more able to withstand a broader range of salinity. A salinity incursion into the Delta in fall followed by a dry or below normal outflow winter may allow *Corbula* to not only settle within the X2 area but also to survive at that location into the following spring and fall even if the salinity declines. Therefore,

we need to determine the distribution and grazing rate of *Corbula* and *Corbicula* in fall and in the following spring.

Comments: This project produce results that should be integrated with results from proposed projects by Stillman, Kimmerer & Sullivan, Gross, MacWilliams & Kimmerer, and Dugdale, Wilkerson & Parker.

Determination of the Causes of Seasonal and Spatial Variation in NH₄⁺ Sources, Sinks and Contributions to Algal Productivity in the Sacramento River, Delta and northern San Francisco Bay Using a Multi-isotope Approach

IEP 2010-179

Point person: Anke Mueller-Solger (DSC)

Lead Agency: USGS

Questions: Can we quantify the seasonal contributions of NH₄⁺ from SRWTP, tributaries, and other sources to critical habitats? Does the phytoplankton species composition in the downstream of SRWTP vary with the concentration of NH₄⁺ or any other constituent in the effluent or tributaries identified by isotope analysis? How do nitrification rates (the main biogeochemical process responsible for lowering NH₄⁺ levels) vary seasonally and in different reaches of the Sacramento River and upper estuary? Which areas act as nutrient sources and sinks? Can we quantify the NH₄⁺ from WWTPs versus agricultural drains to different sites and seasons? How do NH₄⁺, NO₃⁻, and organic matter concentrations vary under different hydrologic conditions?

Description: This project will quantify temporal and spatial variations in the sources, transport, and sinks of NH₄⁺, NO₃⁻, and organic matter at 21 sites in the Sacramento River, San Joaquin River, and Delta sampled as part of the 1-year NH₄⁺ Monitoring Program established and conducted by Chris Foe. We have coordinated with Foe's monitoring program to obtain and archive splits of samples for isotopic and algal speciation analyses. An ongoing SWC-funded pilot study has shown that nutrients and organic matter from different sources, or affected by different biogeochemical processes, have distinctive isotopic compositions that are diagnostic of the different sources and processes.

This project consists of 3 tasks, the first 2 aimed at analyzing the archived samples, and the third to integrate the new data and prepare reports. Task 1 will analyze these samples for δ¹⁵N of NH₄⁺, δ¹⁵N and δ¹⁸O of NO₃⁻, δ¹³C of DOC, δ¹⁸O and δ²H of water, and δ¹³C, δ¹⁵N, δ³⁴S, and C:N of seston. Task 2 (funded presumably by the DWR to Peggy Lehman) will analyze archived lugol-preserved samples for algal speciation. Task 3 (as yet unfunded, for year 2 is to (A) combine the chemical, isotope, algal, and hydrological data to quantify temporal and spatial changes in (1) the relative contributions from different sources of NH₄⁺, (2) the nitrification rate, (3) the uptake of NH₄⁺ and NO₃⁻ by different algal species (and bacteria), and (4) how these critical sources and biogeochemical processes are affected by hydrology and management-driven changes in net flow and water source percentages at different sites and dates; and (B) prepare presentations and reports.

Time period: July 2010 to July 2011 (for tasks 1 and 2; task 3 (currently unfunded) will take about an additional 6 months).

Resources required:

Cost: \$242,000 (task 1 only)

PI(s): Carol Kendall (USGS)

Contract needed / in place: In process.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Roger Fujii (USGS).

Term of contract: To be determined.

Personnel: Megan Young, Steve Silva, and Tamara Kraus (USGS); Marianne Guerin (RMA); Peggy Lehman (DWR); Chris Foe (CVRWQCB)

Equipment: None planned, unless something critical breaks.

Deliverables and dates:

- Quarterly progress reports for task 1 with invoices. Excel data files at project completion.
- Presentations of preliminary data and findings at EET, IEP, CWT, and other meetings during the first year; memo (to the POD CWT?) describing policy implications of project findings, and recommendations for future monitoring efforts at the end of the first year.
- Excel data files containing the phytoplankton data and a report containing the result of the statistical analyses during the first year (task 2).
- Quarterly progress reports for task 3 with invoices during the second year (funding permitting).
- 1 or more USGS-approved journal paper(s) during the second year (funding permitting) for task 3.

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? This project addresses 2 of the priority research topics: habitat effects and food web effects on delta smelt population dynamics. By “habitat effects”, we mean the quantification of abiotic variables such as net flow and residence time within the study area or in sub-regions of that area, and the effect of these variables on nutrient sources and concentrations within the LSZ and elsewhere within the study area (Research Question 1C). The project focus on “food web effects” is at the level of nutrients and primary productivity (Research Question 2C). This proposed project also supports 3 of the 4 key research priorities from the 12 framework-panel-identified topics listed as part of the “CALFED Science Program Issue Summary” on the role of ammonium

(http://science.calwater.ca.gov/pdf/publications/Ammonium_one-pager_070109.pdf):

- Use models to explore the transport of ammonia/ammonium within the Delta and effects on the amount of, type of, and growth of algae, the base of the Delta food web.
- Determine the main sources of ammonia/ammonium (and phosphorus) and trace the fate of these substances within the Delta.
- Explore possible links between specific types of algae and aquatic plants and the amount of ammonium in the water.

Hydrodynamic and Particle Tracking Modeling of Delta Smelt Habitat and Prey

2010-180

Point person: John Netto (USFWS)

Lead Agency: USFWS

Questions: How does the habitat area and volume for delta smelt and other fishes vary with freshwater flow? What patterns of vertical swimming by planktonic organisms in the LSZ result in tidal patterns of vertical distribution similar to those observed? How does the observed tidal vertical migration of planktonic organisms influence their retention and transport in and near the LSZ? How does this retention and transport vary with flow conditions?

Description: This project will use existing modeling tools and existing data to accomplish 2 tasks. The first task is to model the variability of physical habitat with X2 for key fish species

including delta smelt. We will build on our previous efforts, improving the scope and resolution of habitat modeling by using the Unstructured Tidal, Residual, Intertidal and Mudflat (UnTRIM) San Francisco Bay-Delta Model to describe how physical habitat determined from fish distributions varies with freshwater flow.

In the second task, we will investigate the population dynamics of calanoid copepods, the most important food for delta smelt in summer to fall. Specifically, we will use the UnTRIM Bay-Delta model and the Flexible Integration of Staggered-grid Hydrodynamics (FISH) particle tracking model (PTM) to investigate retention processes that affect abundance of planktonic organisms, focusing on the calanoid copepod *Pseudodiaptomus forbesi*. The modeling will be designed to investigate the consequences of tidal vertical migration of copepods in the LSZ for retention of the copepods within that zone. Data from the intensive Entrapment Zone Studies of 1994-1996 will be used to define the range of migratory patterns, and retention under alternative migratory patterns and freshwater flows will be investigated.

Time period: July 2010 – June 2011

Resources Required:

Cost: \$339,000 in 2010.

PI(s): Wim Kimmerer (SFSU)

Co-PI(s): Edward Gross (Bay Modeling) and Michael MacWilliams (River Modeling)

Contract needed / in place: In progress.

Contract manager(s): Erwin Van Nieuwenhuyse (USBR)

Term of contract: To be determined.

Personnel: Rusty Holleman and Sandy Chang (SFSU)

Equipment: None

Deliverables and dates:

- EET presentations (fall 2010 and spring 2011).
- CWEMF presentation on model calibration (February 2011).
- Paper on model calibration (February 2011).
- Paper on habitat analysis (June 2011).
- Paper on copepod retention (June 2011).

What is the relation to POD investigations and effects of fall flow variations on delta smelt?

This project addresses 2 aspects of the POD conceptual model: physical habitat and food supply. Changes in the physical shape or size of the low-salinity habitat have been hypothesized to cause a reduction in abundance of delta smelt or their food when X2 is high or landward (USFWS HSG, 2009). Task 1 of this project will determine how physical habitat of delta smelt, as determined by salinity, responds to changes in flow/X2. This information can be determined only through modeling, and it requires a detailed, state-of-the-art hydrodynamic model.

In Task 2, we will address a particular question on the food supply of delta smelt. A related project will investigate the population processes of the principal food organism, the copepod *Pseudodiaptomus forbesi* and ultimately how they relate to flow conditions. Population dynamics results from reproduction, growth, and development, and mortality. Losses from the population due to transport in the water contribute to total mortality. However, this is complicated by the fact that the copepods are capable of migrating vertically in synchrony with the tides, thereby potentially eliminating the tendency for transport away from the population center. Assessing the effects of this migration on retention is therefore essential for estimating overall mortality, and therefore understanding how population dynamics varies with flow

conditions. This also requires an accurate 3-dimensional hydrodynamic model and a 3-dimensional particle tracking model.

Comments: Task 1 will build on our previous work developing habitat models based on distributions of fish and using the TRIM 3-D model. Task 2 will build on earlier work (Kimmerer, Bennett, and Burau) investigating tidal migrations in the LSZ.

Longfin Smelt Bioenergetics

IEP 2010-181

Point Person: Randy Baxter (DFG)

Lead Agency: UCD

Questions: What are the food and oxygen consumption rates for juvenile and sub-adult longfin smelt? How do these metrics affect longfin smelt growth as assessed by a bioenergetics model? What is the importance of growth in the longfin smelt IBM as related to population dynamics?

Description: In fish bioenergetics studies, the most common parameters investigated in laboratory experiments are those related to consumption and metabolism. To date, no such laboratory experiments have been carried out for longfin smelt, making it difficult to apply bioenergetics data to the longfin smelt IBM. We will measure maximum food consumption, and active and resting oxygen consumption rates at 4 temperatures on hatchery-raised longfin smelt juveniles and sub-adults. These data will be incorporated into a bioenergetics model to determine growth rates over a range of different temperatures.

IBM has been previously developed for longfin smelt to describe the spatial and temporal heterogeneity of parameters affecting fecundity and mortality at all life stages. This IBM is useful for exploring the relative significance of specific factors influencing population numbers of longfin smelt in the Bay-Delta. Specific processes relevant to the IBM include bioenergetics (growth), habitat preference, environmental tolerance, feeding habits, and the age or life-stage dependence of these factors. We will incorporate bioenergetics parameters derived from this study into the IBM to fill in existing key data gaps.

Time Period: May 2010 - April 2011

Resources Required:

Cost: \$128,422

PI(s): Frank Loge, Joseph J. Cech Jr., Nann Fangue and Joan Lindberg (UCD)

Contract needed/in place: In progress.

Contract manager(s): Erwin Van Nieuwenhuyse (USBR) and Frank Loge (UCD).

Term of contract: To be determined.

Personnel: Kai Eder, Erik Loboschefskey, Cincin Young and Dennis Cocherell (UCD)

Equipment: Loligo Respirometer and attendant data acquisition software.

Deliverables and Dates:

- Determine longfin smelt bioenergetic parameters of maximum food consumption and resting and active oxygen consumption rates (November 2010).
- Construct longfin smelt bioenergetics model (February 2011).
- Implement longfin smelt bioenergetics model into the IBM (March 2011).
- Final report (April 2011).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? Temperature changes brought about by

variations in flow influence the food and oxygen consumption rates of Delta fish species.

Growth of longfin smelt, assessed by the bioenergetics model, may thus be affected by variations

in fall flows. Because growth is a major component of the IBM, flow variation is likely to affect longfin smelt population dynamics. These findings will advance the understanding of flow variation effects on Delta fish populations, including delta smelt.

Comments: This work will require cultured longfin smelt obtained from the FCCL located in Byron, California. Development of the IBM for longfin smelt is funded under a separate IEP funded study.

Delta Smelt Acoustic Tag Development

IEP 2010-182

Point person: Ted Sommer (DWR)

Lead Agency: UCD

Questions: What are the overall physical constraints for a Juvenile Salmon Acoustic Telemetry system (JSATs)-compatible acoustic tag suitable for use in delta smelt?

Description: The overall long-term goal of this project is to work towards developing an acoustic telemetry system for delta smelt and to use this system to monitor their location and behavior within the San Francisco Bay-Delta system. For this specific study, we will determine acoustic transmitter constraints of size, shape, and mass needed for successful surgical implantation into delta smelt. The physical dimensions of a tag should neither interfere nor obstruct proper functioning of internal organs and tissues, while minimizing the likelihood of tag expulsion. Additionally, normal swimming behavior and buoyancy control are behaviors that must be preserved through understanding of tag size constraints within the context of sex and life stage of delta smelt. To this end, we will inject smelt internal body cavities with moldable latex to determine the maximum size of a potential tag within a given smelt size class. We will then construct inoperable “dummy” tags comprising 3 size ranges of measured body cavities (upper, middle, and lower ranges) while keeping the tags within accepted tag body burden limits. The tags will be surgically implanted (or injected if small enough) into various smelt at different stages of maturity and their survival will be assessed over 30 days. Short-term tagging effects such as swimming and moving will be assessed during the first 24 hours, while all fish (30-day mortalities as well as survivors) will be examined by a histopathologist for tag-related tissue and organ damage, and assessments of incision and suture will be performed. Based on outcomes from these various analyses an optimal size, shape, and mass of acoustic transmitter will be determined for delta smelt.

Time period: October 2010 - June 2011

Resources required:

Cost: \$177,544

PI(s): Frank Loge and Raul Piedrahita (UCD)

Contract needed / in place: In progress.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Raul Piedrahita (UCD).

Term of contract: To be determined.

Personnel: Kai Eder, Donald Thompson, Joseph Groff and graduate student research assistant (UCD).

Equipment: A 2HP heat pump for a recirculating water system.

Deliverables and dates:

- Measurement of smelt internal body cavity (January 2011).
- Construction of “dummy” tags used for implantation (March 2011).

- Tag implantation and evaluation of biological effects (June 2011).
- Final report (June 2011).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? Many, if not all, of the ecological questions surrounding the recovery of delta smelt require knowledge of where these animals are located seasonally, where they spawn, and corridors of movement. To date, little is known on the spatial and temporal distribution of these animals. Through development and subsequent application of miniaturized acoustic tags to delta smelt, fish movement and behavior can be studied.

Comments: None

Refinement and Application of Novel Molecular and Biochemical Biomarkers to Determine Sublethal Contaminant Exposure and Effects in Archived Delta Smelt Samples.

IEP 2010-183

Point person: Anke Mueller-Solger (DSC)

Lead Agency: UCD

Questions: Can novel biomarker tools be used to monitor and assess sublethal impairment of important life history parameters of delta smelt exposed to ambient Delta water and specific contaminants? Can these tools be utilized for toxicant identification?

What is the temporal and spatial distribution of sublethal impairments and contaminant exposure of delta smelt in the Delta? How do results obtained with novel biomarkers compare to acute toxicity patterns found for delta smelt in the Delta?

Description: The proposed study builds on investigations performed in 2006-2009 to evaluate chemical-related aquatic toxicity of Delta water to early life stages of delta smelt. Results from these studies identified specific areas of concern in the Delta, particularly the Sacramento River at Hood, Cache Slough near Lindsey Slough, and the San Joaquin River at Rough and Ready Island. Current-use pesticides along with contaminants associated with municipal wastewater treatment effluents were labeled as potential toxicants. However, no toxicity identification tools are currently available for this species. Genomic biomarkers measure the functional response of organisms to contaminant exposures. Different chemicals will elicit a different response that is dependent on their mechanisms of action. To date, we have been successful in identifying the mechanisms of action of contaminants of concern through the application of molecular biomarkers and links have been established between gene expression and responses at higher levels of organization.

The effects of environmental contaminants as a factor in the observed decline of delta smelt will be investigated through a single experimental task: refining and applying novel biomarker tools to determine sublethal contaminant exposure and effects in archived delta smelt samples. These goals will be accomplished in 3 interlinked steps: (a) *assessment of gene expression*, focusing on the molecular responses in delta smelt exposed to environmental water samples collected throughout the Delta; (b) *toxicant identification*, based on the functional classification of these gene responses and response profiles generated under a previous IEP POD project, and (c) *site-specific response classification*, based on the generation of genomic profiles constructed by utilizing genomic data obtained for each specific site. Microarrays will provide information on responding genes and allow distinction between expression patterns elicited by different chemicals (expression signatures) or chemical mixtures. Quantitative PCR of select specific biomarker genes and resulting heatmaps will be applied to investigate exposure effects

on muscular, neurological, endocrine, digestive, immune, respiration, membrane disruption and ion exchange alterations in the delta smelt.

Time period: July 2010 - June 2011

Resources required:

Cost: \$220,000 in 2010

PI(s): Inge Werner and Richard Connon (UCD)

Contract needed / in place: In progress.

Contract manager(s): Erwin Van Nieuwenhuysse (USBR) and Inge Werner (UCD).

Term of contract: To be determined.

Personnel: Above named investigators, graduate students and research assistants.

Equipment: No new equipment required.

Deliverables and dates: (assuming a start date of July 1, 2010)

- Study plan, peer-reviewed by the POD CWT (July 2010).
- Quarterly progress preports to the POD CWT (September and December 2010 and March 2011).
- Final report (June 2011).
- 1 or more IEP Newsletter articles (June 2011).
- 1 or more peer-reviewed journal publication(s) (June 2011).

How is this activity related to the POD investigations and what is its relevance to understanding the effects of fall flow variations on delta smelt? This study directly addresses the *IEP Priority Research Topic 1: Habitat Effects on Delta Smelt Population Dynamics*. A gene microarray and novel biomarkers for delta smelt (developed as part of previous IEP POD funded work) will be refined and used to determine exposure to and measure effects of contaminants responsible for causing toxicity to the endangered delta smelt. This study builds on successful previous work funded by IEP POD investigating the effects of environmental contaminants as a factor in the observed decline of delta smelt. This project will provide a sensitive tool to assess the effects of environmental stressors on delta smelt, and to develop site-specific genetic fingerprints for evaluating the effects of seasonal variations in environmental conditions, including flows.

Comments: This study builds on successful previous work funded by IEP POD including toxicity tests with larval delta smelt, determination of toxic thresholds of delta smelt for a number of relevant chemicals, development of a delta smelt specific cDNA microarray, and development of a suite of delta smelt-specific molecular biomarkers for assessing the effects of environmental stressors. Archived delta smelt samples from laboratory tests performed in 2008 and 2009 as part of previous IEP POD funded work will be used for this study.

Disease and Physiology Monitoring in Wild Delta Smelt Adults

IEP 2010-184

Point person: Randy Baxter (DFG)

Lead Agency: USFWS

Questions: What is the influence of pathogens, contaminants, and adverse water quality on delta smelt survival?

Description: The projects objective is to survey subadult and adult delta smelt populations collected in the lower Sacramento River for fish pathogens, tissue abnormalities (histology), energy reserves (muscle triglycerides) and osmoregulatory status (gill Na-K-ATPase). The

minimum target sample is 100 smelt collected by DFG's SKT Survey. A maximum of 20 fish will be collected at any given station.

Time period: Field samples will be collected from January – May 2010.

Resources required:

Cost: This work is not being funded by IEP sources.

PI(s): Scott Foott (USFWS)

Contract needed / in place: N/A

Contract manager(s): N/A

Term of contract: N/A

Personnel: DFG will supply field personnel during the SKT Survey.

Equipment: No new equipment needs to be purchased.

Deliverables and dates:

- Results from laboratory analysis of samples and production of a technical report (August 2010).

Physical Processes Influencing Spawning Migrations of Delta Smelt

IEP 2010-187

Point person: Larry Brown (USGS)

Lead Agency: USGS

Questions: The primary goal of the proposed experiment is to evaluate whether the annual spawning migration of delta smelt is triggered by a sudden decrease in water transparency, i.e. high turbidity.

Description: This study was designed as an outgrowth the hypothesis that that adult delta smelt initiate an upstream spawning migration when the first major storm causes a turbidity pulse above some threshold. The purpose of this element is to test the hypothesis and also determine how delta smelt move upstream, specifically do they take advantage of tidal flows (i.e., tidal surfing). Additional secondary goals/questions will also be addressed to the extent possible. The questions will be addressed using intensive Kodiak trawling before and after the first major storm of the year in coordination with collection of continuous tidal velocity and turbidity data.

Time period: This study is expected to be implemented in the Fall/Winter of 2010. It is unclear if the study will be repeated in additional years. Additional years will likely depend on costs and results obtained in year 1.

Resources required:

Cost: Cost undetermined at this point. Funding expected from DOI through USBR.

PI(s): Jon Burau (USGS) and Bill Bennett (UCD).

Contract needed / in place: Contact/agreement needed

Contract manager: Presumably Erwin Van Nieuwenhuysse (USBR)

Term of contract: Unknown. Anticipate year by year funding.

Personnel: Jon Burau (USGS), Bill Bennett (UC Davis), field help expected from USGS (Sacramento and Columbia River Lab) and possibly IEP (boat operators and crew).

Logistics still being worked out.

Equipment: Boats likely to be borrowed from USGS Columbia Lab. Some equipment may need to be purchased (e.g., nets)

Deliverables and dates:

- Progress report and/or presentation to the Fish Migration PWT, after field work completed and as initial analysis 2011.

- IEP Asilomar presentation (February 2012).
- Final report or journal article, if appropriate (December 2011).

How is this activity related to POD or HSG: The results of this analysis will contribute to understanding how turbidity and possibly other factors influence spawning movements of adult delta smelt.

Comments:

OP and Pesticide use in the Sacramento River and Delta

IEP 2010-188

Point person: Stephanie Fong (CVRWQCB)

Lead Agency: CVRWQCB

Questions: Are past organophosphate (OP) hotspots still exhibiting toxicity in traditional EPA 3-species toxicity tests or to *Hyaletta azteca*? Have the OPs been replaced by other pesticides, and are they causing toxicity?

Description: This study will use traditional EPA 3-species testing in addition to *H. azteca* testing and new, sub-lethal endpoints to assess locations in the Sacramento River and Delta that had a history of OP-caused toxicity.

Time period: March 2010-June 2011

Resources required:

Cost: \$175,000 contract and CVRWQCB staff time.

PI(s): Inge Werner (UCD)

Contract needed / in place: In place.

Contract manager(s): Stephanie Fong (CVRWQCB) and Inge Werner (UCD).

Term of contract: May 2010 to May 2011.

Personnel: Inge Werner and graduate student (UCD), Stephanie Fong (CVRWQCB)

Equipment: No new equipment is needed; these are typical laboratory tests.

Deliverables and dates:

- Peer-reviewed study plan (July 2010).
- Progress reports and/or participation in the POD CWT, quarterly throughout the study.
- Final report (June 2011).

How is this activity related to HSG? The results of this analysis will contribute to the abiotic habitat driver section of the fall X2 and delta smelt conceptual model.

Comments: This contract was combined with the acute and chronic toxicity of contaminant mixtures study for ease of contracting.

Ammonia Literature Review and Synthesis

IEP 2010-189

Point person: Mark Gowdy (SWRCB)

Lead Agency: SWRCB

Questions: What do existing data, studies and literature have to say about the effects of ammonia on aquatic species, and are there any data gaps concerning ammonia concentrations and species in the Delta?

Description: Review existing data, studies, and the literature on the effects of ammonia on aquatic species. Prepare a synthesis report on these studies and determine if there are any data gaps, specifically with respect to ammonia concentrations and species residing in the Delta.

Time period: December 2008 – June 2010

Resources required:

Cost: \$69,000

PI(s): Mike Johnson (UCD)

Contract needed / in place: In place.

Contract manager(s): Mark Gowdy (SWRCB)

Term of contract: December 2008 – June 2010

Personnel: Mike Johnson), Ling Chu, Jennifer Nickell and Zephyr Papin (UCD)

Equipment: No new equipment is needed.

Deliverables and dates:

- Synthesis report on effects of ammonia on species in the Delta.

Comments: None